## A CRITICAL ASSESSMENT OF CNG AS AN ALTERNATIVE FUEL IN PUBLIC BUS TRANSIT IN DELHI, INDIA

by

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#### ABSTRACT

The rapid motorization of transport in Indian cities has led to the implementation of several policies to address motor vehicle emissions, including the conversion of city buses to run on compressed natural gas (CNG) from around the year 2000. As part of this conversion, Delhi Transport Corporation's (DTC) more than 3,000 diesel buses were replaced by Standard (high-floor) CNG buses, and from 2007-08, nearly 4,000 low-floor CNG buses have replaced the Standard CNG buses. CNG has also been implemented in the public bus fleets in other Indian cities.

To help assess the long-term desirability of replacing diesel with CNG in India and similar contexts, and for informing techno-economic and environmental analyses of CNG bus transit operations, I assess the operational and financial performance consequences of, and the cost-effectiveness of emissions reductions due to CNG implementation in Delhi, India.

My research shows that CNG implementation caused a significant reduction in DTC's capacity to deliver transit service in the initial stages of the fuel transition, and also necessitated investments in buses at a considerable cost premium relative to their diesel counterparts. Further, bus life-cycle costs (LCCs) are higher for CNG than for diesel, but CNG negatively affects the LCC of Standard buses proportionately more than for the low-floor buses, for which the LCC is already high. The cost-effectiveness analysis of CNG bus emissions reductions showed that, while Standard CNG buses, relative to their diesel counterparts, offered the most potential in reducing PM and  $CO_2(e)$  emissions in the early 2000s, the most cost-effective choice in tackling these emissions would have been to upgrade to a new fleet of diesel buses running on lower sulphur fuel and using improved exhaust aftertreatment systems. Also, considering the current implementation of CNG in other Indian cities, my analysis suggests that the higher costs of CNG may not justify the environmental benefits compared to available cleaner diesel bus technologies.

The broader question I raise is that the financial situation resulting from these effects due to CNG implementation in Delhi may have detracted from the ability to enhance transit capacity and provide transit service overall. My research also shows the critical importance of the fuel price and fuel economy of CNG, for the competitiveness of CNG relative to diesel buses, and demonstrates the need for careful fuel pricing policies when CNG is implemented in bus transit. Finally, I also demonstrated the need to analyze policies such as CNG implementation broadly, in terms of conflicts and trade-offs between environmental, and other (transit operation, socio-economic and equity) objectives, rather than narrowly in terms of only environmental outcomes.

#### RÉSUMÉ

La rapide motorisation des transports dans les villes indiennes a engendré l'implémentation de plusieurs politiques pour répondre au problème des émissions de gaz véhiculaire, y inclus la conversion des autobus au gaz naturel comprimé (GNC), aux alentours de l'an 2000. Partie impliquée dans cette conversion, Delhi Transports Corporation (DTC) a remplacé plus de 3.000 autobus roulant au diesel par des autobus standard (plancher haut) roulant au GNC et à partir de 2007-2008, près de 4000 autobus à plancher surbaissé ont remplacé les autobus standard roulant au GNC. Les autobus au GNC ont aussi été implémentés dans d'autres villes indiennes.

Pour aider à calculer le taux souhaitable de remplacement à long terme du diesel par le GNC en Inde et dans des contextes similaires, j'examine les conséquences opérationnelles et financières des changements et le coût-avantage des réductions des émissions dû à l'implantations du GNC à Delhi, en Inde.

Ma recherche montre que l'implémentation du GNC a engendré une réduction significative dans la capacité de la DTC de pourvoir le service de transit dans les premiers stages de la transition u combustible, et a aussi demandé des investissements en autobus avec un coût considérablement plus élevés que leurs équivalents roulant au diesel. Plus, le coût du cycle de vie (CCV) des autobus est plus élevé pour les autobus roulant au GNC que pour ceux utilisant le diesel, mais le GNC affecte négativement le CCV des autobus standard proportionnellement plus que ceux au plancher surbaissé, pour lesquels le CCV est déjà haut. L'analyse du coût-avantage des bus roulant au GNC a montré que, alors que les autobus standard roulant au GNC, relativement à leurs équivalents roulant au diesel, offraient le plus grand potentiel à réduire les émissions de matières particulaires (PM) et CO2(e) au début des années 2000, le choix le plus effectif considérant le coût pour faire face au problème des émissions aurait été de passer à des nouveaux bus roulant au diesel avec un taux plus bas de soufre et améliorer les systèmes de post-traitement des gaz d'échappement. Aussi, considérant la courante implantation de GNC dans d'autres villes indiennes, mon analyse suggère que les coûts plus élevés du GNC peuvent ne pas justifier les bénéfices environnementaux, comparés à des technologies des autobus roulant au diesel plus propre.

La question plus élargie que je pose est que la situation financière résultante des actions dues à l'implantation du GNC à Delhi peut avoir endommagé la possibilité d'améliorer la capacité du transport et d'offrir un service de transport en général. Ma recherche montre aussi l'importance critique du prix du combustible et de l'économie de combustible due résultante du GNC pour la compétitivité des bus roulant au GNC relative à ceux utilisant le diesel et montre le besoin d'une politique soigneuse lors de l'implantation du GNC aux autobus. Finalement, je montre aussi le besoin de l'analyse des politiques tels que la large implémentation du GNC en termes de conflits et compensations parmi les objectifs environnementaux et autres (gestion du transport, socio-économique et équité), plutôt que de se limiter seul aux résultats environnementaux.

## TABLE OF CONTENTS

ABSTRACT	i
RÉSUMÉ	ii
TABLE OF CONTENTS	iii
LIST OF TABLES	vi
LIST OF FIGURES	viii
LIST OF TERMS AND ABBREVIATIONS	ix
AUTHOR CONTRIBUTIONS	xv
CONTRIBUTIONS TO ORIGINAL KNOWLEDGE	xvi
ACKNOWLEDGEMENTS	xvii
Chapter 1: Review of issues related to alternative transport fuel use	1
1.1 Introduction	1
1.2 Motorization and impacts	2
1.3 The role of alternative transport fuels in addressing urban transport emissions	8
1.4 Barriers to alternative transport fuel use	10
1.5 Compressed natural gas for urban bus transit fleets	13
1.6 CNG in public bus transit in Delhi	16
Chapter 2: Research objectives and methodological framework	20
Chapter overview	20
2.1 Motivation, knowledge gaps and research needs	20
2.2 Research objectives	25
2.3 Thesis structure, analytical framework, and methodological approach	26
2.4 Data and issues	29
Chapter 3: Natural gas fueled urban bus transit: US and Latin American ex	perience and
lessons for rapidly motorizing countries	31
Chapter overview	
3.1 Methodology and outline	
3.2 Analysis of CNG experience	34
3.2.1 Rationale for CNG implementation in urban bus transit	34
3.2.2 Evaluation of CNG performance on urban bus transit fleets	35
3.2.3 Vehicle and fuelling infrastructure	
3.2.4 Vehicle and fuel systems performance	44
3.3 Discussion and policy implications	67
3.3.1 Summary of key findings	67

3.3.2 Conclusion and policy implications	71
Chapter 4: Operational and financial performance of Delhi's CNG-fueled public bus	
transit fleet: A critical evaluation	74
Chapter overview	74
4.1 Introduction	74
4.1.1 CNG implementation on DTC's bus fleet	77
4.2 Analytic framework, methodology and data	80
4.3 Results and discussion	82
4.3.1 Operational performance	82
4.3.2 Financial performance	89
4.4 Conclusions and implications	97
Chapter 5: CNG and diesel urban buses in India: A life-cycle cost comparison	100
Chapter overview	100
5.1 Introduction	100
5.1.1 CNG implementation on DTC's bus fleet	101
5.1.2 Rationale and objectives	102
5.2 Analytic framework, methods, and data	105
5.2.1 Bus and fuel systems evaluated	105
5.2.2 Analytic framework	107
5.3 Results and discussion	115
5.3.1 LCCs for CNG-Std. versus CNG-LF versus CNG-LF/AC in the Indian context	115
5.3.2 LCCs for CNG versus diesel buses for Std., LF and LF/AC bus configurations in the Inc	lian
context	117
5.3.3 LCCs of CNG and diesel LF/AC buses in India, with US fuel prices	118
5.3.4 Sensitivity analysis	120
5.4 Conclusions, limitations, and implications	123
Chapter 6: Cost-effectiveness of CNG implementation in the public bus transit fleet	in
Delhi	126

Chapter overview	126
6.1 Introduction	126
6.2 Analytic framework, methods and data	
6.2.1 Bus and fuel technologies evaluated	
6.2.2 Cost-effectiveness analysis	
6.2.3 Scenario analysis	139

6.3 Results and analysis	145
6.3.1 Analysis of emissions results for the three scenarios	145
6.3.2 CE ratio analysis	148
6.4 Conclusions and implications	
6.4.1 Summary of key results	
6.4.2 Implications	154
Chapter 7: Conclusion and policy implications	158
7.1 Introduction	
7.2 Key research findings	
7.3 Policy implications of research	
7.3.1 Other issues	
7.4 Scope for future research	171
7.4.1 Trade-off between costs and public transit capacity and its impacts	171
7.4.2 CNG feasibility	172
7.4.3 Fuel supply constraints	174
7.4.4 Natural gas supply scalability	175
References	178
Appendix A: List of sites visited in Delhi for data collection	200
Appendix B: Supplementary material for Chapter 5	202
Appendix C: Supplementary material for Chapter 6	206
C.1: Methane (CH <sub>4</sub> ): tailpipe and fugitive emissions	206
C.2: Black Carbon (BC)	
C.3: Carbon Dioxide (CO <sub>2</sub> )	
C.4: CO <sub>2</sub> equivalent (GWP for BC, CH <sub>4</sub> , CO <sub>2</sub> )	

## LIST OF TABLES

Table 1.1: Key energy, demographic and economic statistics, 2016	6
Table 1.2: Transit bus fleet numbers, India, 2015	14
Table 2.1: Dimensions and perspectives reflected in dissertation	27
Table 3.1: Summary of cases evaluated	33
Table 3.2: Summary of selected bus system specifications	42
Table 3.3: Summary of selected performance parameters of buses	43
Table 3.4: WMATA chassis dynamometer-based emissions, 2000-2004 vintage buses,	CBD
drive cycle	43
Table 3.5: WMATA chassis dynamometer-based emissions, 2003-2006 vintage buses,	various
test cycles	51
Table 3.6: Simulated fuel costs based on Table 3.3 parameters	65
Table 4.1: CNG bus technologies at DTC	79
Table 5.1: Summary of key bus systems evaluated in India	106
Table 5.2: LCC model parameters in Equation (1)	108
Table 5.3: Operating and financial assumptions used in LCC model	109
Table 5.4: Comparison of bus life-cycle costs (\$/km)	115
Table 5.5: Differential life-cycle costs, by bus type and cost category	116
Table 6.1: Potential CE ratio outcomes	133
Table 6.2: Present value of bus LCCs, \$-per-bus, 12-yr service life, 12% disc.rate	135
Table 6.3: Choice of greenhouse gas equivalent emission pollutants	136
Table 6.4: Emission factors for buses, g/km	138
Table 6.5: Emission factors for cars, M2Ws, and M3Ws, g/km	138
Table 6.6: Impact of bus ridership migration to private modes of transport (Scenario 3).	143
Table 6.7: Annual vehicles emissions in Delhi, g/pass-km	144

Table 6.8: Cost-effectiveness ratio calculation breakdown, Scenario 1	149
Table A1: Sites visited for data collection at DTC in Delhi, India	200
Table B1: Description of parameter assumptions used in the LCC model	202
Table C1: Heavy-duty vehicle emission norms, India vs. European Union	209
Table C2: $CO_2$ emission factor estimation for CE analysis, g/km	210
Table C3: CO <sub>2</sub> (e) emission factor estimation	210

## LIST OF FIGURES

Figure 1.1: Vehicle fleet numbers	
Figure 1.2: Total petroleum demand, by sector, World	7
Figure 1.3: Uncompressed natural gas buses in China	12
Figure 2.1: Trip modal shares in India, by city size	22
Figure 3.1: CNG versus diesel retail fuel prices in the US	63
Figure 4.1: Service provision and utilization	84
Figure 4.2: Fleet fuel economy	86
Figure 4.3: Reliability	87
Figure 4.4: Key operating expenses	92
Figure 4.5: Comparative fuel prices – diesel and CNG	92
Figure 4.6: Maintenance costs, key categories, per kilometre	95
Figure 4.7: Operating costs per passenger kilometre and operating ratio	96
Figure 5.1: Annual maintenance costs	112
Figure 5.2: Comparison of bus life-cycle costs	117
Figure 5.3: CNG and diesel fuel prices and ratios, India and USA, 2002-2015	120
Figure 5.4: Sensitivity of LCC to fuel economy and fuel price in India	121
Figure 5.5: Sensitivity of LCC to fuel economy and fuel price – effect of discount rate	123
Figure 6.1: Comparison of passengers carried and load factor at DTC	139
Figure 6.2: Cost-effectiveness ratio, \$-per-emissions reduction, Scenario 1	151
Figure 7.1: Natural gas supply and demand, India, 2016	176
Figure A1: DTC bus depot layout, Sukhdev Vihar depot, 2010-11	201

## LIST OF TERMS AND ABBREVIATIONS

\$	United States Dollar (\$), or USD; unless stated otherwise, all financial figures with \$ symbol are expressed in US Dollars
\$/L	United States Dollars (\$) per litre
% wt.	Percentage with respect to weight
AMC	Annual Maintenance Contract (DTC); see Footnote (3) of Table 5.3.
Αντα	Advanced Vehicle Testing Activity, US Department of Energy
BC	Black Carbon
BEST	Brihan Mumbai Electric Supply & Transport Undertaking
ВМТС	Bangalore Metropolitan Transport Corporation
BRT	Bus Rapid Transit
BS-I to IV	Bharat Stage I (to IV); specifications for emissions from new vehicles set by the Bureau of Indian Standards, and equivalent to Euro-I (to IV) emission standards (CPCB, 2010)
BTU	British Thermal Unit
BTU/cu.ft.	British Thermal Unit per cubic feet
BTU/GAL	British Thermal Unit per Gallon
САРКМ	Passenger Carrying Capacity-Kilometres
CBD	Central Business District Cycle; a chassis dynamometer emissions testing procedure for heavy-duty vehicles
CE	Cost-effectiveness
CH₄	Methane
CI	Compression Ignition (engine)
CIRT	Central Institute of Road Transport, Pune, India
CNG	Compressed Natural Gas
CNG-LF	Low-Floor CNG bus (see Table 4.1 and Table 5.1 for vehicle specifications)
CNG-LF/AC	Low-Floor air-conditioned CNG bus (see Table 4.1 and Table 5.1 for vehicle specifications)

CNG-Std.	Standard CNG bus (see Table 4.1 and Table 5.1 for vehicle specifications)
со	Carbon Monoxide
CO <sub>2</sub>	Carbon Dioxide
CO <sub>2</sub> (e)	CO <sub>2</sub> -equivalent emissions
CPI and CPI-IW	Consumer Price Index; In India, CPI-IW for Industrial Workers
CWI	Cummins Westport Inc.
D15	Diesel with sulphur content of 15 ppm (or 0.0015% on a weight basis, of Sulphur in diesel; 15 mg/kg)
D30	Diesel sulphur content of 30 ppm (or 0.003% on a weight basis, of Sulphur in diesel; 30 mg/kg)
D500	Diesel sulphur content of 500 ppm (or 0.05% on a weight basis, of Sulphur in diesel; 500 mg/kg)
DDC	Detroit Diesel Corporation
DDC S50	Detroit Diesel Corp. engine model
Diesel-LF	Low-Floor diesel bus (see Table 5.1 for vehicle specifications)
Diesel-LF/AC	Low-Floor air-conditioned diesel bus (see Table 5.1 for vehicle specifications)
Diesel-Std.	Standard diesel bus (see Table 4.1 and Table 5.1 for vehicle specifications)
DMRC	Delhi Metro Rail Corporation
DOC	Diesel Oxidation Catalyst; exhaust aftertreatment system
DPF	Diesel Particulate Filter; exhaust aftertreatment system
DTC	Delhi Transport Corporation
E85	Ethanol; high-level ethanol-gasoline blend containing 51%-83% ethanol
EF	Emission factor(s); for vehicles (e.g., in g/km)
EFS	Engine and Fuel System
EGR	Exhaust Gas Recirculation System
EPA	US Environmental Protection Agency
EPA 1998, 2004	US-EPA Emission Standards for Heavy-Duty Diesel Engines
Est.	Estimated

ETC	European Transient Cycle; a chassis dynamometer emissions testing procedure for heavy-duty vehicles
EU	European Union
EURO (I to VI)	European emission standards for new vehicles (implemented in stages over the years; stages I to VI) set by the European Committee for Standardization, CEN. For concise information see <www.dieselnet.com></www.dieselnet.com>
FE	Fuel Economy
GDP (PPP)	Gross Domestic Product measured in Purchasing Power Parity terms
GEF	Global Environment Facility Program (trust-funded grant financing scheme, World Bank)
GHG	Greenhouse Gas
GPS	Global Positioning System
GVW	Gross Vehicle Weight
GWP	Global Warming Potential
НС	Hydrocarbon
НР	Horsepower, measure of power
IGL	Indraprastha Gas Limited, Delhi
INR	Indian Rupee
IPCC	Intergovernmental Panel on Climate Change
IPT	Intermediate Public Transit; for example, auto rickshaws, taxis, and for hire minibuses.
JD	John Deere
JNNURM	Jawaharlal Nehru National Urban Renewal Mission (India)
kg	Kilogram
km/diesel eq. L	Kilometres-per-diesel equivalent litre
km/h	Kilometres-per-hour
km/L	Kilometres-per-litre
L	litre
LCC	Life-Cycle Cost

LF	Low-Floor (bus)
LF/AC	Low-Floor Air Conditioned (bus)
LWV	Light-weight Vehicles (GVW < 3850 kg)
m²	Metres squared
M2W	Motorized Two-wheeled Vehicles (e.g., motorbike)
M3W	Motorized Three-wheeled Vehicles (e.g., auto rickshaws)
Man Bus	Manhattan Bus Test Cycle; a chassis dynamometer emissions testing procedure for heavy-duty vehicles
МВ	Mercedes Benz
MBRC	Miles between roadcalls
МСМА	Mexico City Metropolitan Area
MCS	Mexico City Schedule: chassis dynamometer testing procedure for heavy-duty vehicles
MJ	Mega joule
мт	Metric Tonnes
MTC-CNI	Metropolitan Transport Corp. Ltd., Chennai
Mtoe	Million tonnes of oil equivalent
MX1	Chassis dynamometer testing procedure; part of MCS cycle (see MCS)
MX2	Chassis dynamometer testing procedure; part of MCS cycle (see MCS)
МХЗ	Chassis dynamometer testing procedure; part of MCS cycle (see MCS)
MY (e.g., MY05)	Model Year; that is, the year of a vehicle's model
N <sub>2</sub> O	Dinitrogen Oxide (nitrous oxide)
NCR (Delhi)	National Capital Region (of Delhi)
NG	Natural Gas
МНС	Non-Methane Hydrocarbons
non-OECD	Non-OECD (countries); see OECD
NO <sub>x</sub>	Nitrogen Oxides
NREL	National Renewable Energy Laboratory (US Department of Energy)

NY Bus	New York Bus Cycle; a chassis dynamometer emissions testing procedure for heavy-duty vehicles
NYCT	New York City Transit, New York, NY
<b>O</b> <sub>3</sub>	Ozone
oc	Organic Carbon
OECD	Organisation for Economic Co-operation and Development of 36 member countries, most with high income per-capita
OEM	Original Equipment Manufacturer
pass-km	Passenger-kilometre(s)
PKMS	Passenger-kilometre(s)
PM	Particulate Matter (typically defined by size in aerodynamic diameter)
PM <sub>0.1</sub>	PM with aerodynamic diameter of 0.1 microns and smaller
PM <sub>2.5</sub>	PM with aerodynamic diameter of 2.5 microns and smaller
PM <sub>10</sub>	PM with aerodynamic diameter of 10 microns and smaller
ррт	Parts per million
PPP	Purchasing Power Parity (see GDP-PPP)
RAVEM™	Ride-Along Vehicle Emission Measurement system by Engine, Fuel, and Emissions Engineering Inc. of Sacramento, California
RTP	Red de Transportes de Pasajeros del D.F., Mexico
SCR	Selective Catalytic Reduction (SCR) catalyst; exhaust aftertreatment system
SI	Spark-Ignited (engine)
SO <sub>2</sub>	Sulphur Dioxide
Std. or Std. bus	Standard [size bus]; applicable in the Indian context; see Tables 4.1 and 5.1 for standard CNG and diesel bus specifications
SunLine	SunLine Transit Agency, Palm Springs, CA
TERI	The Energy and Resource Institute
Therm	Unit of heat energy (= 100,000 BTU)
TRIPP	Transportation Research and Injury Prevention Programme, IIT-Delhi

ULSD	Ultra-Low-Sulphur Diesel; US-EPA defines ULSD as a diesel fuel with sulphur content lower than 15 ppm; e.g., see <https: diesel-fuel-standards="" www.epa.gov=""></https:>				
US or USA	United States of America				
USD	United States Dollar or \$ (see \$, above); unless stated otherwise, all financial figures with \$ or USD are expressed in US Dollars				
USDoE	US Department of Energy				
USDoT	US Department of Transportation				
VKMS	Vehicle-Kilometre(s)				
voc	Volatile Organic Compounds				
who	World Health Organization (United Nations)				
WMATA	Washington Metropolitan Area Transit Authority				
WMATA Cycle	Washington Metropolitan Area Transit Authority Cycle; a chassis dynamometer emissions testing procedure for heavy-duty vehicles				
µg/m³	Micro-grams per cubic metre				

#### **AUTHOR CONTRIBUTIONS**

My doctoral dissertation is structured following a manuscript-based format and integrates four analytical Chapters (3 to 6), in addition to the review of issues related to alternative transport fuel use and the overall problem statement (Chapter 1), the thesis research objectives and methodological framework (Chapter 2), and a synthesis of the key conclusions and policy implications (Chapter 7). Chapters 4 and 5 have been published in peer-reviewed journals. I provide here details of my contributions to the research activities related to and the writing of the chapters in my dissertation.

Chapter 4 is a co-authored paper titled "Operational and financial performance of Delhi's CNG-fueled public bus transit fleet: A critical evaluation" that has been published in the *Transport Policy* journal (Krelling & Badami, 2016). I was the primary author of this publication; conceptualization of the paper was done collaboratively with the second author, Madhav G. Badami, while I designed the analytic framework, carried out the fieldwork for data collection, conducted the analysis, and wrote the paper. Madhav G. Badami provided extensive and detailed feedback and comments for the revision and editing of the final paper.

Chapter 5 is a co-authored paper titled "CNG and diesel urban buses in India: A lifecycle cost comparison" that has been published in the *International Journal of Sustainable Transportation* (Krelling & Badami, 2019). I was the primary author of this paper, and conceptualized the research project, designed the analytic framework, determined data needs, carried out the fieldwork for data collection, conducted the analysis, and wrote the paper; Madhav G. Badami provided extensive and detailed feedback and comments for the revision and editing of the final paper.

I conceptualized the research, designed the analytic framework, collected the data, and conducted the analysis for all of the other chapters in my dissertation, as well as writing them. Madhav G. Badami contributed to their conceptualization, and provided extensive and detailed feedback and comments for revising and editing them.

#### CONTRIBUTIONS TO ORIGINAL KNOWLEDGE

The elements of my thesis that are original scholarship and contributions to knowledge, center on the fact that most research published to date on the implementation of CNG in Delhi's public vehicles has focused on its emissions outcomes with little attention devoted to its operational or financial aspects. Particularly, analyzing the operational and financial performance of CNG-fueled buses in Delhi's public bus transit fleet -- as I have done in Chapters 4 and 5 -- is important because it is this performance that critically determines the vehicle operator's policy responses, which in turn will crucially determine the extent to which implementation, and the associated emissions reductions, actually occur and are successful. This work will help assess how the conversion from diesel to CNG affected bus fleet operations and finances. In addition, I contribute to an important policy-analytic objective when I evaluate the cost-effectiveness of the implementation of CNG in Delhi in Chapter 6. This is an important contribution to knowledge, because I use a framework that integrates operational and financial dimensions with the existing emissions performance research in comparing the related life-cycle costs against their respective emissions outcomes. As a whole, the analyses I conduct in this dissertation will hopefully be useful to policy and decision makers and urban bus transit operators in contexts similar to India's, by drawing lessons for the long-term desirability of large-scale conversions of bus fleets to CNG, for comparison with other CNG bus transit fleets, and for informing techno-economic and environmental analyses of CNG bus transit operations.

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For my PhD fieldwork, with the access provided through Prof. Badami's professional network of colleagues in India, I was able to count on various people and institutions for logistical support. This support included sponsoring of research visas, providing office space and network access, helping in gathering of data, being available for extensive discussions and meetings, and providing access to seminars and conferences, all of which helped me understand the complexity of the research problems being investigated and thus helped shape my research. In this respect, I would like to thank Professors Geetam Tiwari and Dinesh Mohan of the Transportation Research and Injury Prevention Programme (TRIPP), IIT Delhi, Professor Sanjivi Sundar of The Energy and Resource Institute (TERI), and Dr. Rajiv Seth of TERI University, for hosting me during my field-work. In addition, I owe a debt of gratitude to Mr. V.K. Sehgal and Mr. S.P. Sethi for providing access to data at Delhi Transport Corporation (DTC). Finally, I would not have been able to conduct my field-work in India, if not for Suparna, Manna and Zorro, who kindly opened their home to me for the long duration of my stays in Delhi; their friendship and help were priceless.

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This work is dedicated to my wife and son, Nadja and Matthew.

#### Chapter 1: Review of issues related to alternative transport fuel use

#### **1.1 Introduction**

The principal motivation guiding this research is the significant impact of urban transport on air pollution in rapidly motorizing low- and middle-income countries, with serious health and welfare effects for urban populations and the environment, and the prospect of continued and fast growth in vehicle activity, which will tend to aggravate these impacts. The central objective of this dissertation is to assess the operational and financial impacts of, and the cost-effectiveness of emissions reductions due to, replacing diesel bus fleets with compressed natural gas (CNG) in India, in order to support decision- and policy-making related to the use of alternative transport fuel systems in low- and middle-income country contexts. To this end, my doctoral thesis entitled "A Critical Assessment of CNG as an Alternative Fuel in Public Bus Transit in Delhi, India," will address a range of dimensions concerning CNG implementation. For this research, I have focused mostly on the particular case of Delhi Transport Corporation (DTC) in Delhi, but also have drawn on the experiences of other public transit bus operators in India, and even internationally, for comparison and critical evaluation.

Over the past 30 years, many high-income countries, as well as low- and middle-income countries, have implemented a number of alternative transport fuel technologies. These actions were mainly driven by the rapid rise in urban transport activity and the negative impacts associated with this trend, especially in terms of deteriorating urban air quality due to the increase in transport emissions. The rapid increase in motorization coupled with the heavy dependence of road transport on petroleum-based fuels raises a number of questions, not only in terms of the impacts of vehicle activity on urban air quality, but also in terms of a range of other issues, such as the capacity of countries in obtaining petroleum in the future (e.g., what are energy security implications of no diversification?), the challenges associated with oil substitution in the transportation sector, and the economic and other implications of fuel systems choices. Moreover,

despite the many policy efforts across the globe to diversify transport energy sources, transportation continues to be predominantly dependent on oil for supply of most of its energy needs. For low- and middle-income countries, transportation energy challenges are even more critical, since rapid economic development and increased preference for personal mobility has significantly pushed up personal motor vehicle use, and with this, the impacts of motorization have grown even more. Yet, despite the challenges and incremental costs associated with alternative transport fuel use, a few large developing nations, such as India and Brazil, have already adopted them at a significant scale in attempts to address urban air pollution, and to promote energy security.

Below, I present a detailed discussion of the above and other issues, in order to justify public policy attention to the use of CNG in India, which forms the focus of my dissertation. In Chapter 2, I present the rationale, objectives, and methods adopted across my dissertation; in Chapter 3, I present a comprehensive and critical review of the literature concerning the implementation of CNG in public bus transit fleets in the US and Latin America; in Chapters 4 to 6, I critically analyze various aspects of CNG implementation in India; and, in Chapter 7, I conclude my research, discussing the broader policy implications of my findings.

#### **1.2 Motorization and impacts**

The global motor vehicle fleet has grown substantially from 246 million vehicles in 1970 to 1.3 billion in 2015 (Davis et al., 2017). Moreover, the magnitude of these figures is only compounded by the fact that this data does not include motorized two-wheeled vehicles (M2W), which account for a substantial portion of vehicle stocks in many rapidly motorizing low- and middle-income countries. For example, in India, M2Ws accounted for approximately 75% of all registered motor vehicles in 2016 (MoRTH, 2018). Figure 1.1 shows the trend in worldwide vehicle fleet numbers since 1990, when detailed information is available for selected countries.



Source: Davis et al. (2017): Table 3.2 and Table 3.3.

#### Figure 1.1: Vehicle fleet numbers.

Analysis of data from Figure 1.1 shows that low- and middle-income countries are the main contributors to the growth in worldwide motor vehicles in the past couple of decades. From 1990 to 2015, the motor vehicle fleet in nations like Brazil, Argentina, Russia and Pakistan increased approximately 3 times; Asian countries such as India, Indonesia, and Malaysia saw motor vehicle numbers grow 8 times, and China by 25 times (Davis et al., 2017). During this same period, vehicle fleet numbers in the US, Japan, France, UK, Germany and Canada grew only 1.4 times (Figure 1.1). Particularly India and China accounted for 28% of worldwide fleet growth from 1990 to 2015, but only represented 16% of the world's vehicle fleet in 2015 (Davis et al., 2017).

While the challenges in road transport are many, in areas as diverse as road safety, energy security and climate change, the impacts of air pollution and congestion caused by increased motorization have attracted much policy attention. There is a significant, and increasing, contribution from motor vehicles to total air pollution loads, especially in urban areas (CAA, 2012), and substantial evidence that exposure to various air pollutants emitted by vehicles are directly linked to increased human morbidity and mortality risks (WHO, 2013). High concentrations of fine particulate matter of 2.5 microns or less in diameter (PM<sub>2.5</sub>), a key component of motor vehicle emissions, are strongly associated with adverse health outcomes (Pope et al., 2011), and data shows that ambient particulate matter levels in large cities of low- and middle-income countries, and particularly in East and South Asia, are far higher than those observed in high-income countries, and vastly exceed international air quality guidelines (WHO, 2014a). In 2012, air pollution was estimated to have caused approximately 3.7 million premature deaths, with a high portion of these deaths (88%) having occurred in low-and middle-income countries, with the greatest number of these deaths in the Western Pacific and South-East Asia regions (WHO, 2014b).

In India, 78% of cities had PM<sub>10</sub> levels that were at or above the annual national ambient standard of 60 µg/m<sup>3</sup> in 2010, with 63% of this group exceeding the standard by at least 1.5 times (CPCB, 2012: Tables 2.4 and 2.11). In the case of Delhi, the focus of this dissertation, the annual average PM<sub>10</sub> concentration in 2010 was 261 µg/m<sup>3</sup> (CPCB, 2012: Table 2.7), which is 4.4 and 13 times the Indian and WHO standards<sup>1</sup>, respectively. What is also very concerning is evidence showing that populations located along roadways receive exposure levels that can be on average 1.5 times greater than those reported by centrally located measuring stations (Apte et al., 2011). This suggests an uneven distribution of air pollution impacts. The populations most exposed and at greater risk -- that is, those with lower incomes, such as people working or living near roads, such as non-motorized commuters, rickshaw operators and users, and street vendors -- who are also least able to cope with illness are the ones most exposed to and affected by health effects due to air pollution.

What is alarming is that the outlook for urban air quality is bleak in regions such as East and South Asia since many cities continue to experience an increasing trend in

<sup>&</sup>lt;sup>1</sup> WHO's air quality guideline is an annual mean of 20 µg/m<sup>3</sup> for PM<sub>10</sub> and 10 µg/m<sup>3</sup> for PM<sub>2.5</sub> (WHO, 2014a).

particulate matter concentrations, primarily due to increased vehicle activity (CAA, 2012; WHO, 2014a; Guttikunda & Mohan, 2014). Further, while local air pollution impacts are very important, there are other important global consequences of road transport emissions, as this sector is the fastest growing contributor to global CO<sub>2</sub> emissions, most of which will increasingly be emitted by China and India over the next 25 years (CAA, 2012). It is for all these reasons, but mainly for local air quality effects, that natural gas -- as well as other alternative fuels -- has been considered as a petroleum substitute in urban areas.

Another important consequence of rapid increase in motor vehicle activity is the increase in road traffic congestion. Rapidly motorizing developing countries typically lack adequate road infrastructure capacity to support the increase in vehicle activity, and therefore a key consequence is the rapid increase in congestion. The outlook on congestion for countries like India is not positive, since, while car ownership rates still lag those observed in richer countries, public transport systems have not been able to meet growing urban passenger trip demand, only exacerbating motorization trends (WSA, 2008a). Congestion has serious implications for the efficiency and effectiveness of transport systems, with broad impacts on living conditions of urban populations, access of people to economic opportunities, flow of goods within cities, and economic activity. Also, emissions and fuel consumption are seriously exacerbated by congestion.

In terms of aggregate energy use, the transportation sector represents 28% of total global energy consumption, with most of this energy (75%) accounted for by road transport (IEA, 2018). A comparison of countries at different stages of economic development shows that there is a significant rich-poor energy dichotomy in road transport, considering that the per-capita road transport energy use in OECD countries is 5.4 times greater than non-OECD countries (Table 1.1). In the future, this gap will likely narrow as non-OECD countries develop their economies, and GDP-per-capita increases; a trend that is corroborated by the rapid increase in vehicle numbers in low-and middle-income countries. This trend will enhance the importance of road transport

as a major energy-consuming sector of developing economies, which in turn will push demand for energy resources significantly upward.

Table 1.1: Key energy, demographic and economic statistics, 2016								
	Total Final	Transport	Road	Road Trpt.	Population	GDP		
	Energy	Energy	Transport	Energy		(PPP)		
	Consumption		Energy	per-capita		per capita		
	(Mtoe)*	(Mtoe)*	(Mtoe)*	(1,000 toe)**	(millions)	(US \$)		
World	9,555	2,748	2,055	277	7,429	14,703		
OECD	3,669	1,238	1,091	849	1,284	49,034		
non-OECD	5,488	1,112	964	157	6,145	9,796		
China	1,978	299	246	178	1,379	19,450		
India	572	90	82	62	1,324	7,905		

Table 1.1: Key energy, demographic and econor	mic statistics, 2016
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(\*) Mtoe = Million tonnes of oil equivalent

(\*\*) toe = Tonne of oil equivalent

Source: IEA (2018)

Coupled with the prospect of a significant increase in road transport energy demand in low- and middle-income countries, there is the challenge of the substantial dependence on oil in this sector, as petroleum accounted for roughly 94% of all energy needs of road transport, worldwide (IEA, 2018). This oil dependence is quite significant with road transport consuming nearly half of all petroleum used by all sectors of the global economy (IEA, 2018). More importantly, as shown in Figure 1.2, the share of road transport in oil consumption has grown considerably over the last four decades; thus, these trends in oil consumption by road transport establish an important nexus between rapid motorization and energy security issues, in addition to urban air quality and climate change challenges mentioned. As countries like India continue to experience this process of rapid motorization, given rapid economic development and structural transformation of their economies, road transport energy challenges and constraints will only continue to be exacerbated.



Figure 1.2: Total petroleum demand, by sector, World

India With 1.3 billion people, India is home to around 18% of the world population but only consumes 6% of total world energy (Table 1.1: Total Final Energy Consumption vs. Population). India's share of global road transport energy consumption is even lower, at 4% (Table 1.1), despite India's road transport energy consumption having increased more than 700% from 1981 to 2013 (OECD/IEA, 2015). Within India, the share of road transport in total oil consumption is 44%, compared to 59% in OECD countries (IEA, 2018). But with the importance of road transport only expected to grow as India continues its rapid motorization trend, the country's dependence on foreign oil imports is expected to be exacerbated; India, which is the fourth largest consumer of oil in the world, has increased its total net oil imports from 42% of total domestic oil demand in 1990 to approximately 76% of total domestic demand in 2015 (EIA, 2016). This oil-dependency shows the immense challenges for India, in terms of energy security and associated economic vulnerabilities, given oil price fluctuations. All of this demonstrates the strategic importance of road transport from an energy security perspective, and highlights the challenges to policymakers considering continued rapid

expansion of vehicle activity with little to no meaningful diversification to other energy sources.

While the reasons for India's current low relative level of road transport energy consumption may be multifaceted and complex, key factors constraining energy usage are poverty, low average incomes, and the large proportion of rural populations with low access to motor vehicles (Goldemberg et al., 2000). India's per-capita GDP is less than half of China's (Table 1.1), while the percentage of India's urban population (32%) is significantly lower than China's (53%) (World Bank, 2015). China's per-capita road transport energy consumption is almost 2.8 times greater than India's, with per-capita road transport energy figures for OECD countries being 13 times greater than India's (Table 1.1).

As India's economy continues to grow, the scale of its future transport energy needs, and particularly oil needs, will be quite substantial, both in absolute terms and relative to the rest of the world, with significant domestic and global implications in terms of a range of issues, as outlined and discussed above, but especially in terms of economic (i.e., pressure on global petroleum demand) and environmental impacts (i.e., greenhouse gas (GHG) emissions and deterioration of urban air quality), similar to the profound way in which China's rapid economic growth and shift from rural to urban populations has impacted its domestic economy and along with it that of the entire world (e.g., see China: Building the dream, 2014).

# 1.3 The role of alternative transport fuels in addressing urban transport emissions

It is for the purpose of addressing urban transport impacts, such as mitigating urban air pollution and enhancing energy security that alternative transport fuels -- predominantly ethanol and compressed natural gas (CNG), but also, biofuels, hydrogen gas (H<sub>2</sub>), and electricity -- have been studied and used throughout the world (IEA, 2013). Today, biofuels represent the highest proportion of alternative transport fuel use followed by

CNG. The US, for example, which has the largest vehicle fleet globally, had approximately 1.2 million vehicles running on alternative fuels in 2011, the majority (72%) of which were regularly using high-ethanol gasoline blends (E85), but also a sizable fleet of electric (6%) and CNG (10%) vehicles (Davis et al., 2015). The large share of ethanol powered vehicles in the US, reflects policies targeted at promoting ethanol, dating back to the oil crisis in the early 1980s, when an incentive was introduced and 300 million litres of ethanol were produced in 1981 for use as a transportation fuel; ethanol consumption then accelerated in the 1990s with various air quality concerns requiring the use of re-formulated gasoline (blended with ethanol) in order to reduce vehicle tail pipe emissions and improve air quality (Wang et al., 2007). By 2017, annual ethanol consumption in road transport in the US had increased to approximately 54 billion litres, showing the scale and magnitude of US ethanol policy (EIA, 2018).

During the same period, Brazil also promoted a large-scale biofuels program, initially driven in the late 1970s and early 1980s, by energy security and economic concerns, given the oil crisis of the period; by 2010, ethanol consumption in road transport had reached 22 billion litres annually, accounting for approximately 22% of all road transport energy use in Brazil (UNICA, 2014; OECD/IEA, 2015). Meanwhile, India has also taken substantial measures to promote the use of alternative transport fuels, which were largely driven by deteriorating urban air quality. In India, natural gas has been the most prevalent fuel alternative used in road transport with its share of road transport energy being 2.6%, compared to 2% globally; in 2016, natural gas consumption in road transport in India corresponded to 5.1% of all natural gas consumed globally within road transport (IEA, 2018).

Specifically looking at natural gas, there are many benefits for its use in road transport. In addition to reducing petroleum consumption, natural gas has been widely used for its ability in reducing motor vehicle emissions of critical air pollutants that adversely impact human health. Natural gas has excellent technical qualities as a spark-ignition (SI) engine fuel, with a higher mass-based energy content than gasoline and a much higher octane rating than gasoline, enabling higher compression ratios, and therefore higher thermal efficiencies than with gasoline in SI engines (Faiz et al., 1996). These properties, coupled with the lower carbon-to-hydrogen ratio of natural gas, potentially allow lower carbon dioxide emissions to be achieved per unit of energy than with the liquid petroleum fuels (Faiz et al., 1996; MJB, 2007).

Natural gas is a cleaner burning fuel compared to gasoline or diesel, and above all particulate matter (PM) mass emissions are significantly reduced, as air-fuel mixing is not an issue. Additionally, because no cold start enrichment is required, natural gas significantly reduces carbon monoxide (CO) and reactive non-methane hydrocarbons (NMHC) emissions (Kathuria, 2005; Narain & Krupnick, 2007). On the other hand, since natural gas is predominantly methane, emissions of this significant climate-forcing agent, which include fugitive emissions due to fuel system leakage, are higher than with gasoline or diesel (Reynolds & Kandlikar, 2008). Meanwhile, higher nitrogen oxides (NOx) emissions from natural gas engines could be expected due to higher combustion temperatures and because catalytic control is difficult owing to the exhaust gas having low reactivity (because of low NMHC emissions); however, lower NO<sub>X</sub> levels can be achieved relative to gasoline and diesel engines with lean burn natural gas engines with advanced technology such as electronic fuel injection and three-way catalysts (Posada, 2009).

#### 1.4 Barriers to alternative transport fuel use

Despite their technical advantages, potential benefits, and policy actions promoting alternative transport fuels, they account for only 6% of energy used in road transport globally, with the remainder 94% of energy needs in road transport being provided by petroleum based fuels (i.e., mostly gasoline and diesel) (IEA, 2018). Specifically for natural gas, usage still represents a mere 2% of total world road transport energy use, on an energy basis (IEA, 2018). In the US, despite the substantial number of vehicles running on alternative fuels (1.2 million) they represent only a small fraction of the estimated 269 million registered vehicles in that country (Davis et al., 2015; FHWA,

2017). Thus, alternative transport fuel use continues to be extremely limited, representing only a niche within road transport. These low adoption rates demonstrate that there are still significant barriers preventing wider use of fuel alternatives. Key adoption barriers are mostly related to higher costs of fuel production, storage, distribution, and end-use systems; as well as various considerations such as changes required to existing fuel distribution systems, and the availability of alternative fuels and vehicles (Faiz et al., 1996; IEA, 2013).

For natural gas, its low adoption rates in road transport raise questions on why this is so, since natural gas is an abundant energy source, with vast upstream, downstream and distribution networks already set up, and it is widely used in providing energy to other sectors of the economy. Natural gas is a feedstock for 15% of all worldwide final energy needs, in all sectors of the economy combined, and in particular is a feedstock for 23% of all global electricity generation (IEA, 2018). Therefore, one has to reflect on the specific challenges preventing natural gas from being more widely adopted in road transport. From a technical perspective, an important barrier is that natural gas has a significantly lower volumetric energy density, despite having a higher mass-based energy content relative to gasoline and diesel, under normal atmospheric conditions (e.g., see Figure 1.3). This is a key disadvantage for a fuel in road transport applications, since a high volumetric energy density is needed in order to maximize payload and range. In order to overcome this critical constraint for road transport applications, natural gas has to be compressed to a pressure of 200 bar in order to make it portable, thus the term compressed natural gas (CNG). This increases fuel storage system weight, with critical trade-offs in terms of decreased fuel economy, decreased vehicle range, increased refuelling frequency, reduced payload capacity, and cost implications.



Source: Service (2014).



It is largely because of the lack of a reliable widespread system of supply of natural gas -- or other alternative fuels for that matter -- at the vehicle operator level and, most importantly, also due to the need for fuels to have a high energy density on a volume basis that alternatives to petroleum continue to account for such a low share of energy in road transport as opposed to other sectors. The development of a widespread fuel distribution network, including compression and fast refuelling facilities, requires substantial and capital-intensive investments (Lowell et al., 2007). These capital needs and technical requirements raise a policy conundrum as to how to start and promote use of a new fuel, especially at large scale, while also ensuring that its adoption gains traction. The problem is that individual users or transit operators might be unwilling to switch to the new fuel, and manufacturers would be unwilling to put on the market natural gas vehicles, in the absence of a widespread fuel supply and distribution network (e.g., see OTA, 1994).

More broadly, in terms of national energy policy, a key barrier for natural gas adoption in road transport may be the impact of its large-scale adoption on feedstock availability (say for electricity generation), given the large scale of road transport energy needs. A key macro energy policy challenge is in balancing competing uses for the fuel in terms of electricity generation, transport, heating, and other needs. Consider that, of total global natural gas supplied in 2016, approximately 41% was used for electricity and heating generation, with the remainder of this feedstock being mostly used for industrial purposes (18%) and residential, commercial and other uses (26%) (IEA, 2018).

For low- and middle-income countries, the challenges of substituting petroleum with alternative transport fuels are especially significant, because of the incremental costs associated with replacing existing and complex fuel supply chains and the financing that is required. However, despite these challenges and incremental costs associated with alternative fuels, a few developing nations have already resorted to them, as was shown by the examples of Brazil and India discussed above.

#### 1.5 Compressed natural gas for urban bus transit fleets

Given the barriers discussed in the preceding section, natural gas use has been largely restricted to captive urban bus transit fleets that are operated within a limited range from a centralized depot or depots, where it is more feasible to establish and operate capitalintensive refuelling infrastructures and where buses return every day to be re-fueled. From an emissions perspective, urban buses are well suited for using CNG fuel systems, since industry standard buses are usually diesel-operated, and typically account for the bulk of PM from urban transport, even though they account for a small share of total vehicle-kilometres (Bose & Sundar, 2005).

For these reasons, notwithstanding the low share of natural gas in road transport, it is a significant alternative fuel used in urban transit buses; adoption rates have actually increased over the years, relative to traditional diesel bus numbers. For example, in 2002 in the US, 88% of large buses (i.e., with length 35 feet and over, which account for the bulk of all public transit buses) were diesel powered, 10% were CNG, while other fuel/propulsion technologies accounted for just 2% of buses; in 2016, large diesel-

powered buses in the US accounted for 65% of all buses, CNG 19%, hybrid-diesel 12%, and other fuel/propulsion technologies 4% (NTD, 2017). Meanwhile, roughly, 8% of all buses operated by publicly-owned bus transit fleets in India used CNG in 2015 (Table 1.2).

Table 1.2. Transit bus neet numbers, india, 2015							
Type of bus fuel system	Rural + Hill	Urban	Total				
All types	117,179	25,166	142,345				
CNG powered	1,734	9,619	11,353				
CNG % of All types	1.5%	38.2%	8.0%				

Table 1.2: Transit bus fleet numbers, India, 2015

Source: CIRT (2017): "Fleet strength as on March 2015" (p.19).

Since bus transit fleets are almost invariably diesel powered, and natural gas is a poor diesel engine fuel, natural gas engines on buses are typically spark-ignited (SI), being either retro-fitted or dedicated, along with an on-board natural gas fuel system. Natural gas has a higher octane rating, and therefore enables a higher compression ratio (and fuel economy) than gasoline in SI engines, but which is still much lower than in compressed-ignited (CI) diesel engines. Furthermore, SI engines, unlike CI diesel counterparts, are characterized by poor part-load efficiencies because of throttling losses. Consequently, thermal efficiency (and related vehicle fuel economy) can be considerably lower for SI CNG engines relative to equivalent performance CI diesel engines. This is especially so in the case of vehicles on which existing diesel engines are sub-optimal, and because of other factors such as the additional weight of the CNG fuel tanks (Faiz et al, 1996).

Notwithstanding the above, optimized and technologically improved lean-burn heavyduty natural gas SI engines can in fact achieve diesel efficiencies and CO<sub>2</sub> levels (Posada, 2009). Further, the use of natural gas in engines significantly reduces black carbon (BC) emissions, which is a strong climate-forcing agent and when accounted for, can potentially enable net GHG emissions reductions -- nearly 20% GHG emissions reductions for CNG-based buses -- relative to their diesel counterparts, in contexts such as India (Reynolds & Kandlikar, 2008). Lastly, according to Rabl (2002), while CH<sub>4</sub> and CO<sub>2</sub> emissions are typically higher for CNG buses relative to an equivalent EURO II diesel engine, PM, NO<sub>X</sub> and air toxics are significantly reduced, with important social benefits (primarily due to reduced health costs) on CNG bus systems.

More broadly, with respect to the use of alternative transport fuels in bus transit systems, it should be considered that over the past couple of decades, fuel and emission standards in the US and Europe have become increasingly more stringent, which in turn has pushed fuel and vehicle technological improvements forward in terms of diesel bus emissions (Barnitt, 2008; Posada, Chambliss, & Blumberg, 2016), and thus reduced the justification for CNG bus use (Hesterberg et al., 2008). In contrast, many low- and middle-income countries continue to extensively use higher polluting diesel buses without any effective aftertreatment technologies. While more stringent diesel engine and fuel regulations have been enacted in some of these countries, it is very easy and sometimes guite common for fuels to be adulterated with additives or emission-control systems to be removed for better vehicle driving performance, thus severely compromising the emissions results sought. This is another advantage of natural gas over liquid fuels such as diesel, since natural gas adulteration is difficult (or costly) to accomplish, as opposed to liquid fuels. The consequence of diesel fuel adulteration is the adverse impact on vehicle emissions. Fuel adulteration is usually driven by the economic incentive operators have, in light of potential fuel cost savings, and is a challenge that could be present in any country context, as was the case in Japan until the early 2000s (Wagner & Rutherford, 2013). In the Indian context, the widespread availability of highly subsidized kerosene, intended as a cooking fuel, is a major challenge, given the economic incentive for adulteration of diesel with this fuel. Whether or not this is (or was) a pressing issue, fuel adulteration in Delhi is not commonly cited in the literature as a motivation (or benefit) leading to the adoption of natural gas policy that will be later discussed, but is nonetheless occasionally discussed in India (CSE, 2002a; Gandhi, 2011).

Therefore, in contexts where natural gas is accessible at lower costs and where less stringent vehicle and fuel emissions standards are in place, and where refining and supply of low-sulphur diesel not adequate to meet demand for fuel, there is likely to be more interest in CNG vehicles (Posada, 2009). Under these circumstances, cleaner fuels, such as CNG, still play a role, and can result in substantial emissions reductions and thus help in mitigating air pollution problems relative to diesel. For example, while methane emissions will be higher for CNG systems compared to diesel, particulate matter (PM) emissions will likely be considerably lower (Hesterberg et al., 2008) but, proper vehicle maintenance is important for achieving these results. Indeed, PM emissions reduction is normally among the key justifications for pursuing CNG, as was the case in Delhi, India, discussed in this dissertation.

#### 1.6 CNG in public bus transit in Delhi

Indian cities have been characterized by poor air quality since the 1990s. In Delhi, suspended particulate matter levels have exceeded World Health Organization (WHO) guideline limits almost daily since the 1990s. Levels of PM<sub>10</sub> (particulates below 10 µm diameter), which are strongly linked with respiratory and cardio-vascular illnesses and deaths, also exceed the WHO limits (CPCB, 2015). A global survey of urban air pollution (WHO, 2014a) showed that Delhi had the highest annual average levels of fine particulates (PM<sub>2.5</sub>), which pose the most serious health risk. In response to this problem, a wide range of policies has been implemented since the early 1990s to address air emissions from urban transport. Delhi being the national capital, and given its serious air quality problems, many of these policies were first implemented there and in the other major metropolitan centres, and then in the rest of the country in a phased manner. These policies have included increasingly stringent vehicle emission and fuel quality standards, vehicle inspection and maintenance (I&M) to control in-use emissions, and the phasing out of old commercial vehicles (CSE, 2002b; BIS, 2002; TERI, 2002; Kojima, Brandon & Shah, 2000). A Supreme Court of India ruling in 1998 mandated that all public and for-hire motor vehicles (buses, taxis and auto rickshaws) in Delhi be powered by compressed natural gas (CNG) (Supreme Court of India, 1998).
As a consequence of this ruling<sup>2</sup>, all of the city's urban transit buses had to be converted to run on CNG over a highly compressed time frame, by March 31, 2001. However, because of resource, logistical, and institutional challenges, CNG implementation on Delhi's buses began only in 1999-2000. Significant implementation challenges were mostly linked to the need to establish upstream and downstream fuel networks in the city, with reasonable quality and reliability of supply in addition to providing vehicles and parts. Domestic bus manufacturers were unable to supply the required numbers of conversion kits in time, due to logistical, cost, and institutional constraints, and as a result, the Supreme Court of India had no choice but to extend the original deadline for implementation of its order several times. In the process, millions of commuters were stranded on several occasions, as buses in non-compliance were forced to cease operating, and bus, taxi and M3W vehicle operators went on strike, to protest the costs of implementing the order, and difficulties such as the lack of widespread availability of CNG for refuelling (Bell et al., 2004; Kathuria, 2005; Narain & Bell, 2005). A key implementation issue related to the proverbial "chicken-and-egg" problem, discussed above, was referred to in Delhi's context as a "sequencing problem" by Bell et al. (2004). In the case of bus fleet conversion to CNG in Delhi, this sequencing problem occurred since demand for buses was dependent on the availability of financial resources for conversion as well as the availability of reliable refuelling infrastructure and vehicle technology; meanwhile, the suppliers of the new vehicle technology and infrastructure wanted assurances of demand for the technology, without which they were reluctant to invest in production, despite government legislation requiring such conversion.

Accomplishing the CNG policy mandate in Delhi required a close collaboration and coordination of key stakeholders in this process, such as vehicle manufacturers, fuel distributors and retailers, the environmental protection agency, and various government

<sup>&</sup>lt;sup>2</sup> According to the ruling (Supreme Court of India, 1998), no 8-year old buses could ply in Delhi except on CNG (or "other clean fuels") beyond April 1, 2000, and further, the entire bus fleet in Delhi was required to be "steadily converted" to run on CNG by March 31, 2001.

ministries, among others. In order to break the supply-demand vicious-cycle, bus transit operators, vehicle manufacturers, and refuelling infrastructure providers created a task force to set a timeline for cooperation and bus test trials, thus resulting in a successful transition to the new fuel system (details of the CNG implementation experience, from the perspective of Delhi's principal bus transit operator, is described in detail in Chapter 4). Notwithstanding the many difficulties, by 2004, more than 85,000 public vehicles of buses, taxis and M3Ws in Delhi were reportedly operating on CNG (Patankar & Patwardhan, 2006).

Delhi, with around 20,000 buses, has one of India's largest bus fleets; also, the publicly owned Delhi Transport Corporation (DTC) has the second largest publicly operated urban bus fleet in the country, with around 4,700 CNG buses (CIRT, 2017). DTC serves the National Capital Territory of Delhi, as well as neighbouring cities in surrounding states, and carried approximately 4 million passengers daily in 2014-15 (CIRT, 2017). CNG implementation caused a significant reduction in the capacity to deliver transit service at DTC in the initial stages of the mandated fuel transition, as noted. The first CNG buses in DTC's fleet -- the standard CNG buses -- replaced more than 3,000 diesel buses of similar configuration from around 2000 to 2004. Starting in 2007-08, DTC introduced low-floor CNG buses to replace the standard CNG buses, 10% of which were over DTC's target service age of eight years in 2006-07 (CIRT, 2008). The lowfloor CNG buses were introduced to offer improved accessibility and quality of service. Between 2007 and 2011, 3,700 low-floor CNG buses were put into operation, of which 33% were air-conditioned, on which higher fares were charged. Of DTC's operational buses in 2015-16, around 13% were standard CNG, 87% were low-floor CNG buses (GNCTD, 2017).

In conclusion, given the many issues discussed in this chapter, which outlined a rationale for CNG use in bus transit, the challenges and constraints this alternative fuel system faces, and considering the particular long-term and large-scale case of CNG use in Delhi, I have chosen in this dissertation to explore the use of natural gas-fueled urban bus fleets with a focus on the Indian context, and particularly Delhi given its

natural gas ruling. The focus of this dissertation is all the more significant and timely since the impacts of rapid motorization are shared by many other developing countries similar to India and alternative fuels are increasingly seen in India, as well as other countries, as part of the solution to a cleaner and less carbon intensive future. Alternative transport fuels have the potential to contribute to reductions in emissions of local air pollutants and energy security. As was discussed here, the central issue is that increased motorization has also raised the need for significant expansion of supporting transport infrastructure capacity, which requires massive investment expenditures in capital scarce contexts, and a need for rapid increase in supply of energy services, all of which create tremendous challenges for developing countries and therefore merit careful and comprehensive policy-relevant research.

Therefore, many important issues need to be considered when contemplating use of related alternative transport fuel technologies, such as the feasibility and cost-effectiveness of large-scale implementation and use of alternative transport fuels vis-à-vis traditional fuels, given the scale of road transport energy needs (Table 1.1) and expected continual growth in these needs, as discussed in the previous section. Furthermore, considerations have to also be contextually sensitive to issues important to low- and middle-income countries, such as technology availability, reliability, performance, serviceability, and affordability, as well as infrastructural capabilities, and government policies and regulations relating to energy, fuels and vehicles.

# **Chapter 2: Research objectives and methodological framework**

#### **Chapter overview**

In this chapter I outline the overall motivation, knowledge gaps, research needs and objectives, analytic framework, and methodological approaches used in the dissertation. The key problem motivating my dissertation research is the significant impact of transportation on urban air pollution, with serious consequences for urban populations and the environment, and the prospect of continued and fast growth in vehicle activity in developing countries, which will tend to aggravate these impacts. While Chapter 1 described this problem in broad terms, this chapter presents a rationale for public policy attention to a specific set of issues concerning compressed natural gas-fueled public bus transit fleets in India. Given the dissertation's structure, written as four independent but related substantive chapters that address different aspects of CNG use in bus transit systems, the methodology section of this chapter will discuss the overall methodological approach used in the dissertation as a whole and the four independent research projects carried out in terms of questions and tasks addressed. Detailed descriptions of the methodologies and analytic frameworks used are presented in each of the substantive analytical chapters (3 to 6).

#### 2.1 Motivation, knowledge gaps and research needs

As I argued in Chapter 1, India is an important context for the study of transportationenergy issues given the rapid growth in motorization the country continues to experience and the resulting range of negative impacts, amongst which a special cause for concern is poor urban air quality, and associated human health and environmental impacts. Since the 1990s, in response to the urban air pollution problem caused by motor vehicle emissions, substantial policy measures have been applied nationally. Also, targeted policy measures have been applied in Delhi, including the implementation of CNG in the early 2000s, since the city is the country's political capital and one of its largest metropolitan areas. The implementation of CNG in Delhi was a significant achievement not only because of the scale of conversion of its public vehicle fleet to run on an alternative transport fuel, but also the limited timeframe in which the conversion was carried out, the various competing stakeholder interests that had to be reconciled in order to get the goal accomplished, the limited logistical capacity available to build a fuel supply infrastructure for CNG (e.g., natural gas distribution and refuelling), and the lack of an established market for CNG fuel and vehicles. In 2000, there were 30 CNG filling stations supplying approximately 8,000 kg/day of natural gas, and by 2004 Delhi had a public CNG fleet of nearly 85,000 vehicles that consumed approximately 772,000 kg/day of natural gas through 124 filling stations (Patankar & Patwardhan, 2006).

For my PhD research, Delhi's CNG policy outcomes were evaluated specifically for bus transit in the city, given the importance of buses in Delhi in providing affordable transport and economic opportunity to a majority of urban passengers. Though bus transit shares have been adversely affected by personal modes of motorized transport in recent years in Delhi, public transit accounts for 43% of all trips in the city, while car and motorcycle modal share of all passenger trips are still relatively low at 19% (WSA, 2008a); this pattern is also shared by other large Indian cities (i.e., with populations > 8 million) which are characterized by a good supply of formal public transit in the form of buses, metro, and/or trains (Figure 2.1). Further, the focus of my dissertation research on CNG in bus transit systems is justified, as discussed in Chapter 1, by the fact that captive urban bus transit fleets that operate within a limited range from centralized depots are ideally suited to CNG implementation, making CNG-fueled urban bus transit worthy of policy attention. Finally, as also discussed, the bulk of urban transport PM emissions come from diesel heavy-duty vehicles, including bus transit.



*Source*: WSA (2008a): Fig. 2.9 (p.36).

Figure 2.1: Trip modal shares in India, by city size.

Broadly speaking, a large-scale conversion to an alternative transport fuel system, as was the case of CNG in Delhi, depends on a wide range of issues and must contend with many constraints that can vary significantly from context to context, such as the technology's reliability, performance and serviceability. Conceptually, Faiz et al. (1996) argue that typical constraints and issues related to alternative transport fuel adoption include: (i) higher costs associated with fuel production, storage, distribution, and vehicle purchase and operation; (ii) end-use considerations, such as changes required to fuel distribution system, marketing, availability of end-use systems or user acceptance of new alternative transport fuel adoption; and (iii) the abundance of the alternate fuel, technology and infrastructural capabilities, relative prices across the different road transport fuels, fuel safety, the technical quality of vehicles, and government policies and regulations relating to energy, fuel and vehicle taxation, and emission standards.

The literature on the implementation of CNG in Delhi's public vehicles has focused almost exclusively on its emissions effects (e.g., Kathuria, 2005; Chelani & Devotta, 2007; Narain & Krupnick, 2007; Reynolds & Kandlikar, 2008; Reynolds Grieshop, & Kandlikar, 2011); there is also literature on the emissions effects of a wider range of

policy measures (e.g., Jalihal & Reddy, 2006; Kumar & Foster, 2009; Goel & Guttikunda, 2015; Aggarwal & Jain, 2016; and Jain, Aggarwal, Sharma, & Kumar, 2016). However, little if any attention has been devoted to systematically analyzing other critical issues, such as those related to CNG's operational or financial performance; the cost-effectiveness of emissions outcomes; and the long-term financial viability of CNG systems in Delhi. In this regard, see also Khan (2015), an international review of CNG implementation in transport, which focuses exclusively on the emissions outcomes of CNG. As well, much of the literature on this topic in China also focuses on the environmental outcomes of CNG, and issues such as bus emissions estimation, total emissions inventory estimation, and well-to-wheels emissions assessments (e.g., Karman, 2006; Wang et al., 2015a; Qiu et al., 2016; Wang et al., 2018). Further, while there is limited published research related to other aspects of CNG implementation (e.g., Wang et al., 2015b; and Hao et al., 2016), it focuses mainly on issues of logistics, infrastructure requirements, and vehicle and fuel pricing, but lacks attention to the operational and financial aspects of CNG. The recent experience with CNG implementation in Pakistan highlights the importance of looking at other important issues, such as those related to fuel supply logistics and fuel costs. The government there has promoted the use of natural gas in transport, and the number of vehicles running on CNG is approximately 3 million (Khan, 2014). However, due to the importance of natural gas feedstock for other uses, such as in energy generation and industry (e.g., for fertilizer production), and since natural gas demand far exceeds supply, the government started to ration natural gas supply for CNG transport in 2010, creating immense hardship for the many transport users that rely on CNG (Khan, 2014). Furthermore, this is an ongoing problem, given the increased reliance on natural gas for the country's total energy needs on the one hand, since this fuel is important for various sectors of the economy (and not just transport), and the continued shortage of natural gas supply vis-à-vis demand, on the other (Rehman, 2019).

In the Indian context, Sen et al. (2007) point to the lack of sufficient focus on the role of decision-makers in the creation of financially viable and self-supporting urban transport systems. Further, it can be argued that, while the emissions outcomes of CNG

implementation are important from a societal perspective, and are an important focus for policy evaluation, it is also important and useful to critically evaluate this, and indeed, any such policy from the perspective of vehicle users and operators, because it is the policy responses of these actors that crucially determine the extent to which implementation, and the associated emissions reductions, actually occur and are successful.

More particularly, analyzing the operational and financial performance of bus fleets on CNG or any other alternative fuel is important because it is this performance that critically determines the bus operator's policy responses. Such an analysis is also useful from a policy perspective, given the need to provide quality, convenient and affordable bus services within the constraints of limited budgets on which public transit operators typically rely, to prevent the migration of ridership to private motor vehicles, with all of their negative impacts. In this regard, note that, while buses and other public transit modes still account for a significant share of passenger trips in Delhi, their mode share has been declining significantly, due to the growing role of personal motor vehicles.

Delhi has a population of approximately 17 million and a vehicle fleet (including motorcycles) of more than 10 million (GNCTD, 2017). Because Delhi now has many years of experience with CNG, and at such a large scale, this policy is a unique case for investigation of challenges and evaluation of outcomes related to this decision. Delhi's use of CNG shows that there are many issues related to policy implementation that are not always anticipated, but that had a profound impact on policy outcomes (CSE, 2001; Sanghi et al., 2001; Bell et al., 2004; Patankar & Patwardhan, 2006). Thus, the need for systematic and comprehensive assessment is important, especially in low- and middle-income countries where the priorities are many, governmental and transit agencies financial resources are constrained, and with lower per-capita disposable incomes of transit users, which only increase the importance of finding affordable, cost-effective, and financially sustainable solutions to transport challenges. This dissertation will use Delhi's experience with CNG bus transit in order to fulfill this important research need, especially since no systematic post-implementation studies have been conducted to

critically analyze this experience in an integrated manner covering the range of issues discussed above.

# 2.2 Research objectives

Given all of the above, the overall objective of my PhD research is to assess the operational and financial performance consequences of, and the cost-effectiveness of emissions reductions due to, replacing diesel bus fleets with CNG in Delhi, India, in order to support decision- and policy-making related to the use of alternative transport fuel systems in low- and middle-income country contexts, like India, and help inform whether this alternative transport fuel system can cost-effectively mitigate urban air quality problems. Based on this assessment, it is hoped that key lessons can be drawn for the long-term viability of a large-scale conversion to CNG in other bus transit fleets, and for informing techno-economic and environmental analyses of CNG bus transit operations. Following are the key research questions that I address in my PhD dissertation:

- What are the key lessons that can be learned from the international experience with CNG in urban bus transit fleets for rapidly motorizing low- and middle-income countries contemplating use of natural gas in order to address urban air quality, energy security or climate change concerns?

- What were the CNG implementation issues in Delhi in terms of the associated infrastructure, logistical and institutional challenges, and what was the operational and financial performance of its natural gas-fueled public bus transit fleet? Furthermore, and related to the prior question, what are the implications of CNG implementation for the ability to provide convenient and affordable public transit service?

- What are the long-term financial implications of CNG implementation in the bus transit fleet, and what are the key influencing factors? What are the lessons for the long-term viability of large-scale conversions of bus fleets to CNG?

- What was the cost-effectiveness of CNG implementation on bus public transit in Delhi? Was the conversion of the bus public transit fleet to CNG for achieving emissions reduction justified from this perspective? Further, considering Delhi and the broader Indian context, in terms of its cost-effectiveness, would CNG merit adoption today vis-à-vis cleaner diesel technological options currently available?

# 2.3 Thesis structure, analytical framework, and methodological approach

My doctoral dissertation is structured following a manuscript-based format and integrates four analytical Chapters (3 to 6), in addition to the review of issues related to alternative transport fuel use and the overall problem statement (Chapter 1), the thesis research objectives and methodological framework (Chapter 2), and a synthesis of the key conclusions and policy implications (Chapter 7). As previously indicated, two of the analytical chapters (Chapters 4 and 5) have been published in peer-reviewed journals.

Overall, each analytical chapter explores key dimensions of and perspectives on CNG implementation in Delhi's public bus transit fleet, as shown in Table 2.1, and taken together, address the above research questions. Each of the analytical chapters -- besides Chapter 2 -- starts with an overview, which provides a logical progression between chapters, and contains an introduction, literature review, methodology, research results and analysis, and conclusion.

	(Chapter 3)	(Chapter 4)	(Chapter 5)	(Chapter 6)
	Natural gas fueled	Operational and	CNG and Diesel	Cost-effectiveness
	urban bus transit: US	financial performance	Urban Buses in	of CNG
Analytical chapter:	and Latin American	of Delhi's CNG-fueled	India: A Life-cycle	implementation in
	experience and	public bus transit fleet:	Cost Comparison	the public bus
	lessons for rapidly	A critical evaluation		transit fleet in Delhi
	motorizing countries			
Poreportivo	Policy & planning /	Policy & planning /	Policy & planning /	Policy / Societal
reispective.	Bus fleet operator	Bus fleet operator	Bus fleet operator	Fullcy / Sucletal
Dimension:				
Technological	х	х	Х	
Institutional	x			
Operational /	×	×	×	
infrastructure	×	×	X	
Economic /	×	×	v	v
financial	X	*	X	X
Environmental	Х			Х

#### Table 2.1: Dimensions and perspectives reflected in dissertation

**Research projects - Description** In the first analytical chapter (Chapter 3), I conduct a comprehensive critical review of the literature concerning the performance of CNG as an alternative fuel system in urban bus transit fleets based on the experience of different fleet operators in the US and Latin America. Drawing on the key lessons from this review for CNG evaluation and the important issues raised, I analyze in Chapter 4 the operational and financial performance of DTC with CNG in Delhi to assess how the switch from diesel to natural gas affected this performance. Both these research projects, in turn, helped provide the basis for modelling the life-cycle costs of operating CNG in the Indian context, in Chapter 5. Lastly, based on the life-cycle costs estimated in Chapter 5 and coupled with my estimation of the emissions outcomes of CNG implementation in Delhi, I conducted a systematic evaluation of the cost-effectiveness of reducing key pollutants in this city in Chapter 6. Specifically, these are the four independent research projects that were carried out for each of the analytical chapters:

(i) In Chapter 3, I investigate key lessons that can be drawn from the experience related to CNG implementation in bus transit in the US, Mexico, and Chile, which was then critically analyzed in terms of the motivations for CNG implementation from the operator's perspective; the process of evaluation that was used, in terms of the criteria and issues that were considered; the choice of and rationale for the vehicle and refuelling infrastructure configurations selected; and CNG fleet performance, in terms of

parameters such as vehicle performance, fuel economy, exhaust emissions, and bus fleet and fuel system costs. The outcome of this research was to identify key challenges and draw lessons for rapidly motorizing developing countries also contemplating use of natural gas in order to address urban air quality, energy security or climate change concerns. Particularly, in this research project I sought to outline the requirements for a successful CNG fuel system implementation in bus transit fleets and how its viability can be more effectively evaluated based on the critical study of international experiences. This research also helped to identify critical issues for policy analysis, related to CNG implementation in urban bus transit, which were investigated in the subsequent substantive chapters.

(ii) In Chapter 4, I evaluate the operational and financial performance of DTC's bus fleet from 1989-90 to 2010-11 -- that is, from ten years prior to CNG implementation until 10 years after -- to assess how this performance was affected by the fuel switch, as well as the introduction of low-floor CNG buses. Key lessons from this research are drawn to see how CNG affected DTC's financial situation and as a result its overall capacity to provide transit services. I explore and raise questions on how policies, such as CNG in Delhi, may affect important trade-offs between environmental, and other transit operation, socio-economic and equity objectives.

(iii) In Chapter 5, I evaluate the life-cycle costs (LCCs) of Standard, Low-floor and Lowfloor air-conditioned (AC) CNG buses and their diesel counterparts in India. These evaluations were based on actual (on-road) bus performance data for DTC and other public bus transit agencies, in order to closely reflect fuel and bus technologies, operating conditions, and costs that are prevalent in urban India. Given the importance of fuel economy and fuel price for fuel costs, I analyze the sensitivity of the LCCs to these factors. Also, to assess what might happen if fuel prices were largely market driven, I evaluate the LCCs of Low-floor AC CNG and diesel buses in India, but with US fuel prices. Lastly, I analyze how the discount rate affects these assessments. (iv) In Chapter 6, I analyze the cost-effectiveness of Standard, Low-floor and Low-floor AC CNG buses, relative to diesel, to evaluate Delhi's natural gas directive of the early 2000s and the continuation of CNG policy in urban bus fleets in this city, given newer natural gas and diesel bus technologies available in India today. Using cost and emissions estimations, I quantify the incremental cost-effectiveness (CE) ratio of CNG and cleaner diesel buses relative to standard diesel buses in the form of dollars-peravoided emissions ratios. The CE numerator includes the total cost of ownership of each bus technology, thus reflecting incremental life-cycle costs from Chapter 5, of CNG and cleaner diesel buses over baseline Standard diesel buses. The CE denominator reflects incremental emissions reductions, based on the differences between emissions of baseline conventional Standard diesel buses and emissions of CNG and cleaner diesel buses in India. I evaluate emissions impacts in terms of PM and NO<sub>X</sub>, -- which are health critical pollutants -- and CO<sub>2</sub> equivalent emissions. In addition to the CE analysis, I also use scenarios to evaluate the broader impact of CNG policy measures on ridership across the transit system as a whole. With use of these scenarios, I evaluate how higher life-cycle costs of CNG may affect the supply of buses, and as consequence, the impact on the overall effectiveness of emissions reductions due to these incremental costs.

#### 2.4 Data and issues

Chapters 4 to 6 focus on the Indian experience, for which much information, both quantitative and qualitative, was needed and gathered during two field visits to India (Jan-Apr/2010 and May-Aug/2011). These field visits took place predominantly in the Indian capital region, Delhi, to determine (at a first instance) information availability, identify information sources, and gather data that was needed to accomplish the specific research goals of the analytical chapters; but also, the visits enabled the establishment of contacts with experts and decision makers in government, research and teaching institutions, and industry, in the subject areas that were relevant to this dissertation. In this first assessment trip, preliminary information was collected in the areas pertaining to the specific research outputs in Chapters 4 to 6 but also, in this

connection, a number of CNG-related stakeholder organizations were visited, mostly in Delhi, that allowed for the contextualization of the dissertation research in a broader policy-oriented perspective.

During the second field visit, information gathering was finalized. Detailed data pertaining to operational and financial statistics of DTC's performance was collected from secondary sources. Also, various visits were conducted at DTC to gather key operational and financial statistics that are representative of the performance of diesel and CNG use in bus transit operations in Delhi (used in Chapters 4, 5 and 6). Data on CNG implementation was also collected directly at DTC. Most information collected at DTC was quantitative, such as time-series performance statistics, collected either via access to reports from the company or via select field visits to a representative sample of bus. Details of DTC sites visited, their location in Delhi, and the typical depot layout can be found in Appendix A. Descriptions of methods and data collection are also provided in each analytical chapter (4 to 6).

Lastly, it should be noted that, while I reviewed a wealth of (quantitative and qualitative) information for this dissertation research and that most of the data required to achieve specific research objectives was available for the context of interest (India), in certain cases data were not available and had to be estimated. Since collecting these data through primary data collection methods was beyond the intended scope of individual research projects or even for the dissertation as a whole, as discussed in each analytical chapter, in these circumstances, parameter estimation was based on best available evidence for similar contexts, for which details of the assumptions used are clearly outlined and referenced in the methodological sections of each analytical chapter.

# Chapter 3: Natural gas fueled urban bus transit: US and Latin American experience and lessons for rapidly motorizing countries

#### **Chapter overview**

As discussed in preceding chapters, India mandated the implementation of CNG to mitigate urban air pollution problems. However, carrying out such a mandate requires massive investments in vehicles, fuel systems, and refuelling infrastructure, and therefore presents a substantial challenge (or even barrier) to CNG adoption, especially from a transit operator's perspective, if they have to bear these costs. In addition to these implementation challenges and costs, there are also important operational issues regarding the use of CNG in providing transit services, such as vehicle performance on the road -- including its reliability -- emissions performance, and the operating costs, among others. Specifically, analyzing the operational and financial performance of bus fleets on CNG is important because this performance will be a key determinant the operator's policy responses. In Delhi's case, no systematic post-implementation study has been carried out, in terms of these range of aspects, as most research to-date has focused largely, if not exclusively, on emissions outcomes. Most importantly, as will be shown in later chapters, despite CNG implementation, Delhi urban air quality has been deteriorating significantly, all of which raises important questions, such as the effectiveness of CNG in reducing pollution; the cost-effectiveness of CNG measures; if the same outcomes, and other objectives, could have been achieved by other means; and, how viable CNG is, given investments and the socio-economic context. All these are important questions, especially when considering the issues discussed in Chapter 1, such as existing barriers to alternative fuel implementation, the proverbial "chicken and egg situation" to initiate these policies, and the fact that even now alternative fuels only account for 5-6% of all fuels consumed in road transport globally. With respect to alternative transport fuel use, even though the share of natural gas is small relative to petroleum, CNG propulsion system is the most prevalent alternative to diesel (particularly for urban public bus transit fleets).

In view of the foregoing, the objective of this chapter is to investigate the lessons that can be learned from international experience with CNG use in urban bus transit fleets, and to explore the particular relevance of these lessons for rapidly motorizing low- and middle-income countries which are implementing, or might wish to implement, this fuel for addressing urban air quality concerns. To this end, this chapter comprises a critical review of the performance of CNG based on the experiences of five urban bus transit fleet operators in the US, Mexico, and Chile, in order to see how CNG implementation can be more effectively evaluated, with respect to the bus fleet operator's financial, emissions, and operational goals, and what is required for this implementation to be successful.

#### 3.1 Methodology and outline

Three US and two Latin American CNG urban bus transit fleets are used for the analysis (Table 3.1). While the fleets in New York (NYCT), Palm Springs (SunLine Transit Agency) and Washington D.C. (WMATA) were chosen because they have been operational for a decade or more, the Mexican (MCMA) and Chilean cases were chosen for being part of a limited group of countries in the global South in which alternative fuel systems, including CNG, were evaluated for application in urban bus fleets. These cases are relevant to CNG in the Indian context, the primary focus of my dissertation, because the CNG bus technologies implemented in the US and Latin America that I discuss are similar to the bus technologies that were implemented in India in the early 2000s, which will be analyzed in Chapters 4, 5 and 6. More importantly, I will use the US and Latin American experiences to draw lessons for the evaluation and implementation of CNG in the Indian context. For each of the US and Latin American cases, the following issues are critically discussed: the motivation for CNG implementation from the bus fleet operator's perspective; the process of evaluation that was conducted, in terms of the criteria and issues that were considered; the choice of and rationale for the vehicle and fuelling infrastructure configurations selected; and CNG fleet performance, in terms of parameters such as vehicle performance, fuel economy, air pollutant and GHG emissions, and bus fleet and fuel system costs. Finally,

note that these cases were selected also because of the comprehensive, published literature on the assessments conducted therein, along the range of dimensions that I am focusing on.

Although this chapter focuses on the performance of CNG relative to diesel as a fuel for urban bus transit, evaluations of diesel-hybrids are also discussed, because they have been implemented, in some instances for quite a while and to a considerable extent, including in the cases discussed here, and will likely become more prevalent in the future, including in low- and middle-income countries like India. Also, note the focus on comparisons between different fuel systems on similar configuration buses under similar operating conditions in the US fleets, with additional comments on the performance of these systems in the Mexican case, in which these conditions are different (for example, while the absolute capital and operating costs for the different fuel systems are lower for MCMA relative to the US fleets, useful cost comparisons can still be made among these systems).

CA	SE	No. of CNG buses <sup>(1)</sup>	% of fleet	1 <sup>st</sup> adoption
1.	New York City Transit, New York, NY (NYCT)	481	12%	1995
2.	SunLine Transit Agency, Palm Springs, CA (SunLine)	66	96%	1994
3.	Washington Metropolitan Area Transit Authority (WMATA)	459	31%	2002
4.	Mexico City, Mexico (MCMA)	30	n/a	evaluation <sup>(2)</sup>
5.	Santiago, Chile	0	0	evaluation <sup>(3)</sup>

 Table 3.1: Summary of cases evaluated

<sup>(1)</sup> Fleet numbers source as follows: NYCT (NYCT, 2011), Sunline (SunLine, 2011), WMATA (WMATA, 2010), MCMA (RTP, 2010), Santiago (Universidad de Chile, 2007); <sup>(2)</sup> Project: climate-friendly measures in transport; <sup>(3)</sup> Project: sustainable transport and air quality

Finally, while this chapter seeks to learn from the international experience with CNG in urban bus fleets, there is of course the more fundamental question of whether CNG is worth implementing in urban transport, to address urban air quality, energy security and climate change concerns, specifically in the LIC context (and whether urban transport is the best sector in which to implement CNG, given the other important sources of air pollution, and related health and welfare concerns). While these larger questions are beyond the scope of this chapter, CNG, diesel and diesel-hybrid urban buses were

evaluated in terms of operating costs, based on fuel price sensitivity analysis of selected cases. The conclusion contains a summary of the key lessons from the experience in the five cases, and the results of the sensitivity analysis.

#### 3.2 Analysis of CNG experience

#### 3.2.1 Rationale for CNG implementation in urban bus transit

In the US, transit operators have been under pressure, from the public and the EPA, to reduce bus emissions in urban areas; EPA have progressively tightened emissions requirements for urban transit buses, particularly for NOx and PM, the pollutants of primary concern, since 1990 (e.g., see Barnitt, 2008; EPA, 2011). For the WMATA, the rationale was primarily to reduce NOx emissions, given Washington DC's ozone problems (Melendez et al., 2005). In NYCT's case, they created their Clean Bus Program in 1992, and stopped buying new diesel buses in 1999. NYCT introduced CNG through a small pilot project in 1995, as part of a clean fuel bus program, with a fleet of 486 CNG buses in two bus depots in 2011 (NYCT, 2011). Since 2000, they decided to purchase only low emission buses, including those operating on CNG, and -- motivated by fuel economy and operating costs as well -- diesel-hybrids, besides other actions (Barnitt, 2008). It may be that funding, such as by that from the US Department of Energy, was also a motivating factor for conversion to CNG.

In Mexico, the Third Air Quality Management Plan prepared by state and city authorities identified the significant role of the transport sector in air pollution and the likely costeffectiveness of focusing on this sector to improve air quality and minimize health impacts (SMA, 2006b; World Bank, 2002a; World Bank, 2002b). A comprehensive strategy for addressing vehicular air pollutant and GHG emissions, including the evaluation of clean fuels, was therefore developed. In Santiago, the environmental and economic impacts of clean bus technologies were evaluated, and, as in Mexico, this effort was intimately linked to both the city's urban transport plan and the metropolitan region's Air Pollution Prevention and Clean-up Plan that sought to reduce PM10 and NOx emissions from transport (OECD, 2005; SECTRA, Plan de Transporte Urbano para el Gran Santiago 2000-2010). In both cities, the CNG evaluation was conducted as part of a comparison of different fuel and vehicle technologies to improve urban air quality and reduce greenhouse gas (GHG) emissions. Pre-evaluation and field tests for CNG fuel systems in both cities were funded by the Global Environment Facility (GEF), in conjunction with local government sources (Graftieaux et al., 2003).

# 3.2.2 Evaluation of CNG performance on urban bus transit fleets

It is important to highlight the different objectives and contexts of the evaluations conducted in the five cases reviewed. In the three US cases, performance evaluations were sponsored by the US Department of Energy's (USDoE) Advanced Vehicle Testing Activity (AVTA) programme to assess advanced propulsion technologies, including CNG, based on the actual experience and costs of selected bus transit operators (Barnitt, 2008). While the Mexican and Chilean cases were both pre-implementation evaluations of CNG and other alternative fuel systems for buses, field tests were actually conducted for determining their feasibility under city operating conditions in Mexico City. The evaluation of CNG in Santiago, Chile only relied on desk studies, despite initial plans for field tests. Nevertheless, this evaluation is also useful to look at for the criteria that were considered. Finally, note that though the evaluations analyzed in this chapter are from the mid-2000s, for buses of vintages ranging from mid-1990s to mid-2000s (Table 3.2), they are very relevant for the purpose of this thesis since these buses, with corresponding emissions standards, are still pertinent and in use in countries like India<sup>3</sup>. Furthermore, it is for these vintages of vehicles that long-term performance information and insights are available in terms of fuel economy, reliability and operating cost parameters, and how performance changes over longer periods of time, in these evaluations.

The purpose of the in-use performance evaluations of alternative fuel and advanced technology vehicles in bus fleet applications sponsored by USDoE's AVTA programme -

<sup>&</sup>lt;sup>3</sup> For example, in India, only in 2017 was Bharat Stage IV (BS-IV) emission standard applied nationally (Posada et al., 2016).

- including those conducted by Barnitt and Chandler (2006) and Barnitt (2008) at NYCT, Chandler (2006) and Chandler and Eudy (2007) at SunLine, and by Chandler, Eberts and Melendez (2006) and Melendez, et al. (2005) at WMATA, which are the studies used in this chapter -- was to "provide comprehensive, unbiased evaluations" in order to enable "fleet owners and operators make informed purchasing decisions" regarding these technologies (Barnitt & Chandler, 2006); therefore, these assessments should be of interest to any bus fleet operator considering a fuel system change. The AVTA initiative was carried out by the National Renewable Energy Laboratory (NREL), and followed standardized procedures of data collection and analysis developed there (NREL, 2002; Barnitt, 2008). It was a pre-requisite for all buses being assessed to have been in operation for at least a year in their fleets.

The evaluation of the bus fleet performance referenced above at NYCT focused on the key issues of reliability, fuel economy, vehicle maintenance, and vehicle and fuel infrastructure capital and operating costs, for a statistically significant sample of CNG and hybrid buses during a 12-month period during 2004-05 (Barnitt & Chandler, 2006; Barnitt, 2008). Data was collected for the various parameters which importantly influence performance outcomes, listed in Table 3.2. While the evaluation frameworks in the assessments of the CNG bus fleets at WMATA and SunLine referenced above essentially included identical parameters, the time period of evaluation was 2000-04 for SunLine, and 2003-04 for WMATA. In WMATA's case, the performance evaluation was supplemented with a measurement of emissions from lean burn CNG and ultra-low sulphur diesel (ULSD) buses on a chassis dynamometer (Melendez et al., 2005).

While the evaluations of the US bus fleets conducted by Barnitt (2008), Barnitt and Chandler (2006), Chandler (2006), and Chandler et al. (2006) considered infrastructure capital costs, and vehicle capital and operating costs (including fuel and maintenance costs), as well as factors such as miles between road calls, all of which are helpful in analyzing the performance of alternative fuels, these parameters were assessed at a given point in time; the evaluations did not include an assessment of life-cycle costs, which the AVTA programme did not require. Consideration of the costs (and other

impacts) related to the purchase, operation, maintenance, servicing and replacement of engines, vehicles, and fuelling infrastructure (or in the case of hybrids, batteries and battery conditioning stations) on a life-cycle basis is important for assessing the long-term viability and cost-effectiveness of, and therefore enabling informed fleet operator decision-making related to, alternative fuel systems, which is after all the primary objective in conducting such evaluations. These factors are particularly important, given that the challenge for transit operators considering an alternative fuel to mitigate vehicular emissions is that that fuel must be proven to be superior to cleaner diesel options combined with exhaust treatment, which do not require major changes or incremental costs to operator's refuelling infrastructure. Indeed, exhaust gas treatment technologies such as diesel particulate filters (DPFs) and exhaust gas recirculation (EGR) have already been implemented in conjunction with ULSD, in order to meet increasingly stringent EPA emission standards for PM and NO<sub>x</sub> (Barnitt, 2008; EPA, 2011).

Also, note in this regard that incremental costs associated with the purchase of new CNG buses compared to diesel buses can range from \$25,000 to \$50,000 in OECD countries (IEA, 2002) and approximately \$40,000 in the US (Lowell et al., 2007). Most of the incremental vehicle costs, whether for retrofitting or buying new vehicles, relate to the on-board CNG tanks. Yang et al. (1997) estimate that these tanks can represent two-thirds of total retrofit costs. Furthermore, as discussed in Chapter 1, CNG fuel tank technology has important implications for vehicle performance factors such as payload, fuel economy, refuelling frequency and range, as well as for the refuelling infrastructure. Natural gas compression speed is one of the major factors in determining total refuelling infrastructure costs (Heath et al., 1996). For an urban bus fleet of 100 buses using a fast-fill system comparable to diesel in an indoor refuelling facility, additional costs can amount to \$2 million, which represents an incremental cost of \$20,000 per bus just for the refuelling infrastructure (Heath et al., 1996; ARCADIS, 1998; Lowell et al., 2007).

In the Mexican case, following on from the Third Air Quality Management Plan discussed earlier, field tests were conducted to evaluate the performance of various

alternative bus fuel systems, under real operating conditions, for the overall purpose of assessing the cost-effectiveness of each option in mitigating air pollutant emissions in Mexico City's Metropolitan Area (MCMA) (Vergara & Haeussling, 2007; SMA, 2006a). The field tests included a total of 22 buses of varying dimensions, emission standards and fuel systems, including lean burn CNG, diesel-hybrid, and diesels, with the diesels using a fuel of varying sulphur content ranging from the standard 350 to 15 ppm (parts per million), and additional emission controls (specifically, catalytic DPFs). Note that fuel cells and LPG were eliminated from the field tests, the former because of capital cost considerations, and the latter due to fugitive emission concerns. Of interest for this chapter are the two CNG bus models supplied for the field tests by Ankai from China and Busscar from Brazil (Table 3.2) and the two diesel buses (Volvo 12 and Mercedes Benz 12) from Mexico City's operating fleet of buses at RTP (Red de Transportes de Pasajeros del D.F.), that were used as a baseline with which to contrast the emissions and operating performance of the alternative fuel systems. The main issues that were evaluated included bus and engine performance, fuel consumption and economy, capital and operating costs, noise, and PM, NOx, CO, HC, VOC and CO<sub>2</sub> emissions (SMA, 2006a). The evaluation also included on-board vehicle emissions measurements conducted by Weaver and Almanza (2006) along a BRT (Bus Rapid Transit) test route for a year during 2004-05, and chassis dynamometer tests in 2004 conducted by Western Virginia University (WVU, 2005). Note, however, that no particular attention was given to refuelling infrastructure requirements or costs.

For Santiago, while the GEF funded project specifically required that CNG buses be field-tested (World Bank, 2003; Graftieaux et al., 2003) in order to integrate a thorough environmental-economic assessment of them as part of transportation decision-making (Deuman, 2003; Universidad de Chile *et al.*, 2003; Universidad de Chile, 2007), only desk surveys were conducted to evaluate CNG. Data for the desk studies came from two sources, prior field tests conducted in 1999 in Santiago by the Chilean Ministry of Environment Commission, results of which are available in "Anexo C.2" of Universidad de Chile (2007), and the Mexican field tests discussed in this chapter (SMA, 2006a).

To date, there is no evidence of CNG adoption in Santiago (AGN, 2016), though the public transit system operator, Transantiago, was considering this fuel system closely and looking for adequate public funding for a potential project as recently as 2010 (NGV Journal, 2010). In the case of Santiago, which had sought to promote CNG, an ineffective incentive structure for private fleet operators prevented the technology from being adopted; a group of experts critiqued the fuel tax structure, which favored diesel over CNG (MTT, 2008). MTT (2008) argued that distortions in the tax structure should be eliminated in order to promote a more efficient transport system, by incorporating all externalities of each fuel option in the decision-making process. According to MTT (2008), the transport fuel tax structure gave no incentives for fleet operators to explore lower emission technologies and so was mostly unsuccessful because of the low economic feasibility of the proposed changes. Meanwhile, in Mexico City, the public transit operator RTP introduced 30 CNG buses in November of 2010, at the approximate cost \$270,000 per bus (RTP, 2010). An interesting aspect of the Mexico City CNG project is that while RTP has not yet invested in a dedicated CNG refuelling infrastructure, which is currently being provided by a third-party supplier outside the bus depot, they are considering such infrastructure if CNG adoption were to gain "scale" (RTP, 2010; NGV Journal, 2011a, 2011b).

# 3.2.3 Vehicle and fuelling infrastructure

The space needed for high-powered compressors and buffer tanks that ensure fast refuelling of buses can be quite substantial; additionally, compliance with building and fire safety codes has to be considered in densely populated areas. At both NYCT and WMATA, CNG buses are operated from two carefully selected bus depots, since not all depots are suitable given these considerations (Chandler et al., 2006; Barnitt, 2008). Despite implementing CNG, WMATA has maintained a diversified approach, by including standard diesel buses (55% of the bus fleet), clean diesel buses that combine ULSD with DPFs (8%) and diesel-hybrids (6%) along with CNG buses (31%) in its fleet in 2010 (WMATA, 2010). Note that while this data relates to buses used in the dates described, these technologies are still widely available and used today in contexts such as India, and which will be analyzed in detail in the following chapters.

Similarly, CNG buses represented 12% of NYCT's fleet (Table 3.1), with other technologies, particularly diesel-hybrids, also being used to achieve its air quality objectives. Note that this diversified approach, and the use of hybrids in particular has been motivated not only by the infrastructure and space constraints discussed above, but also by the fact that hybrids offer significant fuel economy advantages relative to both diesel and CNG, while also producing low emissions. In the case of SunLine, it opted for a full fleet conversion in 1992 and 46 new CNG buses were put into service in 1994. In 2011, SunLine had 66 CNG and 3 hydrogen fuel-cell powered buses (SunLine, 2011).

NYCT's two depots operating CNG buses had 163 and 318 buses running on this fuel respectively in 2011 (NYCT, 2011). The capital cost for the refuelling station and facilities improvement at the first depot was \$7.4 million, of which \$2 million was spent on blasting through solid rock in order to cover high-pressure gas lines (Barnitt, 2008). WMATA's nearly 400 CNG buses were equally distributed between two separate facilities. The costs for one of these facilities was \$15.6 million (Chandler et al., 2006), but most of this amount is likely not fuel related, as the 40-year-old facility had to be upgraded. The cost for the refuelling infrastructure at this WMATA depot alone was \$4 million. At SunLine, meanwhile, no modifications were necessary in their outdoor bus storage facility, showing that CNG refuelling infrastructure investments are climate and space dependent.

All three US fleets outsourced the operation of their refuelling infrastructure. NYCT has a fast-fill system which allows buses to be fueled in 3-10 minutes; its operation and maintenance was outsourced at a variable cost of \$0.25 per therm of natural gas consumed, which amounted to 25% of fuel feedstock price. WMATA also outsourced fuel infrastructure operations, but at a fixed rather than at a variable rate, in addition to paying the related electricity charges. Interestingly, WMATA's refuelling costs also amounted to roughly 25% of feedstock fuel price. Given the magnitude of refuelling costs, operators should consider, in addition to the natural gas price, the refuelling infrastructure setup, operating and maintenance costs, electricity charges, and the modalities and costs of out-sourcing.

As for the natural gas engines on the buses, the key specifications are summarized in Table 3.2, for each fleet. Engine details are also provided for diesel and diesel-hybrid buses operated concurrently by them. Note that the power rating for the NYCT diesel-hybrid buses is for the motor driving the wheels, not for the engine. In MCMA's case, as many as 22 different kinds of bus models were evaluated on both road and chassis dynamometer tests. Comparison with the US cases is challenging, given that most of the MCMA test vehicles were relatively new and had been supplied by OEMs for testing purposes. To ensure comparability, only road test data for two each of the diesel and CNG vehicles, which had similar characteristics to the US vehicles (based on passenger capacity, and vehicle and engine size), have been included in the analysis.

While engine power ratings were similar -- or marginally higher -- for CNG relative to diesel buses of the same vintage, power ratings appear to have increased since the 1990s, for engines operating on both fuels. As importantly, from the point of view of fuel economy, the compression ratios of the (spark-ignited) CNG engines were of the order of 10:1, going up to 11:1 over the years, as against around the typical 16:1 for the compression-ignited diesel engines. Again from the point of view of fuel economy, note that the curb weight was higher by around 10% and 12% respectively for the CNG and diesel-hybrid buses relative to their diesel counterparts at NYCT; however, these differences might be due in part to both the CNG and diesel-hybrid buses, but not the diesel bus, being low-floor vehicles. Further, note that the passenger capacity was lower for the CNG bus, and even lower for the diesel-hybrid bus, relative to the diesel bus, perhaps to maintain a similar gross vehicle weight for the two alternative propulsion systems. At WMATA, interestingly, while the curb weight was only marginally higher for the CNG bus, the passenger capacity and the gross vehicle weight were considerably lower, than for the diesel bus; it is not clear why this is so.

		NYCT		WMATA			SunLine MCMA		/IA <sup>(1)</sup>
	Diesel	CNG	Hybrid	Diesel	CNG	CNG	CNG	Diesel	CNG
Model Year (MY)	1994 & 1999	2002	2002	2000	2001	2004	1994	2004	2004
Fuel Type	D30	CNG	D30	D15	CNG	CNG	CNG	D15	CNG
Manufacturer / Model	Orion V (high floor)	Orion VII (low floor)	Orion VII (low floor)	Orion VI (low floor)	NewFlyer (LF) C40	NewFlyer (LF) C40	Orion V (high floor)	Mercedes & Volvo	Busscar + Ankai
Engine (model / max HP)	DDC S50	DDC S50G	Cummins ISB	DDC S50	CWI C Gas Plus	Deere 6081HFN04	Cummins L10-240G	OM924-LA & D7C	Cummins B5.9-230G
	275	275	270	275	280	280	240	280 & 300	230
Compression ratio	16.5:1	10:1	16.3:1	15:1	10.1:1	11:1	N/A	19.5:1 <sup>(3)</sup>	10.5:1
No. of cylinders	4	4	6	4	6	6	6	6 <sup>(3)</sup>	6
Engine displacement	8.5L	8.5L	5.9L	8.5L	8.3L	8.1L	N/A	7.3L <sup>(3)</sup>	5.9L
Air system		Turbo	ocharged		Turbocharged			Turbocharged <sup>(3)</sup>	
Emissions control / Standards	Retrofit DPF (Johnson Matthey)	None	Engelhard DPX	Engelhard DPX	Oxidation catalyst	Oxidation catalyst	N/A	EPA1998 & EUROIII	EPA'04 + 2WC
Fuel Capacity (litres)	473	473 <sup>(2)</sup>	379	473	606 <sup>(2)</sup>	606 <sup>(2)</sup>	473 <sup>(2)</sup>	300 <sup>(3)</sup>	480 & 580 <sup>(2)(4)</sup>
Curb Weight (metric tonnes)	12.9	14.2	14.4	13.3	13.6	13.6	13.4	10.2	10.5
Gross Vehicle Weight (GVW)	18.1	19.3	19.3	19.3	18.4	18.4	18.4	16.0	15.9
Payload	5.2	5.1	4.9	6.0	4.8	4.8	5.0	5.8	5.4
Length (metres)	12.2	12.2	12.2	12.2	12.2	12.2	12.2	12.0	11.2
Passenger capacity	75	73	70	88	60	60	67	90	91
Bus purchase cost (\$ '000)	290	313	385	300	340	340	225	<b>74</b> <sup>(5)</sup>	134 <sup>(5)</sup>
Average operating speed (km/h)	10.2	10.2	9.9	18.7	18.7	18.7	23.5	Avg.(slo Avg.(fas	w)=17 to st)=65 <sup>(6)</sup>

# Table 3.2: Summary of selected bus system specifications

<sup>(1)</sup> Various models were used in MCMA evaluations. CNG buses include Busscar & Ankai models. Diesel includes Volvo 12m D7C engine and Mercedes Benz 12.4m OM924-LA engine model. The two diesel buses were selected to match passenger capacity of two selected CNG buses.

<sup>(2)</sup> CNG fuel capacity expressed as diesel energy equivalent litres.

<sup>(3)</sup> For diesel buses, data refers only to Volvo 12m bus.

<sup>(4)</sup> 480L for Busscar model; 580L for Ankai model.

<sup>(5)</sup> Dollar amounts converted from Mexican pesos.

<sup>(6)</sup> Average slow and fast speeds: route speeds varied from 17km/h near commercial/industrial areas to 65km/h (Weaver & Almanza, 2006). Sources (NYCT, WMATA & SunLine converted from imperial system): NYCT (Barnitt & Chandler, 2006); SunLine (Chandler, 2006); WMATA (Chandler et al., 2006); MCMA (SMA, 2006a).

# Table 3.3: Summary of selected performance parameters of buses

	NYCT				WMATA			MCMA <sup>(1)</sup>		
	Diesel	Diesel	CNG	Hybrid	Diesel	CNG	CNG	CNG	Diesel	CNG
Number of buses evaluated	9	9	10	10	5	5	5	31	2	2
Model Year (MY)	1994	1999	2002	2002	2000	2001	2004	1994	2004	2004
Fuel economy (km/l) <sup>(2)</sup>	0.97	1.01	0.72	1.36	1.21	0.99	1.02	1.42	2.32	1.78
Fuel price (\$/diesel litre equivalent) <sup>(2)</sup>	0.47	0.47	0.46	0.47	0.35	0.35	0.35	0.29	0.67	0.42
Total operating costs (\$/km)	N/A	N/A	1.44	1.11	0.66	0.68	0.71	0.44	0.33	0.28
Fuel costs (\$/km)	0.48	0.47	0.64	0.35	0.29	0.35	0.35	0.22	0.29	0.25
Maintenance costs (\$/km)	N/A	N/A	0.80	0.76	0.37	0.32	0.36	0.22	0.04	0.03
Engine and Fuel system costs (\$/km)	N/A	N/A	0.19	0.23	0.08	0.08	0.08	N/A	N/A	N/A
Reliability (kms between roadcalls)	4,275	9,213	9,235	8,323	5,223	7,709	8,449	21,185	N/A	N/A
Engine and fuel system related roadcalls) <sup>(3)</sup>	8,794	19,743	15,285	13,079	20,458	28,909	29,573	N/A	N/A	N/A

<sup>(1)</sup> Average performance of select models for three sets of field tests conducted at MCMA (Tables 7.2 & 7.3 from SMA, 2006a). All dollar figures were converted from Mexican pesos. Diesel buses were selected to match CNG passenger capacity: Two CNG buses (Busscar & Ankai); Two diesel buses (Volvo 12m with D7C engine; Mercedes Benz 12.4m with OM924-LA engine).

<sup>(2)</sup> CNG calculated on an energy equivalent basis of diesel litres.

<sup>(3)</sup> Engine and fuel system related breakdowns include engine, fuel system, exhaust, non-lighting electrical, air intake and cooling, but not transmission; for hybrid, electric propulsion is also included.

Sources (NYCT, WMATA & SunLine converted from imperial system): NYCT (Barnitt & Chandler, 2006); SunLine (Chandler, 2006); WMATA (Chandler et al., 2006); MCMA (SMA, 2006a).

## Table 3.4: WMATA chassis dynamometer-based emissions, 2000-2004 vintage buses, CBD drive cycle

Vehicle	No. of buses	СО	NOx	NO	CH <sub>4</sub>	NMHC**	PM	CO <sub>2</sub>	FE
Configuration	tested	(g/km)	(g/km)	(g/km)	(g/km)	(g/km)	(g/km)	(g/km)	km/l***
MY 2000 / DDC S50 w/ DPF (DIESEL)	2	0.12	15.29	11.00	0.00	0.00	0.01	1,963	1.30
MY 2001 CNG–CWI w/ oxidation catalyst (CNG)	4	0.34	11.81	9.76	10.75	0.68	0.01	1,403	1.28
MY 2004 DDC S50 w/ EGR and DPF (DIESEL)	3	0.21	11.12	7.21	0.00	0.00	0.02	2,079	1.23
MY 2004 CNG–Deere w/ oxidation catalyst (CNG)	3	0.09	5.64	4.78	6.59	0.34	0.00	1,350	1.34

CBD = Central Business District Drive Cycle

\*\* Non-methane hydrocarbons; THC for Diesel buses

\*\*\* km per energy equivalent diesel litre. Fuel economy was computed using a carbon balance, fuel properties, and measured emissions data. The carbon compounds (CO<sub>2</sub>, CO and HC) emitted in the exhaust were measured, and the fuel consumption was calculated using a carbon balance equation.

Source: Average testing results per vehicle model converted to metric, Table 7 in Melendez et al. (2005).

As for emissions ratings, presumably, the engines in the various cases shown in Table 3.2 for NYCT, WMATA and SunLine conformed to the EPA transit bus emissions requirements that applied to the relevant model year, or even a later year. Note that the EPA emission requirements for transit buses, which have been tightened since the early 1990s, became especially stringent for NOx and PM since around the mid-1990s (Barnitt, 2008; Lowell & Kamakate, 2012). Further, the EURO requirements for these two critical pollutants have generally lagged behind the EPA's (on the other hand, the EURO requirements have been considerably more stringent than EPA's, for CO) (Lowell & Kamakate, 2012; Posada, 2009). In any case, the increasingly stringent EPA requirements for PM and NOx since 1996, and even more particularly, those for 2007-2010, required DPFs, and ULSD (30 and 15 ppm sulphur since 1998 and 2006, respectively), in addition to EGR, on diesel buses (Barnitt, 2008). Finally, it appears (Table 3.2) that CNG requires simpler (and less expensive) emission control measures to achieve these stringent requirements than diesel (or even diesel-hybrids).

The CNG buses in the US cases (Table 3.2) vary considerably in relation to their cost. NYCT's 73-passenger capacity CNG bus was approximately 8% more expensive than a comparable baseline diesel bus, but 19% less so than a diesel-hybrid produced by the same manufacturer in the same year. At WMATA, CNG buses cost 13% more than a diesel bus of comparable dimensions and weight in its fleet, but also 8% more than NYCT's CNG models of equivalent age. The incremental cost for CNG buses relative to their diesel counterparts at WMATA and NYCT was very close to the 9% average for US bus transit operators (Lowell et al., 2007). Besides, note the lower passenger capacity on the CNG buses.

# 3.2.4 Vehicle and fuel systems performancei) Fuel economy

Natural gas characteristics have a significant impact on vehicle performance. CNG highpressure storage tanks incrementally increase a vehicle's curb weight by up to 1.4 metric tons, or by roughly 10% of a vehicles curb weight, depending on materials used, and reduce its payload capacity (ARCADIS, 1998; Barnitt & Chandler, 2006; Chandler et al., 2006). In turn, the increased vehicle weight negatively affects fuel economy and driving range in comparison to diesel, while also requiring buses to have their suspensions modified, with possibly increased road damage. While the Otto cycle, on which the spark-ignited (SI) CNG engines operate, has a higher thermal efficiency than the diesel cycle at the same compression ratio, diesel engines use much higher compression ratios in practice (Table 3.2), because they compress only air, and are thus not susceptible to auto-ignition (as in the case of SI engines), and are therefore able to achieve higher thermal efficiency relative to CI diesel engines. Further, SI CNG engines have lower part-load efficiency relative to CI diesel engines, due to throttling losses, also with negative implications for CNG fuel economy. All of these factors combined, lower the potential fuel economy performance of CNG buses, on a per-unit energy basis, relative to comparable diesel buses.

Under similar conditions of vehicle age, weight and speed, the CNG buses were 29% and 18% less fuel economical, on an equivalent energy basis, than their diesel counterparts at NYCT and WMATA respectively (Table 3.3). Note that the same EPA emission standards applied in all these cases, since they did not change between 1998 and 2003; further, while the GVW was slightly higher for CNG relative to the diesel bus at NYCT, the situation was reversed at WMATA; and finally, while the engine platforms for the two fuels differed in the case of WMATA, they were the same in the case of NYCT. At MCMA, the two CNG buses (Ankai and Busscar) on average had a 34% lower fuel economy than their two diesel counterparts (Volvo 12 and Mercedes Benz 12) of the same 2004 vintage (Table 3.3).

The significantly better fuel economy for the diesel buses was despite a higher level of exhaust control to achieve the same emission standard (even the 2004 vintage CNG engines at WMATA required only an oxidation catalyst, whereas even the 1994 diesels at NYCT called for particulate filters), and a higher GVW than for the CNG buses at WMATA, and in the case of NYCT, despite being of much older vintage relative to CNG. Even the more advanced 2004 vintage CNG buses had a fuel economy 16% lower than the 2000 vintage diesel buses at WMATA. On average, the fuel economy of CNG buses

was reported to be 25% lower than for diesel buses in the US (Lowell et al., 2007). This significant fuel economy gap may in part be due to the characteristics of the SI cycle on which the CNG engines operate, as discussed earlier.

While engine type and vintage are undoubtedly important factors, operating conditions make a significant difference to fuel economy for both diesel and CNG vehicles<sup>4</sup>. The fuel economy figures were higher -- by 20% for the diesel buses of similar vintage, and by as much as 38% for the CNG buses of similar vintage -- at WMATA relative to NYCT, which had lower average speeds than WMATA. Note that the NYCT driving cycle is characterized by significantly lower speeds, significantly more stops per kilometre, and more acceleration and deceleration, relative to that at WMATA (Nylund et al., 2007). At SunLine, the CNG buses, which operated at a much higher average speed than those in the other two fleets, had a much higher fuel economy, despite their 1994 vintage, than the diesel buses operated by NYCT and WMATA, and even the hybrid bus at NYCT.

So, it appears that higher average speeds significantly improve performance for both diesel and CNG buses, but much more for CNG buses; this is likely because the SI cycle on which the CNG engines operate performs poorly at low speeds and loads, as in the case of the NYCT bus cycle, which is characterized by low average speeds (below 10 km/h) and frequent stop-and go operation. As well, therefore, improved operating conditions narrow the fuel economy gap between diesel and CNG. This effect (and the importance of operating conditions, more generally) is further borne out by the chassis dynamometer tests conducted on the WMATA buses by Melendez et al. (2005) (results from this study is summarized in Table 3.4, above), which in several instances were performed on the same buses as in the on-road tests (Chandler et al. 2006, reported in Table 3.3). Considering only the buses common to these two tests, the fuel economy performance on the dynamometer tests was consistently higher than on the on-road tests, for both fuels. However, the fuel economy improvement on the

<sup>&</sup>lt;sup>4</sup> For example, see comparative fuel economy performance of buses under different city driving cycles in Nylund et al. (2007).

dynamometer tests was considerably higher (30%) for CNG than for diesel (only around 7%), suggesting that CNG-SI engines are far more sensitive to, and are more affected by, changes in operating conditions than are diesel engines.

Similarly, the fuel economy for both diesel and CNG buses at MCMA was significantly higher than for their counterparts at NYCT as well as WMATA, and even the hybrid at NYCT, demonstrating the important role of higher operating speeds but also lower vehicle weight. Further, the fuel economy gap for CNG relative to diesel buses at MCMA was significantly lower on the dynamometer tests, as for WMATA. The MCMA performance figures in Table 3.3 were obtained on the RAVEM cycle, reflecting actual road performance along a BRT route (Weaver & Almanza, 2006); as noted, the two CNG buses (Ankai and Busscar) on average had a 34% lower fuel economy than their two diesel counterparts (Volvo 12 and Mercedes Benz 12) on this cycle. The diesel and CNG buses were also tested on a chassis dynamometer (WVU, 2005), on the Mexico City Schedule (MCS), representing "average" driving conditions there (WVU, 2004), and the European Transient Cycle (ETC); overall, the ETC cycle has a much higher average speed, with significantly lower stop and go, than MCS. For both diesel and CNG, fuel economy was significantly higher (double) on ETC than on MCS; further, the fuel economy penalty for CNG relative to diesel reduced to around 19% on both cycles, from the 34% on-road. Barnitt (2008) also reported higher fuel economy on a chassis dynamometer relative to road tests because of higher speeds, and the lack of airconditioning load, in the former case. All of the foregoing demonstrates the critical importance of making comparative evaluations under operating conditions that are similar and as close as possible to real-life conditions, in terms of vehicle vintage, weight, speed, and engine power among other factors.

Also importantly, while the on-road fuel economy of buses with CNG engines operating on the SI cycle can be inferior to diesel buses, on an energy basis, especially at low operating speeds, Table 3.4 shows that buses with advanced CNG engines can be better in this regard than diesel buses, especially at higher operating speeds and as diesels increasingly require advanced exhaust control to meet stringent emission standards, in particular relating to PM. On the dynamometer tests (Table 3.4), fuel economy on the dynamometer was much higher than on-road for both fuels. Besides, the 2001 CNG buses had only 1.5% worse fuel economy than their diesel counterparts, and the 2004 CNG buses were actually 9% better than their diesel counterparts (Table 3.4). These results are partly due to the improvement in CNG engine technology, but also the significantly tightened EPA emission standards for 2004 (especially in terms of HC and NOx) having called for more advanced exhaust treatment (in terms of EGR and DPF systems) on diesels, with significant negative implications for fuel economy. Similar results were also observed in other chassis dynamometer tests for newer CNG and diesel buses in Wayne et al. (2008). But it is important to keep in mind that, notwithstanding these dynamometer test results, on-road fuel economy might still be better for diesel than for CNG, given the significantly higher fuel economy gap between dynamometer and road tests for CNG relative to diesel.

At the same time, as a comparison of the NYCT hybrid and the WMATA buses in Table 3.3 shows, buses with diesel hybrids have a significantly higher economy than advanced CNG, and indeed, conventional diesels, despite similar exhaust control requirements as the latter, and even at a lower operating speed, and higher vehicle weight -- with this advantage being gained, of course, without any of the costs and other implementation issues related to a change in fuel infrastructure.

# ii) Vehicle emissions

Table 3.4 compares the emissions (and fuel economy) performance, based on chassis dynamometer testing, of 12 40-foot, low-floor WMATA buses representing diesel and lean-burn NG (natural gas) technologies, of 2000/01 and 2004 vintage, using West Virginia University (WVU)'s Transportable Heavy-Duty Vehicle Emission Testing Laboratory (Melendez et al., 2005); the project was led by USDoE's National Renewable Energy Laboratory (NREL).

The test bus configuration details are in Table 3.4. Both the 2000 diesel buses, and all but one of the 2001 and 2004 CNG buses in Table 3.4 were the same as the on-road

WMATA test buses in Tables 3.2 and 3.3. New EPA emission standards came into force for transit buses in 2004, for NOx and HC. Both the 2000 and 2004 diesel buses used a DDC S50 platform, which was the mainstay of the transit bus industry, but was discontinued in 2004, because it could not meet the prospective 2007 emission standards (Barnitt, 2008). Whereas only a catalyzed particulate filter was required to meet the applicable standard in 2000, the 2004 diesel required EGR additionally, to meet the 2004 NOx + NMHC standard. Meanwhile, CNG engines needed an oxidation catalyst to meet both the 2000 and 2004 standards; indeed, the CNG buses in 2001 met the 2004 NOx + NMHC standard (Melendez et al., 2005). The buses were tested on the WMATA cycle, which simulates the on-road operation of WMATA buses. The test buses had GVWs ranging from 18.1 to 19.3 MT (metric tonnes), similar to the US buses in Table 3.2; also, the test weights (curb weight plus half the maximum passenger load) fell within a narrow range, but the NG buses were tested at a higher weight than diesel. The diesel fuel used was ULSD (15 ppm sulphur).

NOx was lower for the 2004 relative to the 2000 diesel bus (because of EGR in response to 2004 emission standard), but CO, PM and CO<sub>2</sub> (and fuel economy) deteriorated, reflecting emission and performance trade-offs due to control technologies. The 2004 (John Deere) CNG was much better than the 2001 (DDC S50) CNG bus on both emissions and fuel economy, with CO, HC, CH<sub>4</sub>, and NOx in particular having improved significantly, without a change in emission controls. Therefore, the overall performance gap between the two fuels widened in favour of NG in 2004, relative to 2000/01. Overall, both PM and NOx, the principal focus of the US-EPA standards, were lower for CNG; in the case of NOx, even more so in 2004 than in 2001, despite EGR on the diesel engine. Diesel fuel economy was slightly (1.5%) better relative to NG in 2001, but considerably (9%) worse in 2004, perhaps because of the additional EGR. CO<sub>2</sub> was considerably lower for NG, by as much as 35% in 2004. On the other hand, HC, particularly CH<sub>4</sub> (important from a GHG perspective), was vastly higher for CNG (both CH<sub>4</sub> and NMHC were zero for diesel).

The dynamometer tests on the WMATA buses show the superiority of NG relative to diesel, with regard to PM, NOx, fuel economy and CO<sub>2</sub>, but note that the 2004 John Deere NG engine was more advanced than its DDC S50 counterpart, which was phased out in that year; besides, the diesel EGR likely affected fuel economy. It is possible that the superior fuel economy for NG might not have been observed on-road; whereas the fuel economy of the 2000 diesel was only 1.5% better than for its NG counterpart on the dynamometer, it was 18% better on the on-road test in Table 3.3, demonstrating the importance of test conditions. Lastly, if NG engine technology has been improving, so has diesel (and diesel hybrid) technology. It is therefore important to compare advanced diesel versus NG technology, as done in Wayne et al. (2008), which also involves dynamometer testing of (2005/06 vintage) WMATA buses.

Table 3.5 compares the emissions and fuel economy performance, based on chassis dynamometer testing, of 12 40-foot WMATA buses representing diesel, CNG, and hybrid-electric diesel technologies, of 2004-06 vintage, using West Virginia University (WVU)'s Transportable Heavy-Duty Vehicle Emission Testing Laboratory (Translab); the project was sponsored by USDoE and USDoT (US Department of Transportation), and conducted in co-operation with WMATA (Wayne et al., 2008). The test buses, details of which are in Table 3.5, comprised two with 2006-year Cummins diesels, six buses with an equal number of Cummins and John Deere lean-burn NG engines of 2004/05 vintage, two buses with 2005/06-year Cummins hybrid-electric diesels, and one diesel bus of 1992 vintage, retrofitted with a DDC S50 engine of 2003 vintage. None of the above buses were common with the 2000-04 WMATA buses used in the dynamometer tests in Melendez et al. (2005) (Table 3.4), or the on-road tests in Chandler et al. (2006) (Tables 3.2 and 3.3). However, as in the case of the dynamometer tests on 2000-04 vintage WMATA buses reported by Melendez et al. (2005), the buses in these tests had GVWs (17.9-19.3 MT) similar to the US buses in Table 3.2; also, the test weights (curb weight plus half the maximum passenger load) fell within a narrow range, and the NG and hybrid buses were tested at a higher weight than diesel.

One bus each from the diesel, (lean burn) CNG, hybrid-electric diesel, and retrofitted diesel groups was tested on 17 different driving cycles, and the other buses in these groups on six of these cycles. In Table 3.5, test results are excerpted from Wayne et al. (2008) for the different technologies, on the NY Bus (the most severe in terms of average speed and frequency of stops), the Manhattan Bus, WMATA, and the ETC-Urban cycles; the average speeds and stops per kilometre, drawn from Nylund et al. (2007), are indicated for each cycle. All the diesels were tested with ULSD (15 ppm sulphur). The Cummins and John Deere lean-burn NG engines were catalytically controlled; the hybrid-electric diesels used a particulate filter; and the DDC S50 on the retrofitted bus for which results are reported in Table 3.5 used a catalyzed particulate filter.

	Diesel	CNG		Diesel-Hybrid	Diesel (retrofit)
	Cummins	Cummins	CNG	Cummins	DDC S50
	2006	2005	JD 2005	2005/06	2003
No. of buses tested	2	3	3	3	2
CO Emissions (Avg., g/km)					
NY Bus <sup>(1)</sup>	5.95	1.44	0.23	0.21	N/A
Man Bus <sup>(1)</sup>	2.74	1.03	0.21	0.03	0.07
WMATA Cycle <sup>(1)</sup>	2.84	0.35	0.11	0.30	0.04
ETC Urban Cycle <sup>(1)</sup>	1.58	0.07	0.02	0.01	N/A
NOx Emissions (Avg., g/km)					
NY Bus	11.71	39.01	23.08	10.37	28.03
Man Bus	7.52	22.00	14.32	6.95	15.04
WMATA Cycle	5.65	13.80	8.54	6.12	13.05
ETC Urban Cycle	4.32	8.64	7.83	4.20	6.90
HC Emissions (Avg., g/km)					
NY Bus	1.45	47.09	33.89	0.13	N/A
Man Bus	0.65	33.75	15.97	0.09	0.19
WMATA Cycle	0.61	18.44	11.18	0.02	0.14
ETC Urban Cycle	0.52	14.73	11.06	0.02	0.07
PM Emissions (Avg., g/km)					
NY Bus	0.25	0.03	0.06	0.03	0.04
Man Bus	0.15	0.01	0.02	0.01	0.01
WMATA Cycle	0.10	0.01	0.01	0.00	0.01
ETC Urban Cycle	0.06	0.01	0.01	0.01	0.02
Fuel Economy (Avg., km/L equiv.) <sup>(2)</sup>					
NY Bus	0.67	0.56	0.57	1.07	0.56
Man Bus	1.19	0.89	1.00	1.66	0.91
WMATA Cycle	1.39	1.14	1.28	1.85	1.09
ETC Urban Cycle	2.36	1.78	1.92	2.66	2.13

Table 3.5: WMATA chassis dynamometer-based emissions, 2003-2006 vintage buses, various test cycles.

<sup>(1)</sup> Average speeds for the driving cycles are: NY Bus, 5.9 km/h; Man Bus, 10.9 km/h; WMATA Cycle, 13.4 km/h; ETC Urban, 22.7 km/h

<sup>(2)</sup> Fuel economy Data for Diesel Cummins 2006 and Hybrid Cummins 2005/06 obtained through direct correspondence with authors (Wayne et al. 2008)

Source: Wayne et al. (2008)

In comparing the NG with the diesel buses, the first point to note is the significant differences in the emissions profile between the two NG engines; while the Cummins engine had lower PM, it had considerably higher NOx, CO, CH<sub>4</sub> and HC levels than the John Deere NG engine, especially on the lower average speed cycles, demonstrating the emissions trade-offs as a result of technological strategies to meet emission standards, on engines of similar configuration, exhaust control, and vintage. The diesel engine had very low HC levels, as expected, and NOx nearly as low as the hybrid, but by far the highest PM and CO emissions of all the technologies. Even on the John Deere engine, which was the better performing NG engine overall, HC (around 90% of which was CH4), NOx and aldehydes were significantly higher relative to the diesel engine (18-25 times in the case of HC, and 1.5-2 times for NOx). On the other hand, the NG engines had very low PM and CO emissions -- 10-25% and 4-25% respectively of the levels on the Cummins 2006 diesel, on the worse performing NG engine in each case. While the emissions results on the NG engines were mixed relative to the diesel, the emission levels on the Cummins hybrid engine were either competitive with or lower -- particularly in terms of HC -- than on the best of the other technologies. Most interestingly, the 2003-year DDC S50 diesel (discontinued in 2004) retrofitted on a 1992 chassis, had -- besides the lowest HC emissions apart from the hybrid -- similar PM and lower CO emissions relative to NG, and indeed, significantly lower PM than the 2006 Cummins diesel -- thereby showing the benefit of DPFs -- and NOx emissions between the levels for the NG buses.

The fuel economy figures in Table 3.5, excerpted from Wayne et al. (2008) for the NY Bus, Manhattan, WMATA and the ETC-urban cycles, were inferred from a carbon balance, and do not account for air-conditioning loads nor road grade in actual operation. The Cummins NG buses had a 16-25% worse fuel economy relative to the Cummins 2006 diesel across these cycles; the John Deere NG engine was also inferior to diesel across these cycles, but about 8-12% better than the Cummins NG engine at higher speeds. The hybrid-electric diesel was on average 25% and 45% better than the diesel and NG engines on all the cycles tested by Wayne et al. (2008). Finally, the retrofitted 2003 DDC S50 diesel was 16-24% worse on the slow speed cycles in Table
3.5, but only 4% worse on the ETC-Urban cycle, relative to diesel; and 2-15% worse at slow speeds, but actually 11% better on the ETC-Urban cycle, relative to John Deere NG engine, which was the better performing of the two NG engines.

Operating conditions, and therefore driving cycles, profoundly affect emissions and fuel economy for all technologies. Emission levels were significantly higher (particularly for CO), and fuel economy was significantly lower, on cycles with low average speeds, high frequency of stops, and high percentage of idling. In the case of PM, however, while emissions were the highest on the NY Bus cycle, the most severe cycle in terms of these parameters, PM showed a mixed trend with respect to speed. In the case of fuel economy, the effect of driving cycle average speed was the highest for the NG and diesel engines; for these two technologies, the fuel economy on the ETC-Urban cycle was over 200% higher than on the NY Bus cycle. This effect was the lowest for the hybrid engine -- the corresponding improvement on the ETC-Urban cycle was 149%. At the same time, even a small increase in average speed produced a considerable FE improvement for all technologies -- fuel economy on the Manhattan cycle (with less than twice the speed, and half the stops per kilometre relative to the NY Bus cycle) was improved by 55% for the hybrid, and by 60-80% for the diesels and NG engines.

Operating conditions by way of test cycles also affect the relative emissions and fuel economy performance of various technologies (Wayne et al., 2008). The performance of the retrofitted DDC S50 diesel relative to the 2006 Cummins diesel and the NG engines at slow versus high speeds, and the differential emissions performance of the two NG engines at low speeds, has already been referred to. In the case of the hybrid, its fuel economy advantage is higher on cycles with lower speeds with frequent stops and a high percentage of idling, since regenerative braking and energy storage are most beneficial under these conditions. Relative to the John Deere engine, the better performing NG engine, the hybrid was 39 and 45% more fuel economical on the ETC-urban and WMATA cycles, but 88% more so on the NY Bus cycle -- that the effect of speed on fuel economy is lowest for the hybrid, as noted earlier, is also related to this differential FE advantage relative to other technologies at various speeds.

Lastly, as noted, the Mexico City study sought to measure emissions from public transit buses with current and advanced engine control, aftertreatment technologies, and low emission fuels, for the purpose of introducing climate friendly and low emission transport measures. The study involved the measurement of emissions in-use, as well as on a chassis dynamometer.

The in-use RAVEM (Ride-along vehicle emission measurement) tests (Weaver & Almanza, 2006) were conducted on 17 buses of 2002-04 vintage -- a sub-set of the 22 buses selected for evaluation in SMA (2006a), as described above -- and of varying dimensions and GVWs, ranging from 14 to 29.5 MT. The buses comprised 10 diesels, including the MB 12 and Volvo 12 in Table 3.2, conforming to EPA 1998 or EURO III emission standards, except one, which conformed to EPA 2004; two diesels conforming to EPA 1998 but with emission controls (catalytic DPFs); three lean burn CNG buses conforming to EPA 2004, and with 2-way catalysts, including the Busscar and Ankai in Table 3.2; and two diesel-hybrids, one equipped with a DPF. The tests were conducted during 2004-05, in normal traffic, and on a traffic-free BRT corridor, at average speeds of 17.1 and 21.4 km/h respectively.

Eight of the RAVEM vehicles were also tested on the West Virginia University (WVU) chassis dynamometer system (WVU, 2005) -- six diesels (including the Volvo 12 in Table 3.2, and another equipped with a DPF); two CNG buses with the same emission controls as above (the Busscar in Table 3.2, and an FAW, of high GVW); and one DPF-equipped hybrid. The diesels were tested with fuels of varying sulphur content (350, 50 and 15 ppm on RAVEM, and 350, 150, and 15 ppm on the dynamometer). The dynamometer tests were conducted on the MCS driving cycle, as well as on its constituent MX1 (low speed), MX2 (medium speed), and MX3 (high-speed, BRT bus only corridor) cycles, and on the European Transient Cycle (ETC). Comparisons of emissions (and fuel economy) across different buses were made on a distance, as well as a passenger-mile and ton-mile basis.

On the RAVEM tests, PM levels were the lowest (0.01-0.04 g/km) for the CNG buses; while they were 0.07-0.45 g/km for the diesels without DPFs, they were only 0.01-0.08, not very different from NG, except for a higher maximum, for the two diesels with DPF, and 0.03 g/km -- competitive with NG -- for the DPF-equipped diesel-hybrids. While NOx levels were the lowest for the CNG buses, they varied widely, and were 30-80% higher than the applicable EPA 1998/2004 standards for the diesels, which yet again shows the importance of in-use testing under real life conditions. CO was very low (0-6 g/km) from the diesel and NG buses. Total VOCs were virtually zero from the diesels; meanwhile, the three NG buses emitted 5-52 g/km CH<sub>4</sub>, but only 0.1-3 g/km ethane, and much lower levels of higher carbon VOCs. Lastly, two 1991-vintage buses from RTP's transit fleet were tested, one with original model engine and the other with a repowered engine meeting the EPA 1998 standard; on the retrofitted bus, PM, CO, and NOx were 88, 86, and 59% lower than for the non-modified 1991-vintage bus; retrofitting old buses with engines with advanced emission controls can achieve significant emission reductions cost-effectively.

On the WVU (2005) dynamometer tests, as in the case of RAVEM, PM on the MCS cycle was by far the lowest (<0.013 g/km) for the CNG buses, followed by the DPF-equipped hybrid (0.025 g/km). Among the diesels, the DPF-equipped bus had the lowest PM level (0.09 g/km); meanwhile, the other diesels (without DPFs) had PM levels ranging from around 0.3 to as high as 1.25 g/km (on the Volvo 12 bus). An interesting point (WVU, 2005) is that sulphur content made little difference to PM. A small fraction of sulphur is converted to sulphates and sulphuric acid, which accounts for a small proportion of PM mass; thus, sulphates may be significant for high sulphur fuels, but not at the levels tested<sup>5</sup>. While NOx levels were the lowest for the hybrid bus, they were the highest for one of the CNG buses, contrary to the RAVEM case; as for the diesels, NOx levels varied widely, as for RAVEM. CO levels were low for the CNG and the DPF hybrid buses; for the diesels, CO levels ranged widely, from even lower than for the

<sup>&</sup>lt;sup>5</sup> WVU (2005), note in this respect that the benefit of low-sulphur diesel is to "allow the use of catalyzed PM filters on the exhaust. This is important, because these catalyzed filters will generally not function well without the low sulphur diesel" (p.33).

CNG buses on the DPF-equipped bus to as much as 20 times higher on the heavier buses and the Volvo 12, showing yet again the importance of technology and maintenance. While HC emissions were expectedly low even for the heaviest diesel buses, they were close to zero for the DPF-equipped diesel and hybrid bus, demonstrating the important role of DPFs in reducing not only diesel PM but also CO and HC. As in the RAVEM case, HC, the bulk of which was CH<sub>4</sub>, was very high (6.3-9.4 g/km) for the CNG buses; but even their NMHC emissions, around 0.63 g/km, were higher than for the diesels.

On the WVU (2005) dynamometer tests, the fuel economy of the diesels declined with increasing weight. The Busscar NG, despite a similar GVW to Volvo 12, the worst performing diesel in its weight class, had a 19% worse FE, and indeed, diesel buses with significantly higher (by 50-70%) GVW had FE similar to the Busscar. Also, the FAW CNG bus, with a much higher GVW, but lighter than the heaviest diesels tested, had a FE 24% lower, and the lowest FE of all vehicles. Meanwhile, the hybrid bus, despite a 9% higher GVW than the Volvo 12, had a 19% better FE, and a 30% better FE than the Busscar NG vehicle, of similar GVW. FE over the ETC was significantly higher than over MCS, which involved the same average speed but 40% fewer stops than MCS; on the ETC, the improvement over the MCS cycle was greatest for the Busscar NG bus, for which FE doubled, lower for the heavy diesels and the hybrid (around 70%), and lowest for the lightest (and most fuel economical) diesels (43%). On the low speed MX1, the NG bus had the lowest FE, as low or lower than the heaviest diesels; on the higher speed MX2, FE was generally higher than on MX1, but the FE advantage for the hybrid was lost somewhat; and FE was the lowest overall on the MX3, the most severe cycle with many stops, despite its high average speed.

All of this shows again the FE deficit for NG relative to diesel; the important role of operating conditions, and that lean-burn NG performs particularly badly under low speed, and transient conditions; and that the advantage for hybrids is the highest at low speeds. Finally, on a passenger-kilometre and tonne-kilometre basis, there is far lower

FE difference between buses than when distance based measures are used; and the hybrid and the heaviest diesel buses had the highest FE on a tonne-kilometre basis.

Importantly, there were large differences in emissions (besides fuel economy) between individual diesel as well as CNG vehicles with similar engine and control technologies, showing the importance of technology and maintenance. Secondly, just as for fuel economy, operating conditions affect emissions significantly. For example, PM was higher on the MX3 cycle relative to MCS; and the FAW CNG bus, which had the highest NOx level on the MCS cycle, performed even worse in this regard on the slow speed MX1, demonstrating the poor low speed NG performance, already noted. Further, emissions on the MX3 cycle were generally 30% higher than on RAVEM; note that while the average speeds are the same, the MX3 cycle involves 66% more stops, and is therefore far more severe, than RAVEM. Lastly, emissions were significantly lower on the ETC, which has much fewer transients, than on the MCS cycle; in the case of PM for example, by as much as 80% for the diesels; and by 50% or more for NOx, for all bus technologies. Similarly, the significant reduction in CO2 emissions (discussed below) on the ETC was in the same proportion as the improvement in fuel economy.

CO<sub>2</sub> emissions reflect fuel economy, as well as the carbon-hydrogen ratio of the fuel. At the same time, because fuel economy is strongly determined by vehicle weight, which varied widely, from 14 to 29.5 MT, as noted, it is desirable to compare CO2 emissions (and fuel economy) on a passenger-kilometre, or even better, on a ton-kilometre basis. On a per-kilometre basis, the three lightest vehicles, all diesels, including the one equipped with the DPF, had the lowest CO<sub>2</sub> levels, followed by the heavier Busscar CNG and the hybrid buses; the three heaviest vehicles, including the FAW CNG vehicle, had the highest CO<sub>2</sub> emissions. On a tonne-kilometre basis, however, the heaviest vehicle, a 29.5 MT diesel bus, had the lowest CO<sub>2</sub>, followed by the FAW CNG, despite its fuel economy being the lowest on a per-kilometre basis. All other things being equal, distance-based emission (and fuel economy) measures favour lighter vehicles, while passenger-kilometre and tonne-kilometre measures favour larger and

heavier vehicles, since fuel consumption and emissions do not increase in the same proportion as vehicle weight (WVU, 2005).

## iii) Reliability and operating costs

This section discusses the reliability and maintenance costs, focusing particularly on engine and fuel system (EFS) related maintenance costs, which are the most important from the perspective of this chapter, for analysis of the chosen cases. At NYCT, reliability in terms of MBRC (miles between roadcalls) -- for which Table 3.3 shows data converted to kilometres between roadcalls -- was the same for the 2002 CNG and 1999 diesel fleets; however, EFS-related reliability was considerably lower for the CNG vehicle, despite the diesel fleet being a few years older at the time of the evaluation. Further, as many as 61% of all road calls were EFS-related for the 2002 CNG fleet, as against 47% for the 1999 diesel fleet. In particular, the non-lighting electrical system (spark plugs, battery, and alternator) was the most responsible for EFS road calls for the CNG 2002 fleet (44%), followed by the engine (31%), and the fuel system (21%). The corresponding shares were 28%, 19% and 19% for the 1999 diesel fleet. In the case of the 2002 diesel-hybrid fleet, both vehicular and EFS reliability were poorer than for the CNG and diesel fleets of similar vintage. The electric propulsion system was the primary cause of failures, accounting for as many as 62% and 39% of EFS-related and total vehicular road calls respectively. Overall, a higher proportion of vehicular road calls were EFS-related, for the CNG and hybrid fleets, relative to the similar vintage 1999 diesel fleet. Incidentally, the 1994 diesel fleet had a significantly lower reliability relative to the 1999 diesel fleet, which had the same engine platform, and was evaluated over the same period, demonstrating the significant deterioration in this regard, for both the vehicle and for EFS, with vehicle age.

At WMATA, both vehicle and EFS reliability was much better for the CNG fleet relative to its slightly older diesel counterpart; the 2004 CNG fleet, powered by a John Deere engine subject to a more stringent emission standard, had a slightly better vehicle as well as EFS reliability than even the 2001 CNG (note however that the fleets had different engine platforms; also, whereas the 2001 CNG fleet was evaluated June 2003May 2004, the 2004 CNG fleet was considerably younger when it was evaluated in May-October 2004). The engine was a major cause of EFS roadcalls for both the 2001 and 2004 CNG fleets (50% and 75% respectively); for the latter fleet, the non-lighting electrical system was also an important cause.

Comparing the performance of CNG and diesel fleets of similar vintage under different operating conditions at different operators shows that, whereas vehicular reliability was significantly worse for both fuel systems at WMATA relative to NYCT, EFS reliability was significantly better for WMATA relative to NYCT, especially for CNG. This is also reflected in the fact that EFS accounted for a significantly lower share of total road calls (26-29%) for the diesel and CNG fleets at WMATA, as against 47% and 60% at NYCT. This suggests that higher operating speeds perhaps help EFS, but not necessarily vehicle reliability. However, note that the vehicular reliability of the SunLine CNG fleet was more than double that for its considerably older NYCT and WMATA counterparts (unfortunately, there is no information on the EFS reliability for the SunLine fleet).

Fuel costs (in \$/km) are of course a function of fuel economy as well as fuel price (Table 3.3 shows fuel price in \$/diesel litre equivalent for all fuel systems). In the case of both NYCT and WMATA, the fuel cost for the CNG fleet is higher in the same proportion as its fuel economy is lower, relative to its diesel counterpart, given very similar prices during the relevant evaluation periods for each fuel for each operator. However, note that the fuel cost is significantly lower for both fuel systems for WMATA relative to NYCT, because of the much higher fuel economy (likely due to more favourable operating conditions) plus the significantly lower for CNG, despite a significantly poorer fuel economy relative to diesel, because of a much lower CNG price relative to diesel in Mexico City.

Next, maintenance costs are presented, again focusing particularly on EFS-related maintenance costs, in \$/km, for the selected cases; note that a \$50/hr labour rate was assumed across the US studies, in order to enable comparison. The WMATA data in

Table 3.3 (Chandler, Eberts & Melendez, 2006) allows the comparison of the maintenance cost performance of CNG and diesel fleets of similar vintage; this is not possible for NYCT, for which maintenance costs are not available for the diesel fleet, because of its age (Barnitt & Chandler, 2006; Barnitt, 2008).

Even though the EFS-related MBRC figures for the 2001 and 2004 CNG fleets at WMATA were much higher than for their diesel counterpart, the EFS-related maintenance cost in \$/km was interestingly the same for all these cases, implying more costly though less frequent repairs for the CNG fleets. The total maintenance cost in \$/km was slightly lower for CNG compared to diesel. The EFS share of total maintenance cost was similar (22-25%) for the diesel and CNG fleets at WMATA; interestingly, the EFS shares of total road calls and maintenance costs were also similar. For NYCT's 2002 CNG fleet, on the other hand, EFS cost was 24% of total maintenance cost, even though EFS accounted for 60% of total road calls. Even so, EFS, along with transmission, was the highest maintenance cost category, followed by the cab, body and accessories. Importantly, preventive maintenance inspection (PMI) accounted for a major share (23-29%) of total maintenance cost at WMATA.

At both NYCT and WMATA, the non-lighting electrical system and the engine were the most important EFS-related maintenance cost categories for the CNG fleets. At NYCT, the non-lighting electrical system accounted for 29% of EFS maintenance cost, followed by the engine (18%), the fuel system (17%), and cooling (15%); these were also the most important causes of EFS road calls, in the same order. At WMATA, the top two categories accounted for 29% and 23% of the EFS maintenance cost for the 2001 CNG fleet, whereas for the 2004 CNG fleet, the engine was the most important cause of road calls, as noted. For the 2000 diesel fleet, the predominant EFS cost category was the cooling system, followed by the engine. Labour was the most important cause of costs, and the above components were generally ranked similarly in labour hours as for EFS cost, for each fleet.

Finally, for the NYCT hybrid fleet, EFS maintenance cost accounted for 29% of total maintenance cost, even though 63% of total road calls. Within EFS, the electric propulsion system was by far the most important cost category (49%), followed by engine (17%), and the non-lighting electric system (12%). EFS maintenance cost was higher for the hybrid than for the CNG fleet because, though the parts costs were similar, labour costs were considerably higher for the hybrid fleet, mainly related to the electric propulsion, followed by the engine. Overall, however, the total maintenance cost for hybrid was lower than for CNG, despite the higher EFS cost, and the lower vehicular MBRC, relative to CNG.

Some interesting trends emerge when comparing the reliability and operating cost performance across the US operators. In the case of EFS maintenance cost, WMATA's reliability (in MBRC) was much better than NYCT's, for both CNG and diesel, and correspondingly, WMATA's EFS costs in \$/km were also much lower. However, WMATA's vehicular maintenance costs in \$/km were also much lower than NYCT's, despite WMATA's much more frequent road calls relative to NYCT's. At the same time, even though WMATA's EFS and vehicle maintenance cost was much lower than NYCT's, WMATA's EFS share of vehicle maintenance cost was the same as for NYCT (22-25%).

In carefully examining the detailed cost breakdowns by vehicle systems for NYCT and WMATA (Barnitt & Chandler, 2006; Chandler, Eberts & Melendez, 2006) it turns out that the labour hours as well as parts expenditure per kilometre are significantly lower for the latter, both for EFS as well as the vehicle as a whole. This factor, which implies effective cost management, as well as the considerably larger role for preventive maintenance inspection at WMATA, as already discussed, might explain the significantly lower maintenance costs for WMATA, even for the vehicle as a whole, despite much more frequent road calls, relative to NYCT. Finally, SunLine's maintenance costs, which are considerably lower than even WMATA's, despite its being a much older CNG fleet, demonstrate the considerable scope for cost control, through effective management, preventive maintenance, and training.

At WMATA, CNG had a higher total operating cost in \$/km, even though lower maintenance cost, relative to diesel; this was due to the higher fuel cost, because of the lower fuel economy, for CNG relative to diesel. At MCMA, on the other hand, while CNG had a similar maintenance cost, the total operating cost per kilometre was significantly lower, despite a 23% poorer fuel economy than diesel, mainly because of a significantly lower unit price of CNG relative to diesel.

In the case of SunLine, CNG operating cost was much lower even than for WMATA, not only because of low maintenance costs, as discussed, but also because of significantly lower fuel cost, both due to a high fuel economy and a very low unit fuel price for CNG. In the US, a survey of bus transit operators showed that fleet managers using CNG considered having a reliable and in-house refuelling system one of the top reasons for a successful CNG experience, along personnel training, commitment to the new fuel system, and a deep understanding of the costs involved (Eudy, 2002). Finally, in the case of the hybrid fleet at NYCT, the total operating cost was considerably (23%) lower than for CNG, due to a lower maintenance cost (despite slightly higher EFS cost), but mainly because of a considerably lower fuel cost, on account of a significantly higher fuel economy for hybrid relative to CNG.

Overall, it appears that variation in the total operating cost across fuel systems for a given operator is due primarily to variation in fuel cost, since maintenance cost varies considerably less. The variation in operating cost across operators is due to variations in fuel economy (due to operating conditions) and fuel unit price, and variations in maintenance costs due to cost management, preventive maintenance, and training.

## iv) Fuel cost sensitivity analysis at NYCT and WMATA

Since the variation in total operating costs across fuel systems is due primarily to variation in fuel costs, this section simulated the impact of changes to CNG and diesel prices (a primary determinant of fuel costs) on total operating costs at NYCT and WMATA. For consistency in comparisons between operators, only the US context was

included in this analysis; NYCT was chosen since it allows for a comparison of costs between CNG and diesel-hybrid buses; WMATA was chosen since it allows for a comparison between CNG and diesel buses; finally, SunLine was not included since it does not operate diesel buses. The sensitivity analysis was based on the original cost estimates for the two selected operators, as provided in Table 3.3.

An overview of CNG and diesel prices in the US, calculated in inflation-adjusted US dollars is presented in Figure 3.1. It can be seen that, the price of each fuel fluctuates substantially, not only over time, but also relative to each other. The average retail fuel price from 2010-2015 was \$0.97-per-litre for diesel and \$0.63-per-diesel-litre-equivalent for CNG (EERE, 2016). If the prices of these two fuels are compared in terms of their difference (ratio), diesel was sold at an average premium of 1.55 times the energy equivalent unit price of CNG in the US over the past five years, from 2010-2015 (Figure 3.1).



Source: Fuel Prices (EERE, 2016), Inflation-adjustment (BLS, 2016) Figure 3.1: CNG versus diesel retail fuel prices in the US

To analyze the impact of CNG and diesel price changes on total operating costs at NYCT and WMATA, bus fleet fuel and operating costs from Table 3.3 were updated using various hypothetical changes to fuel prices. The results of this analysis are presented in Table 3.6. Case 1 reproduces the original fuel and maintenance costs reported for NYCT and WMATA in Table 3.3, but with inflation-adjusted 2015 values using the Consumer Price Index (CPI) (BLS, 2016). Given the significant differences in diesel and CNG prices in Case 1 (Table 3.6), in Case 2, I considered that both operators paid the same fuel price. For Case 2, diesel and CNG fuel prices were taken from the 2010-2015 five-year period (EERE, 2016), as shown in Figure 3.1, and used as the basis of further comparisons<sup>6</sup>. For further comparisons, alternate scenarios were created in Cases 3 and 4, which considered more extreme differences in the price ratio of diesel-to-CNG, based on the historical variations in CNG and diesel prices.

Cases 3 and 4 in Table 3.6 were constructed on likely range of estimates of relative fuel prices<sup>7</sup>, from lower to upper bound values for diesel and CNG prices, in turn, which were based on observed variations in diesel and CNG prices in the US from 2000 to 2015 (Figure 3.1); this was measured by taking the diesel-to-CNG fuel price ratio. Figure 3.1 shows that, over a 15-year time period, the price premium of diesel relative to CNG fluctuates from a minimum of 0.8 to a maximum of 1.8, with an average of 1.4. Based on these variations in the fuel price ratio over the 15-year period, a statistical confidence interval was constructed, that took average diesel-to-CNG price ratio of 1.55 (given diesel at 0.975 \$/litre; and CNG at 0.629 \$/diesel equiv. litre) as the center-point for the sensitivity analysis. The variance observed in the 15-year period (2000-2015) for the diesel-to-CNG ratio was used in constructing confidence interval values, and resulted in a 1.4x lower bound ratio and a 1.7x upper bound ratio<sup>8</sup>. The lower bound

<sup>&</sup>lt;sup>6</sup> This five-year period was chosen, given the substantial variations observed in the ratio of diesel-to-CNG over the years, as shown in Figure 3.1. Further, note that these updated prices are based on retail fuel pump prices, and it could be expected that fleet operators may pay a discounted price to this, given the long-term commitment and bulk quantities of fuel purchased on a regular basis.

<sup>&</sup>lt;sup>7</sup> For this analysis, relative fuel price (i.e., fuel price ratio) is defined as the ratio of the price of diesel (\$/litre) to the price of CNG (\$/diesel equiv. litre), on an energy equivalence basis; historical data for this ratio is available in Figure 3.1.

<sup>&</sup>lt;sup>8</sup> See Footnotes (4) to (6) of Table 3.6 on the methodological details on the construction of these confidence intervals.

(Case 4) represents a price scenario that is less advantageous for CNG (that is, when the price difference of diesel relative to CNG smaller), whereas the upper bound (Case 3) represents the price scenario that is more advantageous to CNG operations (that is, when diesel is relatively more expensive than CNG).

	NYCT									
	Hybrid	% Diff to Case 1 <sup>(1)</sup>	CNG	% Diff to Case 1 <sup>(1)</sup>	Diesel	% Diff to Case 1 <sup>(1)</sup>	CNG	% Diff to Case 1 <sup>(1)</sup>	CNG	% Diff to Case 1 <sup>(1)</sup>
Model Year (MY)	2002		2002		2000		2001		2004	
Fuel economy (km/L) <sup>(2)</sup>	1.36		0.72		1.21		0.99		1.02	
				00/F ·		0.01			4.0	
CASE 1: Original data from Table 3.	3, but infl	ation-adju	sted to	2015 prices	using ave	erage CPI	for 2015	(BLS, 20	16)	
Fuel price (\$/diesel litre							- · -			
equivalent) <sup>(2)</sup>	0.58		0.56		0.47		0.45		0.44	
lotal operating costs (\$/km)	<u>1.36</u>		<u>1.77</u>		<u>0.87</u>		<u>0.87</u>		<u>0.88</u>	
Fuel costs (\$/km)	0.42		0.78		0.39		0.45		0.44	
Maintenance costs (\$/km)	0.94		0.99		0.49		0.41		0.45	
CASE 2: Fuel prices averaged over 2010-2015 period <sup>(3)</sup> [assumed same in NYCT and WMATA]										
Fuel price (\$/diesel litre		• • • • •	L				3			
equivalent) <sup>(2)</sup>	0.97		0.63		0.97		0.63		0.63	
Total operating costs (\$/km)	1.66	22%	1.86	5%	1.29	48%	1.05	21%	1.07	21%
Fuel costs (\$/km)	0.72	69%	0.87	$\frac{11\%}{11\%}$	0.81	109%	0.64	41%	0.62	42%
Maintenance costs (\$/km)	0.94		0.99		0.49		0.41		0.45	/ •
CASE 3: 95% Confidence Interval <sup>(4)</sup> for diesel-to-CNG price ratio (Upper Bound) <sup>(5)</sup>										
Fuel price (\$/diesel litre										
equivalent) <sup>(2)</sup>	1.04		0.60(5)		1.04		0.60(5)		0.60(5)	
Total operating costs (\$/km)	<u>1.71</u>	<u>25%</u>	<u>1.82</u>	<u>3%</u>	<u>1.35</u>	<u>55%</u>	<u>1.03</u>	<u>19%</u>	<u>1.04</u>	<u>18%</u>
Fuel costs (\$/km)	0.77	81%	0.84	7%	0.86	123%	0.61	35%	0.59	37%
Maintenance costs (\$/km)	0.94		0.99		0.49		0.41		0.45	
CASE 4: 05% Confidence Interval <sup>(4)</sup> for discel to CNC price ratio (Lower Bound <sup>1</sup> (6)										
Evol price (\$/discel litre					Junu					
ruer price (\$/dieser nite	0.01		0 65(6)		0.01		0 65(6)		0 65(6)	
Total approximg costs (\$/km)	1.60	1.00/	1 00	70/	1.24	170/	1.00	210/	1.00	220/
Fuel costs $(\protect{P}/\protect{Km})$	<u>1.00</u> 0.67	<u>10%</u> 57%	1.09	<u>/ 70</u> 150/	<u>1.24</u> 0.75	<u>4270</u> 010/	<u>1.00</u>	<u>2470</u> 160/	<u>1.09</u> 0.64	<u>23%</u> 17%
Maintonanaa aaata (¢/km)	0.07	5170	0.90	15%	0.75	9470	0.00	40%	0.04	4170
(1) The percentage difference to Case 1 w	0.94	ted by divid	U.99	value of each	0.49	ir respectiv	0.41	the value	0.45	ne item
is the percentage dimension of case 1 was calculated by uniquity time value of each item in them respective case by the value of the same item is the percentage dimension of the percentage dimens										

Table 3.6: Simulated fuel costs based on Table 3.3 parameters

<sup>(1)</sup> The percentage difference to Case 1 was calculated by dividing the value of each item in their respective case by the value of the same item in the Case 1; e.g., for diesel-hybrids in Case 3, the 25% in Total operating costs was calculated by talking the percentage difference from 1.36 \$/km (total operating costs in Case 1) to 1.71 \$/km (total operating costs in Case 3).

<sup>(2)</sup> CNG calculated on an energy equivalent basis of diesel litres: Diesel 128,700 BTU/GAL and CNG 960 BTU/cu.ft (EERE, 2016)

<sup>(3)</sup> Average fuel from EERE (2016), from 2010 to 2015 (inclusive).

<sup>(4)</sup> 95% Conf. Interval (C.I.) calculated from time series sample variance for each individual fuel (source: EERE, 2016; Reports from Apr/2000 to Apr/2016) and the center point of C.I. based on average prices from 2010-15 (Case 2); i.e., for diesel price, the center value was assumed to be 0.97 \$/L (Case 2) for a range of [0.91,1.04] (based on diesel price sample variance from 2000 to 2015); for CNG price, the center was assumed to be 0.63 \$/L (Case 2) for a range of [0.61,0.66].

<sup>(5)</sup> The Upper Bound scenario took the following values: Diesel (UB) = center value (0.97) + confidence interval (0.069); CNG = center value (0.63) – confidence interval (0.024); the subtraction from CNG's mean price is intentional, in order to convey an increase in relative fuel prices (i.e., increase in the diesel-to-CNG ratio); and not a decrease, if CNG center value + confidence interval had been used.

<sup>(6)</sup> The Lower Bound scenario took the following values: Diesel (LB) = mean value (0.97) - confidence interval (0.069); whereas CNG = mean value (0.63) + confidence interval (0.024); the addition to CNG's center price was intentional, as per logic discussed in <sup>(5)</sup> above.

Analysis of Table 3.6 shows that, for WMATA, the total operating costs for diesel which are approximately on par with the two CNG bus types in Case 1 (original inflation-adjusted data), changes to a cost penalty of \$0.31 to 0.32-per-km when using 2010-2015 fuel prices, since diesel prices increased much more than CNG from Case 1 to Case 2. Even when considering the upper and lower bound estimates, operating costs for diesel buses at WMATA are more likely to be higher than CNG buses by \$0.15 to \$0.16-per-km more expensive in the lower bound Case 4, and by \$0.31 to \$0.32-per-km in the upper bound Case 3 (Table 3.6).

For NYCT, given 2010-2015 fuel prices, total operating costs increase by 5% for CNG systems and 22% for diesel-hybrid systems comparing Case 2 to Case 1 (Table 3.6). As a result, the reported cost advantage of the hybrid over CNG of \$0.41-per-km in Case 1, owing to better fuel economy and lower maintenance costs, decreases to a cost advantage \$0.20-per-km using 2010-2015 fuel prices. For both the lower and upper bound range estimates, diesel-hybrid operating costs were still consistently lower than CNG's. However, it should be noted that hybrid bus purchase costs were 23% higher than CNG buses at NYCT (Table 3.2). Given this discrepancy of lower operating costs but higher bus purchase costs of hybrids over CNG buses at NYCT, it would be interesting to evaluate if diesel-hybrid operating costs are low enough to offset the higher bus procurement costs over the service life of the bus. Considering that the average annual mileage for NYCT buses is around 45,000 km per year (Barnitt & Chandler, 2006), and the expected bus service life is 12 years (Lowell et al., 2007), the hybrid purchase cost premium of \$88,200<sup>9</sup> can be divided into a \$0.16-per-km cost over its expected service life. For NYCT buses, when adding capital cost premium in the various cases in Table 3.6, the lower operating costs hybrids diminish greatly, and are actually less advantageous in Case 3. It should be noted, however, that the preceding analysis of including purchase cost does not account for opportunity cost of capital, which is a substantial aspect to consider, since, on a dollar-for-dollar basis, bus capital costs are up-front costs (or must be financed, which involves added interest costs) and

<sup>&</sup>lt;sup>9</sup> Table 3.2 cost differential (hybrid diesel-to-CNG) inflation-adjusted to 2015 using CPI from BLS (2016).

thus more onerous to the operator relative to operating costs that are recurrent costs and paid throughout the vehicle's service life. Furthermore, refuelling infrastructure costs are significantly greater for CNG technology than for diesel-hybrids, which essentially uses the existing diesel infrastructure. The challenge in accounting for infrastructure costs for CNG is that, apart from timing of disbursements and longer service life of infrastructure than buses, costs also have to be allocated between all those buses that will use the infrastructure. Finally, from Figure 3.1 it can be seen that diesel-to-CNG ratio fell substantially from 2014 onwards, to less than 1 (that is, diesel is less expensive than CNG on an energy equivalent basis), which will tend to magnify the advantage of diesel-hybrid over CNG. Given the complexity of opposing forces from different cost parameters, and to properly address all these issues, a more sophisticated analytic framework would be required, such as in Lowell et al. (2007) that uses life-cycle cost modeling approach to compare various fuel and propulsion technologies for buses. For example, Lowell et al. (2007) accounts for inter-temporal cost schedule variations for different types of technologies, with different capital expenditures for bus and infrastructure investments, different re-build needs throughout the service life of vehicles, and different operating and maintenance expenses for a total value stream of future costs, all of which are then discounted to present values.

## 3.3 Discussion and policy implications

### 3.3.1 Summary of key findings

As discussed, CNG has been used worldwide in road transport, and particularly in urban bus transit, in an attempt to mitigate urban air pollution problems. Despite interest in CNG in terms of mitigating air pollution, adoption has been very limited, which raises important questions regarding the challenges and suitability of using CNG, especially from the perspective of bus fleet operators. Since investigating and understanding the challenges is an important research need, this chapter critically discussed the experience of five fleet operators in the US, Mexico and Chile. Taken together, the cases allowed a comparison of transit bus performance with CNG relative to diesel (and diesel-hybrid), in terms of key parameters including fuel economy and emissions, capital

and operating costs, and reliability; the effect of operating and test conditions on these performance parameters; and the evolution in performance over time, in response to factors such as more stringent emission standards and technological development. With respect to these issues, and based on the five cases, what key conclusions and policy implications can be derived?

Overall, the cases confirm that despite other potential benefits for CNG -- namely energy security and lower greenhouse gas emissions -- urban air quality was the primary rationale for adopting or considering CNG, especially since bus fleet operators have had to increasingly meet more stringent vehicle emissions standards; particularly for NOx and PM as was the case in US.

With respect to **bus and fuelling infrastructure choice**, the experience of NYCT and WMATA showed a diversified approach was used, with the bulk of these fleets continuing to rely on the conventional diesel fuel systems, with exhaust control and cleaner (low sulphur) diesel; this was due to space constraints in dense urban settings that led to high costs of CNG infrastructure and the challenges associated with adhering to safety regulations. At SunLine no modifications were necessary in their outdoor bus storage facility, showing that CNG refuelling infrastructure investments are climate and space dependent. Also, the cases showed that CNG buses required simpler and less expensive control measures to achieve more stringent emission standards than diesel. Yet, despite less expensive emissions control, CNG bus purchase costs were higher relative to comparable diesel buses, but much lower than diesel-hybrids.

In terms of <u>fuel economy</u>, despite other important technical factors, in practice operating conditions made a significant difference to relative performance between diesel and CNG buses. The CNG buses in the cases reviewed had a significantly (25%) lower fuel economy on an energy basis relative to their diesel counterparts in city operation. More importantly, both the US cases and MCMA showed that higher average speeds significantly improved performance for both diesel and CNG buses, but much more for CNG buses. In the case of the diesel-hybrid, its fuel economy advantage was

higher on more severe operating cycles, since regenerative braking and energy storage are most beneficial under these conditions. In WMATA's case, the CNG penalty relative to diesel disappeared when exhaust control measures like DPF and EGR were used to reduce PM and NOx emissions on diesel buses. From an analytic standpoint, the cases demonstrate the critical importance of making comparative evaluations under operating conditions that are similar and as close as possible to real-life conditions, in terms of vehicle vintage, weight, speed, engine power, and emissions control, among other factors.

As for <u>emissions performance</u>, CNG buses confirmed their significant potential for reducing PM and NOx relative to comparable diesel buses. The cases showed that technology and maintenance were key to emissions performance, and that operating conditions also profoundly affected emissions for all bus technologies, with emissions of critical pollutants such as CO, CO<sub>2</sub> and NOx being significantly higher on cycles with low average speeds, high frequency of stops, and high percentage of idling. In the case of PM, however, while emissions were the highest on the most severe cycle in terms of these parameters, PM showed a mixed trend with respect to speed. Further, the MCMA case showed that sulphur content made little difference to PM emissions in diesel buses, at the tested levels. Also, the US and Mexican cases showed that retrofitting older diesel buses with engines with advanced emission controls can cost-effectively achieve significant emission reductions of key pollutants, such as PM, with a performance that is comparable to CNG and newer diesel buses.

In terms of <u>reliability</u>, at NYCT EFS-related reliability was considerably lower for the CNG vehicle, despite the diesel fleet being a few years older at the time of the evaluation. Relative to NYCT, vehicular reliability at WMATA was significantly worse for both fuel systems, but EFS-related reliability was significantly better for WMATA relative to NYCT, especially for CNG. Given differences in operating conditions between the two cities, this suggests that higher operating speeds perhaps help EFS, but not necessarily total vehicle reliability. However, note that the vehicular reliability of the SunLine CNG fleet was more than double that for its NYCT and WMATA counterparts. Even though

the EFS-related MBRC figures for all CNG buses at WMATA were much higher than for their diesel counterpart, the EFS-related maintenance cost in \$/km was interestingly the same for all these cases, implying more costly though less frequent repairs for the CNG fleets. A comparison of reliability and operating cost performance across the US operators showed that, in the case of EFS maintenance cost, WMATA's reliability was much better than NYCT's, as mentioned, for both CNG and diesel, and correspondingly, WMATA's EFS costs in \$/km were also much lower. However, WMATA's total vehicular maintenance costs in \$/km were also much lower than NYCT's, despite WMATA's overall MBRCs performance being worse than NYCT's. At the same time, even though WMATA's EFS and vehicle maintenance cost was much lower than NYCT's, WMATA's EFS share of vehicle maintenance cost was the same as for NYCT. The examination of the cost breakdowns showed that effective cost management, as well as the considerably larger role for preventive maintenance inspection might explain the significantly lower maintenance costs, even for the vehicle as a whole, despite much more frequent road calls at WMATA relative to NYCT. At SunLine, maintenance costs, which were considerably lower than even WMATA's, despite its being a much older CNG fleet, reinforce the argument for cost control through effective management, preventive maintenance, and training.

Total <u>operating costs</u> at WMATA, in \$/km, were higher for CNG despite lower maintenance costs since fuel costs were higher, on account of the lower fuel economy for CNG relative to diesel. In the case of SunLine, CNG total operating cost was also much lower even than for WMATA and NYCT, not only because of low maintenance costs, but also because of significantly lower fuel cost, both due to a high fuel economy and a very low unit fuel price for CNG. In the case of the diesel-hybrid buses at NYCT, the total operating cost was considerably (23%) lower than for CNG, due to a lower maintenance cost, but mainly because of a considerably lower fuel cost, on account of a significantly higher fuel economy relative to CNG. Meanwhile, <u>fuel costs</u> were significantly lower for both fuel systems at WMATA relative to NYCT, because of the much higher fuel economy plus the significantly lower fuel price at WMATA. In the case of MCMA, the fuel cost in \$/km is slightly lower for CNG, despite a significantly poorer

fuel economy relative to diesel, because of a much lower CNG price relative to diesel in Mexico City. Overall, it appears that in the cases analyzed variation in the total operating cost across fuel systems is due primarily to variation in fuel cost, since maintenance cost varies considerably less; in turn, the variation in fuel costs is due to variations in fuel economy (due to operating conditions) and fuel unit price.

The key lesson from the sensitivity analysis is that while CNG fuel economy was generally lower than diesel, CNG fuel prices were low enough, relative to diesel, to make CNG financially viable relative to traditional diesel buses (i.e., operating costs of diesel buses were higher than CNG's, even considering a statistically relevant range of fuel price estimates). Meanwhile, even with updated fuel prices, and the cost advantages of diesel-hybrids over CNG diminishing, operating costs for diesel-hybrids were still consistently lower than CNG's, regardless of the fuel price scenario considered; this suggests that low maintenance costs for diesel-hybrids and the severe operating conditions of New City favor this type of bus fuel system over CNG. From an operational and operating cost perspective, this shows that diesel-hybrids can be superior options to CNG, especially considering that depot modifications and refuelling infrastructure investments are minimal for diesel-hybrids. To make matters worse for CNG's relative advantage, note that since 2015, diesel prices have reduced quite substantially, whereas CNG prices have not, and as a result diesel and CNG fuels are now being sold at near parity, on an energy equivalent basis. Thus, this suggests that the advantages of CNG bus fleets in this analysis will diminish if these fuel price differences persist into the future; whereas the advantages of diesel-hybrids will be enhanced. A caveat regarding diesel-hybrids is the higher bus purchase cost relative to CNG buses, which in turn, are also more expensive than traditional diesel buses.

## 3.3.2 Conclusion and policy implications

From a financial perspective, the complexity of evaluating various alternative fuel systems for buses, considering the above issues, is that these fuel and bus systems have different cost composition (capital and operating costs). A life-cycle cost modeling approach would be more appropriate to compare various fuel and bus systems, since it

integrates for each fuel and bus type, the different capital costs for bus and infrastructure investments, different re-build needs throughout the service life of vehicles, and different operating and maintenance expenses over the entire service life of buses and their supporting infrastructures. By using this framework, it would be possible to better address the question of an optimal diesel-to-CNG price ratio that would justify investment in CNG over diesel (or diesel-hybrid) systems.

The broader question is to incorporate a social perspective into these sets of issues, so that decision-makers are aware of the overall incremental costs associated with the particular fuel choice, and can weigh them against the environmental benefits of each fuel and bus option. Consider that, given the results discussed here, CNG's environmental outcomes can be attained at competitive incremental costs in comparison to diesel and diesel-hybrids, making this fuel an important option in controlling emissions in bus fleets without need for additional exhaust control measures or improved fuel quality. At the same time, the challenge for CNG is that "cleaner" diesel and more advanced diesel engine technologies are making headway in order to comply with stricter emission standards set by environmental authorities, particularly by using low sulphur diesel in combination with exhaust emissions control measures. Further, since diesel technologies have lower vehicle purchase and infrastructure costs, the cost-effectiveness advantages of CNG may diminish into the future. In this regard, consider the renewal of Delhi's public bus transit fleet that started in 2006-2007, with the purchase of 3200 low-floor CNG buses, at a total cost of 16.3 billion Indian Rupees (DTC, 2015), or approximately \$357 million US dollars using the average exchange rate over the period (Reserve Bank of India, 2016). Given the incremental price of a CNG bus over diesel-equivalents, which can be safely estimated at around 20%, this represents a non-trivial added cost of fleet renewal of approximately \$71 million. This raises the questions of how cost-effective air quality measures in Delhi have been: do the PM and NO<sub>x</sub> emissions reduction outcomes justify the incremental price tag? Are there more effective ways of achieving the same emissions outcomes, such as by using ULSD? How many fewer buses are being put into service, and with what effects, given

these incremental costs? These questions will be addressed in Chapter 6 when evaluating the cost-effectiveness of CNG in Delhi.

Considering the above, a critical lesson for policymaking is the need to understand the linkages between various policy goals, such as improving air quality, and the financial and operational challenges related to setting up each fuel option, how different operational and contextual parameters will affect stakeholder behavior, and so forth. For example, if the policy goal is to eliminate market distortions by incorporating key environmental effects of each technological option, policy-making must be based on understanding of the critical incentives that will motivate stakeholder choices. Particularly from the fleet operator's perspective, this means balancing broader public concerns, such as emissions, with operator considerations of financial feasibility and operational performance. This is especially important where, there are fleet operators that are private companies, for whom cost minimization is a top motivating factor in deciding technological options.

# Chapter 4: Operational and financial performance of Delhi's CNGfueled public bus transit fleet: A critical evaluation

#### **Chapter overview**

As shown in Chapter 1, Indian cities, and particularly Delhi, have been characterized by poor air quality since the 1990s, which has led to a wide range of policies implemented nationally in India and in Delhi. Among these policies in Delhi was the Supreme Court of India ruling in 1998 that mandated that all public and for-hire motor vehicles (buses, taxis and auto rickshaws) in the city to be powered by CNG. Because of this ruling, all of the city's buses, both publicly owned and operated by the Delhi Transport Corporation (DTC), had to be converted to run on CNG by 2001, and, as briefly shown in Chapter 1, there were various resource, logistical, and institutional challenges to accomplish this CNG implementation. However, in light of this mandate, and as was discussed in Chapter 2, the literature on the implementation of CNG in Delhi's public vehicles has focused almost exclusively on its emissions outcomes, with little if any attention devoted to its operational or financial aspects. In the case of DTC, analyzing the operational and financial performance of its bus fleets on CNG is important because it is this performance that critically determines the bus operator's policy responses. Building on this important research need and the many important issues raised in Chapter 3 related to CNG implementation and evaluation, I conduct a critical evaluation the operational and financial performance of DTC's CNG-fueled bus fleet in this chapter, in order to assess how the conversion from diesel to CNG has affected fleet operations and finances. This chapter, together with the life-cycle cost assessment of CNG buses in Chapter 5, will help assess the cost-effectiveness of Delhi's CNG policy in Chapter 6.

### 4.1 Introduction

Indian cities have been characterized by poor air quality since the 1990s. In Delhi, for example, suspended particulate matter levels have exceeded World Health Organization (WHO) guideline limits almost daily since the 1990s. Levels of PM<sub>10</sub>

(particulates below 10 microns diameter), which are strongly linked with respiratory and cardio-vascular illnesses and deaths, also exceed the WHO limits (CPCB, 2015). A global survey of urban air pollution (WHO, 2014a) showed that Delhi had the highest annual average levels of fine particulates (PM2.5), which pose the most serious health risk. In response to this problem, a wide range of policies has been implemented since the early 1990s to address air emissions from urban transport. Delhi being the national capital, and given its serious air quality problems, many of these policies were first implemented there and in the other major metropolitan centres, and then in the rest of the country in a phased manner. These policies have included increasingly stringent vehicle emission and fuel quality standards, vehicle inspection and maintenance (I&M) to control in-use emissions, and the phasing out of old commercial vehicles (CSE, 2002b; BIS, 2002; TERI, 2002; Kojima, Brandon and Shah, 2000). A Supreme Court of India ruling in 1998 mandated that all public and for-hire motor vehicles (buses, taxis and auto rickshaws) in Delhi be powered by compressed natural gas (CNG) (Supreme Court of India, 1998).

As a consequence of this ruling<sup>10</sup>, all of the city's buses, including those that were publicly owned and operated by the Delhi Transport Corporation (DTC), had to be converted to run on CNG over a highly compressed time frame, by March 31, 2001. Due to resource, logistical, and institutional challenges, discussed later, CNG implementation on Delhi's buses began only in 1999-2000. In any event, Delhi today has the largest bus fleet in India, currently numbering, according to official statistics, around 60,000 buses (GNCTD, 2012) -- and given that all of the city's buses run on CNG -- one of the largest bus fleets running on this, or indeed any alternative fuel, globally. Further, DTC, the focus of our study, has the second largest publicly owned and operated urban bus fleet in India, with around 5800 CNG buses currently in operation (GNCTD, 2012). It serves the National Capital Territory of Delhi, as well as

<sup>&</sup>lt;sup>10</sup> According to the ruling (Supreme Court of India, 1998), no 8-year old buses could ply in Delhi except on CNG (or "other clean fuels") beyond April 1, 2000, and further, the entire bus fleet in Delhi was required to be "steadily converted" to run on CNG by March 31, 2001.

neighbouring cities in surrounding states, and carried 4.5 million passengers daily in 2010-11 (CIRT, 2012).

The literature on the implementation of CNG in Delhi's public vehicles has focused almost exclusively on its emissions outcomes (for example, Kathuria, 2005; Jalihal and Reddy 2006; Chelani and Devotta, 2007; Reynolds and Kandlikar, 2008; Reynolds et al., 2011; Narain and Krupnick, 2007; Kumar and Foster, 2009), with little if any attention devoted to its operational or financial aspects. Sen et al. (2007) discuss this gap in research, pointing to the lack of sufficient focus on the part of decision-makers on the creation of financially viable and self-supporting urban transport systems. Further, we argue, while the emissions outcomes of CNG implementation are important from a societal perspective, and are an important focus for policy evaluation, it is also important and useful to critically evaluate this, and indeed, any such policy from the perspective of vehicle users and operators, because it is the policy responses of these actors that crucially determine the extent to which implementation, and the associated emissions reductions, actually occur and are successful.

More particularly, analyzing the operational and financial performance of bus fleets on CNG or any other alternative fuel is important because it is this performance that critically determines the bus operator's policy responses. Such an analysis is also useful from a policy perspective, given the need to provide quality, convenient and affordable bus services within the constraints of limited budgets on which public transit operators typically rely, to prevent the migration of ridership to private motor vehicles, with all of their negative impacts. In this regard, note that, while buses and other public transit modes still account for a significant share of passenger trips (43% in Delhi), their mode share has been declining significantly, due to the growing role of personal motor vehicles (WSA, 2008a).

In view of the foregoing, we critically evaluate the operational and financial performance of DTC's CNG-fueled bus fleet, in order to assess how the conversion from diesel to CNG has affected fleet operations and finances. We also critically discuss the implementation experience in terms of the associated infrastructure, logistical and institutional challenges. The 10-plus years of experience accumulated by DTC, a major public transit operator, of such a large-scale conversion of its bus fleet to CNG, provides a valuable opportunity for this retrospective analysis. Apart from addressing an important research need (such a post-implementation assessment has not been reported on so far), our analysis will hopefully be useful to decision makers and urban bus transit operators in contexts similar to India's, by drawing lessons for the long-term viability of large-scale conversions of bus fleets to CNG, for comparison with other CNG bus transit fleets, and for informing techno-economic and environmental analyses of CNG bus transit operations. In particular, our study, coupled with others that have focused on its emissions outcomes, should help assess the cost-effectiveness of Delhi's CNG policy.

## 4.1.1 CNG implementation on DTC's bus fleet

Despite the 1998 Supreme Court mandate to convert the entire bus fleet in Delhi to run on CNG by March 31, 2001, only around 150 CNG buses had been put into service at DTC by then. Implementation gained momentum only in 2001-02, and was completed only by 2003-04 (Figure 4.1a). There were serious logistical and technological challenges related to the large-scale fuel-system conversion that DTC (and even more so, Delhi's private bus operators) faced, particularly with retrofitting diesel engines to run on CNG. Therefore, they mostly opted for factory-built CNG buses, which were supplied by two of the largest Indian bus manufacturers, Tata Motors and Ashok Leyland. More generally, there was a "sequencing problem", namely that of implementation depending on the availability of financial resources for conversion as well as reliable refueling infrastructure and vehicle technology (Bell et al., 2004). The bus manufacturers and infrastructure providers wanted assurances of demand for the technology, without which they were reluctant to invest in production. In order to break this supply-demand vicious-cycle, DTC, the bus manufacturers and Indraprastha Gas Limited (IGL), who provided the dedicated refueling infrastructure in DTC's depots, created a task force to coordinate their respective roles. Despite the challenges, CNG

implementation was accomplished on Delhi's bus fleet over a short period, from 2001 to 2004 (Bell et al., 2004; Patankar & Patwardhan, 2006).

Table 4.1 highlights the key vehicle and engine attributes of the diesel and CNG bus technologies used at DTC during our analysis period. The first generation of CNG buses, which we refer to as the Standard CNG model, were introduced from 2000 to 2004, when they replaced a fleet of similar configuration diesel buses (Table 4.1), with most of these CNG buses being inducted in the 24-month period from 2001-02 to 2002-03. From the transit user's perspective, the quality of the buses changed little due to this introduction, apart from tailpipe emissions due to the substitution of diesel fuel with natural gas. From 2007-08, DTC began a fleet modernization process by launching two procurement cycles for Low Floor (LF) CNG buses. The introduction of LF CNG buses, which commenced towards the end of 2007, was part of a process that sought to replace the fleet of ageing standard CNG buses, 10% of which were over DTC's target service age<sup>11</sup> in 2006-07 (CIRT, 2007), while also offering higher quality bus service generally, and in particular, improved and more convenient accessibility for the aged and infirm, as well as young children. Key bus attribute changes included a 400mm floor height for LF versus 800mm for standard buses, automatic transmission, full air suspension, air conditioning (AC) option for higher fares, and a target service life of 12 years (or 750,000 kilometers). From 2007 to 2011, 3,700 new LF buses were put into operation at DTC, of which 67% were non-air conditioned buses (CNG LF) and 33% were air-conditioned (CNG LF/AC) (Table 4.1). In 2010-11, of the 4,300 operational buses, nearly 46% were standard CNG buses, 53% were LF buses, and less than 1% of the buses were diesel powered, which were in the process of being decommissioned.

<sup>&</sup>lt;sup>11</sup> The target service life of standard CNG bus set by DTC was 8 years or 500,000 kilometres (CIRT, 2012).

	Diesel Standard	CNG Standard	CNG LF	CNG LF/AC
Model Year	1997/98	2001/02	2007/08	2007/08
Fuel Type	Diesel	CNG	CNG	CNG
Engine model max HP	Hino, Leyland 370, Tata 692 95-110	IVECO 8060.05; TATA LPO 1510 CGS/55 120-150	Cummins LB BGe230/30 230	Cummins LB 250
Compression ratio	17:1 - 17.9:1	10.5:1 (Est.)	10.5:1	10.5:1
No. of cylinders	6	6	6	6
Engine displacement	5.9L	5.9L	5.9L	5.9L
Emissions control / Standards	1996 norms	BS-I / BS-II	BSIII Complia	ant; Engelhard
EMISSIONS CONTON / Standards	(pre BS-I)	catalytic converter	Oxidation	n Catalyst
Fuel Capacity	160-165 L	90-96 kg	108 kg	108 kg
Gross Vehicle Weight (tonnes)	15.2	15.3	16.2	16.2
Length (metres)	~10.1	~10.7	12	12
Width (metres)	2.4	2.4	2.6	2.6
Passenger capacity	~65	68	70	70
Inflation adjusted bus cost (USDx10 <sup>3</sup> )	47.4	59.6	126	149
Percentage of fleet on road (2010-11)	1%	46%	43%	10%

## Table 4.1: CNG bus technologies at DTC

Source: Delhi Transport Corporation

The standard CNG buses, introduced in 2000, had a spark-ignited engine with a power rating 31% higher than its standard diesel counterpart, and CNG storage in 8-12 gas tank cylinders located under the passenger floor. Their purchase cost was 26% higher than that of the diesel buses they replaced. The LF CNG buses included an integral body chassis, rather than being built on a truck chassis as was the case for the standard CNG and diesel buses, a larger floor area, rear mounted engines with nearly double the power rating, and roof-mounted CNG storage tanks. Also, these buses conformed to the Bharat Stage III and IV emission standards, which are comparable to EURO-III and IV norms. The LF CNG buses cost USD 126,000-149,000, more than twice the average price of USD 60,000 for a new standard CNG bus, based on recent procurements by other Indian transit operators (Government of India, 2009a, 2009b, 2009c, 2011, and 2012).

Operation, storage and maintenance of DTC's bus fleet are conducted from 49 depots across Delhi. The majority of these depots have dedicated CNG refueling infrastructure on site, which includes diesel-powered compressors, CNG buffer storage tanks, and usually two fast-fill dispensers per depot. Natural gas feedstock costs, staffing, operation, and maintenance of DTC's CNG refueling infrastructure is provided by IGL, which sells CNG to DTC at a negotiated price. Each depot serves only one type of bus technology (i.e., standard CNG, or LF CNG buses). DTC staff maintain the standard CNG buses at the depots, while overhauling and reconditioning is done at two central workshops.

When the LF CNG buses were procured, starting in 2007-08, DTC contracted with the bus manufacturers, requiring them to be responsible for maintaining these buses. This outsourcing was a clear departure from the traditional practice of "in-house" maintenance, as was the case when CNG was first implemented in 2000, and allowed for technological risk to be transferred to the bus manufacturers while allowing DTC to reduce staffing needs substantially and focus staffing resources on transit operations. Under the agreement with DTC, the bus manufacturers employ their staff for routine and preventive maintenance of the LF CNG buses on site at the depots, and bear the associated costs of labour and parts. They carry out overhauls and more specialized maintenance at outside locations.

## 4.2 Analytic framework, methodology and data

In this paper, we present our analysis of DTC's bus fleet performance over the period 1989-90 to 2010-11 -- that is, from roughly ten years prior to CNG implementation until 10 years after -- in order to assess how fleet operational and financial performance was affected by the replacement of diesel with CNG, and changes to bus technology. Our analysis was based on those factors and performance measures most affected by fuel system change; namely, bus transit service provision and utilization (in terms of fleet size and age, carrying capacity-kilometres, and passenger-kilometres); fuel economy (in energy equivalent terms, for diesel versus the CNG operation, and for standard CNG

versus LF CNG versus LF-AC CNG buses); capital and operational expenditures (related to fuel, maintenance and labour); and reliability (in terms of breakdowns, for the fleet, and for standard CNG versus LF CNG buses).

The data for our analysis was based on annual statistics gathered at DTC by the authors during a field trip from 02/2010 to 04/2010, by compiling monthly reports published internally for DTC management, as well as on yearly reports published by the Central Institute of Road Transport (CIRT, 1992-1995; CIRT, 1997-2012). Unless stated otherwise, all annual time series data were for the relevant financial year (so, a statistic for period 2001-02 refers to data from April 1, 2001 to March 31, 2002). Further, the data we used and our calculations focused only on DTC's bus operations, excluding those related to buses they hired (shown as DTC Hired in the CIRT reports).

When calculating operational and financial performance measures on a per-kilometre basis, we used effective or revenue earning kilometres, henceforth referred to as kilometres. Fuel economy figures were as reported by CIRT and DTC (see below), and were based on gross kilometres. CNG fuel consumption was converted to litres diesel equivalent, based on the mass based energy content of the two fuels, and the density of diesel in Delhi (SIAM – Society of Indian Automobile Manufacturers, 2013). All financial parameters reported in Indian Rupees (INR) were, unless otherwise stated, inflation adjusted based on the Consumer Price Index for Industrial Workers (CPI-IW) published by the Labour Bureau of the Government of India (Government of India, 2013), with 2010-11 as the base year (the average annual inflation rate during our 20-year analysis period ending 2010-11 was 7.63%). Finally, values expressed in US dollars (USD) were derived from the inflation adjusted INR values, using the average exchange rate for 2010-11 (Reserve Bank of India, 2013).

While the annual DTC and CIRT data were useful for analyzing financial and operational performance, they were mostly aggregated for the entire fleet. Higher frequency data, available on a monthly basis, were accessed from DTC and used for analyzing fleet fuel economy and reliability, and to compare different bus technologies

that were operational contemporaneously. The reports accessed at DTC provided statistics on aggregate fleet performance, as well as statistics disaggregated at the depot level for key operational and financial parameters such as staffing, routes, fleet size, kilometers operated, key expenditure items, and material consumption (fuel, tyres, oil). Information was also gathered from DTC managers at selected bus depots, relating to systems characteristics, maintenance, refueling infrastructure and vehicle technology, fuel economy (FE) and reliability. The sites at which data was collected were chosen to ensure representativeness of the geographical regions served by DTC's urban transit routes, and accounted for the fact that DTC operated two types of CNG bus technologies in 2010, for which bus attributes varied significantly.<sup>12</sup> Finally, interviews were conducted with DTC management, to elicit their perspectives based on their experience with CNG implementation.

# 4.3 Results and discussion

# 4.3.1 Operational performance

## i) Service provision and utilization

Figure 4.1a shows the buses added and scrapped, and the resulting fleet size and age, while Figure 4.1b shows the changes in carrying capacity-kilometers<sup>13</sup> and passengerkilometers over our 20-year analysis period. From 1989-90 to 1999-00, when DTC's buses were exclusively diesel powered, there was no significant capital investment in fleet renewal, and the average age of DTC's fleet steadily increased to almost 8 years, DTC's benchmark for maximum service life for standard buses. From 1999-00 to 2002-03, as highlighted in Figure 4.1a, DTC's fleet went through a transition period, in which

<sup>&</sup>lt;sup>12</sup> The DTC sites visited for data collection were the Strategic Business Unit, DTC's management office for the LF bus fleet, in Hauz Khas; DTC's Central Workshop I on Banda Bahadur Marg, where planning for the standard buses, and re-conditioning, re-treading and bus body repairs, were conducted; the Kalkaji Depot and Hari Nagar Depot II (operating the standard buses), and the Sukhdev Vihar Depot, Hari Nagar Depot I and Rohini Depot I (operating the LF buses).

<sup>&</sup>lt;sup>13</sup> Capacity-kilometres are calculated by multiplying the total seating and standee capacity of all buses by the kilometres operated (CIRT, 2009).

fleet capacity was affected by, among other factors, the Supreme Court's scrappage and CNG mandates. The scrappage mandate requiring the retirement of buses older than eight years preceded the mandate that all buses be converted to CNG by one year (see Footnote 10).

As a result of these mandates, DTC's bus fleet size, which was already in decline, was reduced precipitously, by about one-third, from 3,000 buses to 2,000 buses, during the transition period (Figure 4.1a). CNG implementation accelerated from 2001-02, as noted, and as new CNG standard buses were inducted into the fleet, the fleet size returned to pre-transition period levels, the fleet composition changed from mostly diesel to mostly CNG, and the average fleet age decreased significantly. DTC's CNG bus technology was predominantly homogeneous, with the fleet being comprised only of standard CNG buses, and vehicle scrappage and fleet renewal being low, during 2001-2009, therefore allowing the use of operational statistics aggregated at the fleet level over this time to analyze the performance of standard CNG buses over their service life. CNG LF bus introduction gained momentum in 2009, and fleet-wide performance statistics start to be increasingly influenced by this technology from this time onward. In 2010-11, DTC's fleet comprised around 6,200 CNG buses, but only around 4,300 were in use due to about 1800 standard CNG buses being over their target service life of 8 years (or 500,000 kilometers) and deemed unfit for operation (CIRT, 2012).<sup>14</sup>

<sup>&</sup>lt;sup>14</sup> While diesel buses were being used in inter-state operations, these buses accounted for less than 1% of the operational fleet in 2010-11, and so have limited impact on fleet-wide statistics.



Figure 4.1: Service provision and utilization

Carrying capacity-kilometres, as well as passenger-kilometres were both already in decline since 1989-90 (Figure 4.1b), but with the sharp reduction in the bus fleet due to the scrappage and CNG mandates, there was an equally sharp (40%) reduction in these measures in the beginning of the transition period. Further, the average load factor<sup>15</sup> was extremely high, reaching almost 100% for two consecutive years. With the induction of standard CNG buses from 2000-01, carrying capacity-kilometres increased steadily until 2005-06, but interestingly, passenger-kilometres did not at the same rate. Beyond 2005-06, and until 2009, when the CNG LF buses were introduced, both capacity-kilometres and passenger-kilometres again dropped precipitously, as the standard CNG buses that were past their target service life were taken out of service, but also possibly because of the significant increase in bus breakdowns during this period, as we discuss in a subsequent section. These measures began to revive again, after the introduction of the new CNG LF buses. Overall, what is striking is the decline in passenger-kilometres over the 20-year analysis period, and that this measure remains flat, even while the fleet size was being augmented significantly with new CNG

<sup>&</sup>lt;sup>15</sup> The load factor is the ratio of passenger-kilometres to carrying capacity-kilometres, expressed as a percentage (CIRT, 2009).

buses.<sup>16,17</sup> This trend likely reflects (and is reflected in) the steady decline in public transit shares reported in WSA (2008a), which is likely due to a range of factors, but also possibly the growing role of the Delhi metro.

# ii) Fleet fuel economy

The average fuel economy of standard CNG buses was 45% lower than that of DTC's diesel-only fleet, in km/L diesel equivalent terms (Figure 4.2a). This significant reduction in fuel economy in equivalent energy terms is likely due to several factors. First of all, while the Otto cycle, on which the spark-ignited (SI) CNG engines operate, has a higher thermal efficiency than the diesel cycle at the same compression ratio, diesel engines use much higher compression ratios in practice, because they compress only air, and are thus not susceptible to auto-ignition (as in the case of SI engines), and are therefore able to achieve higher thermal efficiencies than SI engines. CNG is a much better SI engine fuel than gasoline, because of its very high octane rating, and therefore SI CNG engines can use a compression ratio higher than that of gasoline engines, but still, much lower than that of diesels (in the case of DTC's buses, as can be seen from Table 4.1, 10.5 for the SI CNG engines, relative to the 17.1-17.9 for the diesels which they replaced). Additionally, SI engines are characterized by poor part-load efficiencies because of throttling losses. In the case of the DTC CNG buses, it is also possible that sub-optimal engine technology, as well as the additional weight of the CNG fuel tanks contributed to the significantly lower fuel economy. Yet another important factor is likely the low speed, stop-go characteristics of DTC bus operations; note in this regard that congestion is a major and growing challenge in Delhi, where the average peak-hour speed -- for all modes -- was 16 km/h, according to WSA (2008a). Chassis dynamometer and road tests have shown that the fuel economy of CNG buses with

<sup>&</sup>lt;sup>16</sup> One factor contributing to the low passenger-kilometres from 2000 onward is that the average annual distance covered by DTC's CNG buses was approximately 74,000 kilometres (CIRT 2012), as opposed to 82,338 kilometres on its diesel buses during the 1990s, according to CIRT data. Also, note that the corresponding figure in other Indian metro cities is 85,000 kilometres (as against only 56,000 kilometres in US cities -- Lowell et al. 2007).

<sup>&</sup>lt;sup>17</sup> Note that, while the average load factor for DTC was 71% from 2000-01 to 2010-11, it was 64% for Bangalore's and Mumbai's public bus fleets, and 81% for Chennai's (CIRT, 1992-1995; CIRT, 1997-2012).

similar specifications to diesel buses are typically 4-25% lower depending on the vehicle operating cycle (Lowell et al. 2007; Wayne et al. 2008; Clark et al. 2009), with the gap being greater at slower average speeds.<sup>18</sup>



<sup>(</sup>a) Yearly Fuel Economy for Standard diesel and CNG buses

(b) Monthly Fuel Economy, Standard versus LF CNG buses

# Figure 4.2: Fleet fuel economy

While the fuel economy of DTC's fleet reduced significantly as a result of the switch to CNG on their standard buses, it deteriorated even further with the introduction of LF buses in 2007, a process that gained momentum in 2009, as noted. The key differences in the LF CNG buses relative to the standard CNG buses included a one tonne higher vehicle weight, and a 70% higher engine power rating (Table 4.1). Based on data collected at DTC, the average fuel economy was 17% and 35% lower for the LF (non AC) and LF-AC CNG buses respectively, than for the standard CNG buses during the same evaluation period in 2009-10. In the case of the AC buses, there was greater seasonal variation due to the air conditioning system.

<sup>&</sup>lt;sup>18</sup> The US experience has shown that, because SI CNG engines are far more sensitive to operating conditions than their diesel counterparts, higher average speeds significantly improve performance for both diesel and CNG buses, but much more for CNG buses. Further, the fuel economy of state-of-the-art SI CNG engines actually approach that of diesels, under favourable operating conditions, and particularly when stringent emission standards call for more advanced exhaust treatment on diesels, with negative implications for their fuel economy -- see for example Melendez et al. (2005) in this regard.

#### iii) Reliability

One measure of service reliability is the rate of breakdowns.<sup>19</sup> In order to evaluate the reliability of alternative fuel systems on buses, it would be useful to focus on breakdowns that are related to the fuel system, engine, air intake, cooling, and exhaust. However, the CIRT breakdown statistics (CIRT, 2001-2012) are disaggregated only in terms of mechanical, electrical and tyre related breakdowns, with mechanical breakdowns comprising those related to transmission, engine and brakes. So, for the evaluation of reliability (in terms of breakdowns) of CNG buses relative to their diesel predecessors over our analysis period (Figure 4.3a), we used CIRT statistics, excluding tyre related breakdowns. Since vehicle retirement influenced bus availability, and likely distorted the statistics during the transition from diesel to CNG, the period from 1999-00 to 2001-02 was excluded from this analysis.



Figure 4.3: Reliability

(b) Monthly breakdown rate, Standard versus LF CNG buses

<sup>&</sup>lt;sup>19</sup> A breakdown is defined as stoppage of a bus (while in operation) due to mechanical defects or other failures rendering the bus unfit to operate without attention to it, irrespective of the time involved (CIRT, 2009).

Interestingly, the diesel fleet showed an improving trend in breakdown rates (Figure 4.3a), despite the increasing age profile of the fleet (Figure 4.1a), during 1990-91 to 1998-99. On the other hand, the breakdown rate worsened significantly for the CNG fleet, as the fleet aged, with the average from 2002-03 to 2010-11 being 125% higher than for the diesel fleet. Since effective kilometres per bus on road is a stronger determinant of wear and tear, and therefore breakdowns (and maintenance costs), we have plotted this measure for reference in Figure 4.3a; note that the significantly higher breakdown rates for the CNG buses occurred despite lower effective kilometres per bus for CNG relative to diesel buses, and a significant decline in this parameter from 2005-06. Also worth noting is the significant increase in electrical breakdowns, likely due to the spark plugs and related electrical systems on the SI CNG engines, a problem that has also been observed elsewhere (Chandler, 2006). New York City Transit (NYCT), which operates a fleet of over 4,000 buses, including a large number of CNG buses, has set performance criteria, which among other things require that buses should be operated at least 6,400 kilometers between roadcalls on average (Barnitt and Chandler, 2006). This benchmark, which is equivalent to 1.56 breakdowns per 10,000 km (for all breakdowns)<sup>20</sup>, was exceeded beyond around 2010, when many of the standard CNG buses were past their 8-year target service life, and the LF CNG buses were being rapidly inducted into DTC's fleet (see below).

In Figure 4.3b, we compare the breakdown rates of the standard and LF CNG buses during the time period December 2007 to December 2009, based on monthly breakdown statistics accessed from DTC, which did not discriminate between types of breakdowns, and so include all breakdowns. The breakdown rate for the LF buses, which on average was 2.4 breakdowns per 10,000 km, was 41% worse than for the standard CNG buses (and consistently worse than NYCT's benchmark of 1.56 breakdown-per-10,000 km throughout), despite the very low (1.5 years) average age of the former buses, and roughly 67% of the standard CNG buses being older than 8 years during this period. This poor performance on the part of the LF buses, which possibly is

<sup>&</sup>lt;sup>20</sup> NYCT's CNG buses had an average breakdown rate, for all types of failures, of just 1.1 breakdowns per 10,000 km (Barnitt, 2008).
due to the fact that they were being implemented for the very first time in India when they were introduced at DTC, has raised public concern and negative publicity, as evidenced by media reports (for example, The Times of India, 2013).<sup>21</sup> Our discussions at DTC revealed that the bus manufacturers had likely underestimated, and were overwhelmed by the maintenance costs of the LF CNG buses. As for the drop in breakdown rates for 2010-11 (the last data point in Figure 4.3a), note that the data for 2011-12 (not shown in the figure) in fact goes back up to the 2009-10 level. While Figures 4.3a and 4.3b, taken together, show higher breakdown rates for the LF CNG buses, it might perhaps be best to wait until around 2016-17 in order to properly evaluate LF buses in this regard, over their full service life.

# 4.3.2 Financial performance

# i) Capital expenditures

DTC's capital expenditures during our analysis period from 1989-90 to 2010-11 were mostly affected by the purchase of roughly 3,000 standard CNG buses in 2000-2004, and then by the purchase of 3700 LF CNG buses starting in 2007-08. There was limited impact on DTC finances due to other capital expenditures, such as those tied to the refueling infrastructure or changes to depots and maintenance areas for CNG operation. The refueling infrastructure was put in by IGL, the gas supplier, as discussed earlier, and so DTC did not incur any directly related capital expenditures. However, since DTC pays IGL, which is a publicly traded company, for the fuel that is provided on site, the CNG price likely factors in IGL's recovery of their capital investment, and so, DTC does pay indirectly for the refueling infrastructure. While we did not have access to information on expenditures on depot modifications related to the CNG refueling infrastructure, we evaluated them indirectly based on the changes to asset values of building infrastructure at DTC during the transition from diesel to CNG. Despite the value of building, plant and machinery having increased by 43% from 2001 to 2004

<sup>&</sup>lt;sup>21</sup> Our research at DTC, based on a sample of 300 LF CNG buses, showed that, from February 2009 to January 2010, mechanical, tyre and electrical problems accounted for 45, 27 and 17% of the breakdowns, respectively.

(based on CIRT data), these assets accounted for only 12%, whereas DTC's bus fleet did so for as much as 85%, of DTC's total fixed assets in 2004. This situation is partly due to the fact that Delhi's climate allows for open storage of CNG buses and semiopen maintenance areas, which require only minimal investment in depot infrastructure. In other contexts, especially in colder climates, most depot upgrades related to CNG buses are due to installation of ventilation systems, re-wiring of electrical systems, and gas detection systems to mitigate the risks of gas leakages, explosions and fires (Barnitt & Chandler, 2006; Chandler et al., 2006).

In terms of bus capital expenditures, purchase costs (Table 4.1) were based on information collected at DTC, as previously noted. Standard diesel bus costs were based on the last procurement of this bus type at DTC in 1998-99, while the costs for standard CNG buses are the average price from 2000-2003, when DTC's fleet was being converted to this fuel. The cost premium for the standard CNG bus relative to its diesel predecessor was 26%, mostly related to the fuel system and engine, apart from which both bus types had similar vehicle attributes (Table 4.1). While DTC has not procured standard buses in recent years, their attributes have changed little, apart from vehicle emission standards, partial use of air suspension, and possibly GPS units. The average current cost of a standard diesel bus in India is USD 60,000, and the cost premium of a standard CNG bus is 20%, based on data for Indian transit operators that procured both standard diesel and CNG buses (Government of India, 2009a, 2009b, 2009c, 2011, and 2012). The CNG bus premium has therefore not changed significantly since the early 2000s.

DTC's costs for the non-AC LF CNG buses increased 27%, from INR 4.1 million (USD 123,000) to INR 5.2 million (USD 126,000) per bus, between the first and second procurements in 2007 and 2009; note that inflation was only 8.7% (based on the consumer price index referred to earlier) over the same period. Meanwhile, costs for the LF AC CNG buses actually decreased 9% over the same period to INR 6.2 million (USD 149,000), probably owing to the increased scale of DTC's purchase of 775 buses in the second procurement, relative to the first one, in which only 25 LF AC CNG buses were

purchased. Assessing the marginal cost of CNG LF buses relative to their diesel counterparts at DTC is a challenge because no new diesel buses have been purchased since adopting CNG. Data from Municipal Corporation Ludhiana (2013), which procured similar specification LF non-AC diesel buses at the cost of INR 4.8 million (USD 105,000) per bus from the same manufacturer that supplied most buses at DTC, shows that the cost premium of a non-AC LF CNG bus at DTC is about 20%; note, however, that, apart from basic engine and model specifications in the Ludhiana case, not much is known about other vehicle attributes, which also could affect costs. Data for Indian transit operators that purchased both CNG and diesel buses (Government of India, 2011) shows that the average cost premium for CNG buses, and higher premiums for the less expensive standard CNG buses, relative to their diesel counterparts, thus corroborating the comparison between DTC and Ludhiana. These findings in the Indian context are similar to the cost premiums of 8-20% for CNG buses observed in the USA and Europe (Posada, 2009).

# ii) Operating expenditures

The area graph in Figure 4.4a shows the key operating expenditures from 1989-90 to 2010-11. Counting from the bottom of the figure, the first two wedges, put together, represent the total labour costs; the second and third wedges represent the workshop and maintenance labour plus the expenditure on automobile spare parts, and the third and fourth wedges the total expenditures on all materials, except fuel, which are shown in the top wedge. Taken as a whole, the expenditures in Figure 4.4a represent the total operating expenditures, less taxes, interest, depreciation and other miscellaneous expenses. In Figure 4.4b, we show the fuel costs, and the expenditures on workshop and maintenance labour plus all materials except fuel, per effective (or revenue generating) kilometre (note that the CIRT data does not indicate the expenditures exclusively related to bus maintenance, except for the workshop and maintenance

labour).<sup>22</sup> For our analysis, we also compare the prices of diesel and CNG on an energy equivalent basis over the analysis period, in Figure 4.5a.



Source: India from 1998-99 to 2009-10 (DTC); India from 2009-10 to 2010-11 (CIRT 2011, 2012); India Jun-Jul'12 (MyPetrolPrice.com, 2013); USA: EERE (2013)

(a) Fuel prices paid by DTC, annual average

(b) CNG-diesel price ratio (P<sub>CNG</sub> ÷ P<sub>diesel</sub>)

#### Figure 4.5: Comparative fuel prices – diesel and CNG

<sup>&</sup>lt;sup>22</sup> Data for various transit performance parameters needs to be provided more consistently for different parts of DTC (DTC urban, rural, hired, NCT), both to allow the different parameters to be correlated more effectively, as well as to enable a proper comparison across different transit operators. With specific reference to breakdowns and maintenance costs, data needs to be provided in a much more disaggregated fashion, to enable assessment of alternate fuel systems in terms of these parameters.

The first point to note is that interest payments (not shown in Figure 4.4a) alone amounted to around 50% of total operating expenditures. This high debt burden is mostly due to the cost associated with the purchase of standard CNG buses from 2000-01 to 2003-04, and the LF CNG buses starting in 2007-08. The CIRT data shows that these purchases were financed mainly through government loans, at annual interest rates of 10.5-14.5%, and with maturities around 13 years. Given the large CNG bus procurements (which involved significant cost premiums relative to their diesel counterparts), the high interest rates, and short maturity periods, DTC's liabilities have grown exponentially since CNG was introduced. The second key point is that labour expenses, of which those related to administration and bus operations form the lion's share -- and have been increasing continuously, and particularly rapidly since around 2007-08 -- account for the bulk of operating expenditures less taxes, interest, depreciation and other miscellaneous expenses. An important factor in this regard is labour productivity. In 2010-11, the average DTC depot, with 830 personnel, serviced 110-130 buses, for an average of 7.5 personnel-per-bus, as against an average of 6.2 personnel-per-bus in the other large public bus transit operations in Bangalore, Chennai, and Mumbai, and 2.5 personnel-per-bus in the USA (CIRT, 2012; Lowell et al., 2007; Clark et al., 2009).

Thirdly, note that total fuel expenses, being a function of fleet fuel economy, fuel price and total fleet activity, are largely driven by the variations in the last factor over the analysis period. On the other hand, fuel costs per kilometre, which are a function of fleet fuel economy and fuel price, appear to have been driven largely by the increase in diesel price for the all-diesel fleet until CNG introduction in 2000-01 (Figure 4.4a), given that fleet fuel economy varied only slightly, from 3.72 to 3.8 km/L between 1989-90 and 2000-01 (Figure 4.3a). Note that, even though capacity-kilometres and passengerkilometres were dropping significantly, load factors were as high as 87% on average, during this period. Fuel costs per kilometre started to increase rapidly with CNG introduction; indeed, they were 70% higher for the CNG fleet from 2003-04 to 2010-11, that is, after the transition to CNG had been completed, than for the all-diesel fleet from 1989-90 to 1998-99. This was despite the high load factors and the increase in diesel prices up to the transition period, on the one hand, and the significant reduction -- by as much as a third – in CNG prices, from the time of introduction of this fuel, in 2000-01, until around 2008-09 (Figure 4.5b), besides a much lower (69%) average load factor for the CNG buses than for the all-diesel fleet (Figure 4.1b). A key contributory factor was of course the much poorer fuel economy of CNG buses relative to diesel buses -- a situation that was exacerbated by the increase in CNG price from around 2009, just as the LF CNG buses, with even poorer fuel economy than the standard CNG buses, were being introduced.

A note on CNG prices: in the US context, natural gas feedstock, as represented by the average city-gate price (EIA, 2013), represented only about a quarter of the final retail price of CNG sold at refueling stations (EERE, 2013). Other major costs include those related to capital infrastructure for refueling stations, energy for compression, fuel retailers' margins, and taxes. All these factors are assumed to be priced into the CNG price paid by DTC to IGL. The average energy equivalent price of CNG hovered just under 50% of the diesel price from 2001 to 2011 (Figure 4.5a). Figure 4.5b compares the CNG price relative to that of diesel in the US and Indian contexts. It appears that, although the CNG-to-diesel price ratio has been higher in the USA than in India, this ratio has converged to similar levels towards the end of 2011.

The expenditures on workshop and maintenance labour plus all materials except fuel, per effective kilometre, show large variations over our analysis period (Figure 4.4b). Interestingly, the graph for this measure appears to be a near mirror image of the graph for total fleet activity (in terms of capacity-kilometres and passenger-kilometres) in Figure 4.1b. This is because workshop and maintenance labour and material expenses – and particularly workshop and maintenance labour, which forms the bulk of these expenses -- are fairly constant regardless of the fleet activity (or even the fuel system), as evidenced by Figure 4.4a. Even so, the average maintenance expenditure per

kilometre for the CNG fleet, from 2003-04 to 2010-11, was roughly 16% higher than for the all-diesel fleet, from 1989-90 to 1998-99. This is despite the fact that the average age of the latter fleet was 6.1 years (with a minimum of 5.6 years), and was almost 8 years before conversion to CNG (Figure 4.1a), whereas the average fleet age decreased to less than 2 years post-conversion, and gradually increased over the years until the procurement of the LF CNG buses.<sup>23</sup> Note also the significantly higher breakdown rates for the CNG buses, despite the effective kilometres per bus on road being lower on average for these buses, and having declined significantly from 2005-06, which we discussed earlier (Figure 4.3a). A final trend worth noting in Figure 4.4b is the decline in the per-kilometre expenditures on workshop and maintenance labour plus all materials except fuel from their peak in 2008-09. This was likely possible, despite the higher breakdown rates for the LF CNG buses, mostly due to the fixed nature of the LF bus maintenance costs under the service contracts, thus highlighting an effective cost management strategy used by DTC to transfer risk on the new bus technology.



Figure 4.6: Maintenance costs, key categories, per kilometre

At the same time, though, fleet age likely contributed to maintenance costs. The period from 2002 to 2009 may be used to understand the evolution of these costs for standard

<sup>&</sup>lt;sup>23</sup> After the fleet had been fully converted to CNG, in 2003-04, standard CNG buses made up over 90% of the operational fleet, with only a marginal presence of standard diesel buses that operated on inter-state routes.

CNG buses, as they aged over their 8-year service life. Analysis of the different components of these costs (Figure 4.6) shows that spare parts expenditures, which were initially low, rose to match or even exceed the labour costs in the last four years of this period.<sup>24</sup>



Figure 4.7: Operating costs per passenger kilometre and operating ratio

As we discussed earlier, passenger-kilometres have declined over the 20-year analysis period. Whatever the cause of this trend, the fact remains that it is happening even as CNG was introduced, and new standard, LF and LF-AC CNG buses were inducted into DTC's fleet from 2000-01 onward, involving significant capital expenditures (and cost premiums over their diesel counterparts), in the process considerably increasing DTC's debt burden. At the same time, operating costs have increased considerably (Figure 4.4a), even without considering taxes, interest, depreciation and other miscellaneous expenses, and even as passenger-kilometres have remained stagnant (Figure 4.1b).

<sup>&</sup>lt;sup>24</sup> Over 1,843 standard CNG buses were older that the targeted service life of 8 years in 2011 (CIRT 2012), and most likely contributed to the high maintenance costs per kilometre (Figure 4.4), as well as the higher breakdown rates in Figure 4.3a.

This effectively means that operating costs per passenger-kilometre have been increasing since CNG introduction; indeed, as shown in Figure 4.7, they have grown steadily over the 20-year analysis period. The net result of this, in turn, is that the operating ratio, which is the proportion of total traffic revenue accounted for by operating expenditures, and is thus an important measure of the operational efficiency of the transit service, has also been increasing steadily over the analysis period, and exceeded 250% during 2007-09 (again, it must be stressed that taxes, interest, depreciation and other miscellaneous expenses are excluded from Figure 4.7).

## 4.4 Conclusions and implications

As we argued in our introduction, the 10-plus years of experience accumulated as a result of the large scale conversion of DTC's bus fleet to CNG provides a valuable opportunity for a post-implementation evaluation of the operational and financial performance of this fuel system, thereby fulfilling an important research need. At the same time, this experience enables lessons to be drawn for the viability of similar large-scale conversions of urban bus fleets to CNG, and for informing techno-economic and environmental analyses of CNG bus transit operations, in India and other rapidly motorizing low and middle income countries. So, what lessons may be drawn from the DTC experience?

The fact that the major technical, logistical and institutional challenges associated with this implementation were overcome, and CNG was implemented on buses (and other public motor vehicles) in Delhi over the space of 3-4 years, despite these challenges, is undoubtedly an achievement, which demonstrates the importance of co-ordination between various actors, including vehicle manufacturers, fuel suppliers, bus and other public motor vehicle operators, and different levels of government.

At the same time, it is important to recognize that the implementation of the scrappage and CNG mandates, which were driven primarily by environmental concerns related to Delhi's air quality, caused serious reduction in the capacity to deliver transit services at DTC (as well as on the part of Delhi's private bus operators, and auto-rickshaws) in the initial stages of the transition to CNG, thereby likely causing significant hardships for commuters (reflected in the very high load factors during 1999-2001), and compromising easy and affordable access to livelihoods and other essential services. Ironically, this drastic capacity reduction also likely caused personal motor vehicle use to increase during this time. These unintended effects demonstrate the importance of anticipating and preparing for them prior to implementation, and also perhaps, of carrying out the implementation in a phased manner.

Above and beyond these effects is the fact that CNG implementation necessitated significant investments in buses at a considerable cost premium relative to their diesel counterparts, besides the investments in fuel infrastructure and depot modifications (which were thankfully not as substantial as in contexts such as New York's). Additionally, operating costs per kilometre grew considerably with the introduction of CNG, firstly due to significantly increased fuel expenditures per kilometre -- despite a low and declining CNG price – because of the lower fuel economy on the CNG buses, and secondly, increased maintenance costs (and breakdowns) per kilometre, despite declining effective kilometres per bus on road. Both of these costs were further exacerbated by the introduction of the LF and LF-AC CNG buses. Taken as a whole, these factors have adversely affected DTC's financial situation.

Further, despite enhancements to capacity-kilometres as a result of the significant investments in standard CNG buses during 2000-04, and then in the LF and LF-AC CNG buses -- which presumably were introduced to make transit more attractive and increase ridership – in 2007-10, passenger-kilometres generally declined over our 20-year analysis period. As a result, operating expenditures per passenger-kilometre, and the ratio of operating expenditures to traffic revenues, both important measures of operational efficiency of the transit service, have progressively worsened. This situation is only likely to be further exacerbated by CNG prices increasing to international levels.

The question that our analysis raises is, apart from its emissions outcomes, how has the significant investment in CNG contributed to, or taken away from, the objective of providing convenient, affordable and viable public transit service in Delhi, which is crucially important first and foremost to cater for the accessibility and mobility needs of the masses, and also to minimize the need for personal motor vehicle activity and its associated impacts, including air pollutant emissions. This question gains particular importance because of the decline in passenger-kilometres at DTC, and the equally significant reduction in public transit modal shares in Delhi among other metropolitan cities in India, over the past couple of decades.

It may be argued that the debt burden, and the overall financial situation due to the increased capital and operating costs per passenger-kilometre due to CNG implementation has in fact detracted from this objective, as well as the ability to enhance public transit capacity and provide widespread coverage region-wide. It is worth investigating whether, if the same investment had been made in emission-controlled diesel buses, a larger capacity to deliver bus transit services might have resulted, thereby better achieving equity, and conceivably, even environmental objectives, by helping avoid a larger number of personal motor vehicle trips region-wide. An additional problem in this regard is of course that the investments are being made without any accompanying TDM measures.

All of the foregoing points have important implications for policy-making and implementation, as well as for policy analysis. They demonstrate the need to analyse and formulate environmental policies such as the implementation of CNG on Delhi's buses broadly, in terms of a wide range of impacts for different groups in society, rather than narrowly in terms of only environmental (in this case emissions) outcomes, and to explicitly consider and address conflicts and trade-offs between environmental, and other (transit operation, socio-economic and equity) objectives.

# Chapter 5: CNG and diesel urban buses in India: A life-cycle cost comparison

#### Chapter overview

In Chapter 5, I estimate the life-cycle costs (LCCs) of diesel and natural gas-fueled public transit bus fleets in India. For this chapter, I draw from data and results in Chapter 4, coupled with on-road bus performance of other key large public bus transit operators in India in order to closely reflect fuel and bus technologies, operating conditions, and costs that are most prevalent in urban India, and build a LCC model for this context. A LCC approach is justified given the complexity of analyzing different fuel systems for buses that have many parameters, such as investment and operating cost requirements, which vary in kind, extent, and temporally over the service life of each fuel and bus technology; and since a LCC framework allows for the evaluation of all these costs using an integrated and comparable metric, the present value of total LCCs. Overall, the analysis of LCCs for CNG and diesel buses in India should provide lessons for assessing the long-term impacts of bus and fuel choices on the viability of CNG use in public bus transit fleets. Further, LCC results from this chapter will be used in Chapter 6 to evaluate the cost-effectiveness of the emissions reductions enabled by CNG public transit buses in Delhi.

#### 5.1 Introduction

The buses in the Indian capital, Delhi, including those operated by the publicly owned Delhi Transport Corporation (DTC), were converted to compressed natural gas (CNG) from 2000 to 2004 (Bell et al., 2004; Patankar & Patwardhan, 2006), pursuant to a Supreme Court of India (1998) ruling, and as part of a suite of technological and regulatory policies to address the poor air quality in this and other Indian cities (CSE, 2002b; TERI, 2002). CNG has also been implemented on buses in other Indian cities (CIRT, 2013).

The implementation of CNG on buses in Delhi and other Indian cities was motivated by the potential environmental and health benefits that CNG offers, by way of reduced emissions of critical air pollutants relative to the conventional diesel fuel that it replaces. CNG, when used in optimized engines, results in significantly reduced emissions of fine particulates, which are strongly correlated with health outcomes including respiratory and cardio-vascular illnesses and deaths, and of other air pollutants such as carbon monoxide, sulphur dioxide, and non-methane hydrocarbons (Kathuria, 2005; Narain & Krupnick, 2007). These benefits are likely to be enhanced in contexts like India, in which vehicle emission and fuel quality standards have been -- at least until recently -- less stringent than those elsewhere (Posada, 2009). As importantly, from the perspective of climate change, CNG has the potential to reduce net carbon dioxide equivalent emissions, relative to diesel in urban buses (Reynolds & Kandlikar, 2008).

# 5.1.1 CNG implementation on DTC's bus fleet

Delhi, with around 20,000 buses, has one of India's largest bus fleets; also, DTC has the second largest publicly operated urban bus fleet in the country, with around 4,700 CNG buses (CIRT, 2017). The first CNG buses in DTC's fleet, referred to in this paper as the CNG Standard model (CNG-Std. -- Table 5.1), replaced diesel buses of the same configuration from around 2000 to 2004. Starting in 2007-08, DTC introduced low-floor CNG buses (hereafter CNG-LF), to replace the CNG-Std. buses, 10% of which were over DTC's target service age of eight years in 2006-07 (CIRT, 2008). The CNG-LF buses, with a 400mm floor height (as opposed to 800mm for the standard buses), were introduced to offer improved accessibility. They also had full air suspension, automatic transmission, and a service life of 12 years (or 750,000 kilometers). Between 2007 and 2011, 3,700 CNG-LF buses were put into operation, of which 33% were air-conditioned (hereafter CNG-LF/AC), on which higher fares were charged. Of DTC's 4,300 operational buses in 2010-11, around 46% were standard CNG, 53% were CNG-LF (CIRT, 2012), and less than 1% were standard diesel buses.

The CNG-Std. buses had an engine power rating 31% higher than that of their diesel counterparts, and CNG stored in 8-12 cylinders below the floor; they cost 26% more than their diesel counterparts. The CNG-LF buses featured an integral body chassis, and rear-mounted engines with nearly double the power rating, and roof-mounted CNG storage. Also, these buses conformed to the Bharat Stage III and IV emission standards, which are based on the EURO-III and IV norms. The purchase cost of the CNG-LF and CNG-LF/AC buses is around USD 126,000 and 149,000 respectively, more than twice the USD 60,000 paid for a new CNG-Std. bus, as data from recent bus procurements in India show (Government of India, 2009a, 2009b, 2009c, 2011, 2012).

DTC's bus fleet is operated from over 40 depots, most of which have on-site CNG refueling infrastructure. This refueling infrastructure is operated by Indraprastha Gas Limited (IGL), which sells the CNG fuel to DTC at a negotiated price. Each depot typically serves only one type of bus (i.e., CNG-Std. or CNG-LF), and re-conditioning and overhauling are done at two central workshops.

From 2007-08, when the CNG-LF buses were introduced, DTC moved away from "inhouse" maintenance, and entered into a contract with the bus manufacturers to make them responsible for bus maintenance. The bus manufacturers conduct routine and preventive maintenance of the buses at the depots, and overhauls and other specialized maintenance at outside locations; they also bear the costs of labor and parts. This arrangement enables DTC to transfer technological risk to the manufacturers, and to reduce staffing levels and focus them on transit operations.

# 5.1.2 Rationale and objectives

Decision-making regarding CNG implementation on urban buses for mitigating urban air pollution needs to consider its environmental and health benefits, as well as its life-cycle (capital, operating and maintenance) costs, over the service life of the buses, relative to their conventional diesel counterparts. There is a considerable literature on the emissions effects of the implementation of CNG, among other measures, in the transport sector in Delhi, and in India more generally. Some studies worth mentioning are Kathuria (2005); Chelani and Devotta (2007); Narain and Krupnick (2007); Reynolds and Kandlikar (2008); and Reynolds, Grieshop, and Kandlikar (2011) in relation to the emissions effects of CNG implementation, and Jalihal and Reddy (2006); Kumar and Foster (2009); Goel and Guttikunda (2015); Aggarwal and Jain (2016); and Jain, Aggarwal, Sharma, and Kumar (2016), on the effects of a wider range of policy measures. However, very little if any attention has been paid to studying the operational or financial aspects of CNG implementation; indeed, there is a major gap in systematically understanding the costs of implementing CNG in the Indian context. The air pollution mitigation effects of CNG are of course important for society at large, especially since air quality improvement was the objective of the policy; however, the perspective of vehicle users and operators is also important. It would be useful to assess how bus fleets on CNG perform operationally and financially, because it is this performance that drives the bus operator's responses, which in turn affect the extent to which CNG implementation, and the resulting emissions reductions, are successful. Such an analysis is particularly useful from the perspective of private bus operators, but it is so for publicly-owned transit operators as well, given the challenge they face in providing quality and affordable bus services within budgetary constraints.

Further, it is useful to evaluate operational and financial performance on a life-cycle basis, because each fuel-bus technology system incurs a range of costs that vary in kind and extent, and temporally, over its service life. Varying patterns of purchase, operation, maintenance, servicing and replacement costs related to the fuel systems, buses, engines, and related supporting (including, importantly, fuelling) infrastructure need to be accounted for. A life-cycle cost (LCC) approach, based on the present value of bus fleet life-cycle costs, would integrate and compare these varying capital and operating cost streams for different fuel-bus technology systems on an uniform basis. This is particularly useful when evaluating CNG bus systems since such systems must be shown to be cost-effective in mitigating emissions, relative to diesel options with exhaust treatment, which do not require major changes to the refueling infrastructure.

The operational and financial performance of DTC's bus fleet was critically evaluated from 2001-02, shortly after CNG implementation began, until 2010-11, and compared with the performance of the fleet when it was diesel powered, from 1989-90 until when CNG was implemented, in Krelling and Badami (2016). That study evaluated how the operational and financial performance of the fleet was affected, as it was converted from diesel to CNG-Std., and then to CNG-LF and CNG-LF/AC buses. In this paper, we evaluate the LCCs incorporating the elements discussed in the previous paragraph related to these CNG bus configurations in Delhi, and compare them with those associated with their diesel counterparts in the Indian context. We base these evaluations on actual on-road bus performance data for DTC and other Indian public bus transit agencies, in order to closely reflect fuel and bus technologies, operating conditions, and costs that are prevalent in urban India. In order to see what might happen if fuel prices were largely market driven, we also assess the LCCs of low-floor air-conditioned CNG and diesel buses in India, but with US fuel prices. We then present sensitivity analyses of LCCs relative to fuel economy and fuel price variations, for the Standard and LF/AC configurations, to show how these key factors affect the viability of CNG bus systems relative to their diesel counterparts. Lastly, we analyze how the choice of a discount rate affects these assessments. No such evaluation has been conducted on CNG urban bus transit fleets in India or similar contexts, to our knowledge.<sup>25</sup>

Our in-depth evaluation of the LCCs of CNG urban buses, as outlined above, is intended to be an input into decision-making regarding the implementation of CNG buses for mitigating urban air pollution, in concert with an assessment of the environmental and health benefits of CNG in this application. We do not address the assessment of the environmental and health benefits of CNG in this application in this paper, because this assessment will involve, among other things, a comprehensive and detailed modeling of emissions of health-critical air pollutants and greenhouse gases for the various CNG bus configurations we consider, over their service life, relative to their

<sup>&</sup>lt;sup>25</sup> A study by TERI (2002) analyzed the "well to wheels" cost of producing and distributing CNG relative to diesel containing 50ppm sulphur, but did not address the life-cycle costs of CNG buses.

diesel counterparts, under real-life operating conditions, and would require a full length paper on its own. Our in-depth evaluation of the LCCs of CNG buses in this paper will hopefully fill an important policy research need, and help inform techno-economic and environmental analyses of CNG urban bus transit operations in India and similar contexts.

# 5.2 Analytic framework, methods, and data

# 5.2.1 Bus and fuel systems evaluated

We conduct our LCC analysis of CNG bus systems in the Indian context, using operational and financial performance data related to the CNG-Std., CNG-LF and CNG-LF/AC buses operated by DTC, over the period from 2001-02, very shortly after CNG buses began to be introduced in Delhi (and the country), to 2010-11 (Table 5.1). DTC's experience with CNG buses over more than ten years since around 2000 makes it a good representative case of large-scale CNG urban bus operations for our analysis in India. While BEST, the public bus operator in Mumbai has also operated CNG buses for nearly as long as DTC, a substantial proportion of its buses continue to be diesel-powered (CIRT, 2017). Most importantly from the perspective of this research, detailed, disaggregated and long-term data on the various parameters necessary for the LCC analysis are not readily available for the CNG bus fleet in Mumbai. Besides, the operating conditions (and therefore bus fleet performance) in Delhi and Mumbai are likely to be fairly similar -- as borne out by the fleet fuel economy of CNG buses being similar in these two cities (CIRT, 2012).

As for comparing the LCCs of the CNG-Std., CNG-LF and CNG-LF/AC buses with their diesel counterparts in the Indian context, we use operational and financial performance data related to the diesel bus fleets in Bangalore, Chennai and Mumbai, since DTC and other bus operators in Delhi have not operated diesel buses since 2003. Note that these

cities have large, well-established public bus transit systems<sup>26</sup>, as in Delhi. Besides, the operating conditions in these cities, in terms of, for example, average driving speeds and bus load factors, are very similar to those faced by DTC in Delhi (CIRT, 2012; WSA, 2008a, 2008b).<sup>27</sup> So, the CNG and diesel bus technologies being analyzed and compared in terms of their LCCs reflect those operated by the largest public transit fleets in the major metropolitan areas of India; and our analysis is based on actual -- and contemporaneous -- on-road CNG and diesel bus performance in these areas.

Table 5.1 summarizes the key characteristics of the CNG and diesel bus technologies evaluated in our LCC analysis.

Table of the Guilling of Key bac eyetenne evaluated in mana								
	CNG Std.	ĈNG LF	CNG LF/AC	Diesel Std.	Diesel LF	Diesel LF/AC		
Model Year (MY)	2001/02	2007/08	2007/08	Various	Various	2005+		
Fuel Type	CNG	CNG	CNG	Diesel	Diesel	Diesel		
Manufacturer / Model	Tata, AL	Tata LE RE LPO 1623	Tata	Tata, AL	Tata, AL	Tata, AL		
Engine (max HP)	120-150	230	250	125-165	180	225-290		
Compression ratio	10.5:1	10.5:1	10.5:1	~17:1	17.2:1	17.3:1		
No. of cylinders	6	6	6	6	6	6		
Engine displacement	5.9L	5.9L	5.9L	5.9L	5.9L	5.9-7L		
Emissions control/Standards	Catalytic converter	BS-III	BS-III	BS-III	BS-III	BS-III		
Fuel Capacity	90-96 kg	108 kg	108 kg	165 L	150 L	150-310 L		
Gross Vehicle Weight	15.3 <sup>°</sup>	16.2	16.2	~15	16.2	16.0		
Length (meters)	~10.7	12	12	~10	12	12		
Passenger capacity	68	70	70	66	70	70		

Table 5.1: Summary of key bus systems evaluated in India

Source: CNG bus data from DTC, Delhi; diesel buses reflect the technologies used at BEST (Mumbai), BMTC (Bangalore) and MTC-CNI (Chennai); data for these operators are based on bus specification guidelines established by the Ministry of Urban Development of the Government of India for the purchase of buses using public funding (MoUD 2012; 2013).

<sup>&</sup>lt;sup>26</sup> The public bus transit counterparts of DTC in Bangalore, Chennai and Mumbai are the Bangalore Metropolitan Transport Corporation (BMTC); Metropolitan Transport Corporation Limited, Chennai (MTC-CNI); and Brihan Mumbai Electric Supply & Transport Undertaking (BEST). Their fleets comprised 3000 or more buses over the period from 2001-02 to 2010-11 (CIRT, 2003-2012).

<sup>&</sup>lt;sup>27</sup> The average speeds were very similar -- 16 km/h, 16 km/h, 19 km/h, and 18 km/h respectively -- in Delhi, Mumbai, Chennai and Bangalore (WSA, 2008a, p. 28); and all these cities have a "plain" terrain (WSA, 2008b, pp. 100, 103, 106, 112). The average age of the bus fleet was 5 years, 7 years, 6 years, and 4 years respectively in these cities, over our evaluation period from 2001-02 to 2010-11 (CIRT, 2003 to 2012). The load factors in these cities were 71%, 63%, 81%, and 64% respectively. Thus, the average fleet age and load factor in Delhi were close to the average of the fleet ages (5.7 years) and the average of the load factors (69%) in the other three cities.

#### 5.2.2 Analytic framework

While our research draws on Schubert and Fable (2005), Lowell, Chernicoff, & Lian (2007), Laver et al. (2007), Clark et al. (2009), and Johnson (2010), among other studies, the model we used to calculate LCCs (Equation (1) below) was based on Lowell et al. (2007) and Clark et al. (2009), given their comprehensive consideration of various parameters. However, it is worth reiterating, as noted earlier, that our evaluations use actual on-road bus performance data for DTC and other Indian public bus transit agencies, and closely reflect Indian fuel and bus technologies, operating conditions, and costs. Each parameter in Equation (1) is explained in Table 5.2.

$$LCC_{i}(y, w_{i}, d, e_{i}, k, r) = w_{0,i}d + w_{1,i}\frac{y}{k} + w_{5}e_{0,i}d + \sum_{t=1}^{T}\frac{(w_{2,i,t} + w_{3,i} + w_{4})y + w_{5}e_{1,i}d}{(1+r)^{t}}$$
(1)

#### i) Operational and financial parameters, and data sources

Table 5.3 summarizes the key operational and financial parameters -- including the infrastructure costs ( $w_0$ ), Bus service life (T), Bus purchase cost ( $w_1$ ), Fuel costs ( $w_3$ ), and labor costs ( $w_4$ ) -- that were used in the LCC model, as well as the data sources and the periods over which the data were drawn. Following this table is a discussion of the key parameters, as well as the specific methods used in their estimation. An indepth description of all assumptions used in the LCC estimation is presented in the Appendix B; this description includes the specific values used for various parameters (e.g., training costs, with hours used,  $e_0$  and  $e_1$ , and hourly wages,  $w_5$ ).

Parameter	Description
LCC <sub>i</sub>	Present value of the total capital and operational costs during infrastructure/fleet setup
	period and the expected service life of the bus fleet, <i>T</i> , in US dollars (\$)
i	Refers to different types of bus fleets operating on diesel and CNG fuel systems shown in Table 5.1
v	Annual fleet vehicle-kilometres (y) = [number of buses in fleet] × k (kilometres per bus)
Wi	Unit costs related to each <i>i</i> , detailed below.
$w_0$	Depot infrastructure costs, measured in \$-per-depot; total depot cost = $w_0 x$ the number of depots, <i>d</i> , required to service the fleet.
WI	Bus purchase cost, measured in \$-per-bus purchased; total bus purchase cost = $[w_l] x$ [the number of buses purchased].
W2	Maintenance cost, measured in \$-per-kilometre; total fleet maintenance cost = $[w_2] x$ [annual fleet vehicle-kilometres (y)].
W3	Fuel cost, measured in \$-per-kilometre:
	$r_{i} = \frac{P_i}{P_i}$
	$FE_i$
	$FE_i$ – bus fuel economy, in kilometres-per-litre (or litre diesel energy equivalent for $CNC$ ).
	Total fleet fuel cost = $[w_i]$ x [annual fleet vehicle-kilometres (v)]
	CNG fuel consumption was converted to litres of diesel on an energy equivalent basis.
	based on Faiz et al. (1996) and CPCB (2010).
$W_4$	Operator labour costs, measured in \$-per-kilometre for various bus fleet types.
	Operator labour represents all labour, other than those involved in maintenance of
	buses (e.g., workshop and maintenance personnel), and includes drivers, conductors,
	traffic supervisors, administrative personnel and others (see personnel costs in CIR I
	Total fleet operator labour costs = $[w_4] \times [annual fleet vehicle-kilometres (v)].$
	Average training costs measured in $^{\circ}$ per beur per employee: broken down by steff
W5	function (administration, maintenance or operator), but salary is assumed to be
	common across bus types. Total fleet training costs = $[w_s]$ x [the number of employee-
	hours required for training]; see e, below.
d	The number of depots required to service the bus fleet, which is important in
	determining depot infrastructure and staffing requirements to effectively service the
	average bus capacity of each depot (integer rounded up)
<i>e</i> <sub>i</sub>	Total fleet employee training hours, broken down by specific employee function (this is
-	treated as a vector, accounting for various specific employee functions); it is calculated
	by multiplying the total training hours (for each function) by the staffing numbers (for
	each function), for individual depots operating any given bus fleet type <i>i</i> .
<b>e</b> 0	Total initial training hours vector, per employee category; also see $e_i$
<b>e</b> 1	Total annual ongoing training hours vector, per employee category; also see $e_i$
t T	Refers to a specific year of operation within <i>i</i> 's expected service life, <i>T</i>
1 k	Appual kilometres per bus: this measure reflects the average revenue corning
n	kilometres operated by each bus in the fleet
r	Discount rate; annual percentage.

# Table 5.2: LCC model parameters in Equation (1)

i	CNG	CNG	CNG	Diesel	Diesel	Diesel
	Std.	LF	LF/AC	Std.	LF	LF/AC
Parameters:						
Fuel type	CNG	CNG	CNG	Diesel	Diesel	Diesel
Depot cost (w <sub>0</sub> ), USDx10 <sup>6</sup>	2.2	2.2	2.2	1.7	1.7	1.7
Service life, years	25	25	25	25	25	25
Annual operation & maint., % of $w_0$	5.5%	5.5%	5.5%	5.5%	5.5%	5.5%
Bus service life (T), years	10	12	12	10	12	12
Bus purchase cost $(w_1)$ , $bus$	59,623	125,744	149,347	48,536	104,661	139,882
Residual value, % of w <sub>1</sub>	13.4%	13.4%	13.4%	14.9%	14.9%	14.9%
Avg. maintenance costs, \$/km <sup>(1)</sup>	0.157	0.185	0.198	0.136	0.156	0.163
Maint. evaluation period	2002-11	AMC (3)	AMC (3)	2002-11	Letimeted (4)	
Maint. source data	CIRT	DTC	DTC	CIRT	<<< Esum	
Fuel costs (w <sub>3</sub> ), \$/km	0.186	0.237	0.296	0.222	0.318	0.397
Fuel Economy (FE), km/L <sup>(2)</sup>	2.12	1.66	1.33	3.81	2.67	2.14
FE evaluation period	2002-11	2009-10	2009-10	2002-11	Est. (5)	2006-11 <sup>(5)</sup>
FE source data	CIRT	DTC	DTC	CIRT		CIRT
Fuel price (P), \$/L	0.480	0.480	0.480	0.860	0.860	0.860
Fuel price evaluation period	2011-15	2011-15	2011-15	2011-15	2011-15	2011-15
Fuel price source data	<<	MyPetrolPrice.com (2016)				>>
Operator labour costs (w <sub>4</sub> ), \$/km	0.442	0.442	0.442	0.442	0.442	0.442

**Source**: CIRT (2012) & DTC: CNG, DTC (Delhi); India diesel: average performance of BEST (Mumbai), MTC-CNI (Chennai), BMTC (Bangalore) (1) Average maintenance expenses reflect the average costs throughout T, for each bus technology and includes costs with workshop and maintenance labour, plus materials (lubricants, spares, tyres, batteries, etc.); actual cost profile used in LCC model is shown and discussed below (see Figure 5.1).

(2) Fuel Economy for CNG calculated on a litre diesel energy equivalent basis.

(3) AMC: Annual Maintenance Contract between bus manufacturers and DTC, for maintenance costs for systems overhaul as well as workshop labour and consumables throughout the 12-year service life of buses.

(4) Maintenance cost data not available for LF and LF/AC diesel buses; estimation methods discussed in the Appendix B.

(5) Fuel economy for LF diesel buses: estimation methods discussed in the Appendix B.

Fuel economy for LF/AC diesel buses: data is a representative value based on Volvo buses at BMTC, MTC-CNI, and BEST.

We used various data sources for the in-use fleet operational and financial performance parameters of the LCC model to evaluate CNG and diesel buses in India. The CNG bus performance data was based on statistics gathered by the first author at DTC and on annual reports produced by the Central Institute of Road Transport (CIRT, 2002–2012). For diesel bus performance in the Indian cities of Mumbai, Chennai and Bangalore, data was also drawn from the Central Institute of Road Transport annual reports (CIRT, 2002–2012) for BEST, MTC-CNI, and BMTC, respectively, during the same period as for DTC.

The LCCs in Equation (1) represent the total costs of the bus fleet over the service life, T, of each bus system (*i*). In order to compare the various bus systems on an uniform basis, and because of the variable size of the fleets representing these systems in our study, we calculated the LCCs for 100 buses for all of the bus systems, keeping in mind that this is the approximate number of buses serviced in each depot at DTC; incidentally, 100-bus depots are typical of US transit operations (Lowell et al., 2007). This assumption also means that the number of depots (d) in our LCC calculations was

1. Further, the LCCs were estimated for a 12-year service life, for both diesel and CNG bus systems. Since the Standard buses had a service life of 10 years (Table 5.2), the capital cost of these buses was converted into an annualized figure (after accounting for the discount rate), and then pro-rated over a 12-year period. Operational costs were calculated based on costs per kilometre ( $w_2$ ,  $w_3$ , and  $w_4$ ) and on annual fleet vehicle-kilometers (y), which in turn were based on the average annual kilometers operated per bus (k). (k) was assumed to be 70,000 kilometers, based on the average distances operated by buses in the Indian cities (CIRT, 2012). Finally, note that we present the LCC results for all the bus systems, as shown in Equation (2) below, in US dollars-per-kilometre<sup>28</sup>.

Average Life Cycle Costs for each bus system = 
$$\frac{LCC_i}{y \cdot T}$$
 (2)

**Depot costs** Depot infrastructure needs for CNG bus fleets include those related to the refueling infrastructure and changes to vehicle storage and maintenance facilities. Data collected at DTC showed that, in the Indian context, a bus depot with the capacity for servicing 100 CNG buses cost approximately USD 2 million; more recently built depots cost DTC an average of USD 2.2 million. These costs include those related to building infrastructure (storage areas, administration, workshop, etc.) required for bus operation, but exclude costs related to the CNG refueling infrastructure, which the majority of DTC's depots have on site.

The refuelling infrastructure, which includes CNG compressors, buffer storage tanks, and fast-fill dispensers, was put in and is operated by Indraprastha Gas Limited (IGL), the local natural gas utility, which sells CNG to DTC at a negotiated price, as noted earlier. This price factors in the capital costs of the refuelling infrastructure, as well as the natural gas feedstock costs, and the costs of operating and maintaining the CNG refuelling infrastructure, incurred by IGL.

<sup>&</sup>lt;sup>28</sup> We use US Dollars (USD), rather than Indian Rupees (INR), for reporting costs in order to facilitate comparison of our findings to other contexts.

Therefore, in order to avoid double counting of costs, we did not add the fuel refuelling infrastructure costs to *w0*, the depot infrastructure costs in our model, but rather accounted for this factor, and the other costs associated with the refuelling infrastructure discussed above, by using the fuel prices faced by DTC. Comparable depot costs in India for exclusively diesel-based bus operators range from USD 1.3 to 2.1 million in 2009 (Government of India, 2009a, 2009b). So, the depot infrastructure cost in the model (*w*<sub>0</sub>) was set to USD 2.2 million for CNG and USD 1.7 million for diesel. Depot infrastructure was assumed to have a 25-year service life in all cases, based on the literature. Since the LCC estimates were made over an uniform 12-year period for all the bus systems, as already discussed, the up-front depot cost was converted into an annualized figure (after accounting for the discount rate), and then pro-rated over a 12-year period.

**Bus purchase costs and residual value** Data for Indian transit operators that purchased both CNG and diesel buses (Government of India, 2009a, 2009b, 2009c, 2011, 2012) shows that, considering all bus purchases on the whole, the average cost premium for CNG buses was 7-25%, with higher premiums for the less expensive standard CNG buses, and lower premiums for the more costly low-floor CNG buses; incidentally, these figures are similar to the 8-20% premiums for CNG buses in Europe and the USA (Lowell et al., 2007; Clark et. al., 2009; Posada, 2009). For our LCC calculations, CNG-Std., CNG-LF and CNG-LF/AC bus costs (Table 5.3) were based on data collected on recent purchases at DTC, and inflation-adjusted. In order to ensure the best possible comparative purchase costs for their diesel counterparts, we used data for other large urban bus operators in India for the Std. buses, and for the Municipal Corporation Ludhiana (2013) for the LF and LF/AC buses, because their diesel buses are of the same type and model as the CNG buses at DTC.

At the end of the fleet's useful service life, bus costs,  $w_1$ , are subtracted by a residual value (see Appendix B for details), based on the fact that buses typically operate beyond, and thus have (market) value at the end of, their service life, *T*. For example,

Laver et al. (2007) show that buses in the US operate 15 years on average, even though the target service life is 12 years. The residual value is treated in the LCC model as a negative cost (i.e., a revenue to the operator) received at year T+1.

**Fuel prices** The average CNG and diesel prices for the period 2011-15 reported in Delhi (MyPetrolPrice.com, 2016) were used in the LCC estimations for all the cases, since Delhi has an important and sizable market for both fuels, and in order to evaluate all of the cases on an uniform basis in terms of fuel price (note that there have been large fluctuations in fuel prices over the last few years – Figure 5.3).

**Maintenance costs** The maintenance expenditures reported in Table 5.3 include costs associated with both labor and parts, and are the average of the actual maintenance expenditure ( $w_{2,i,t}$ ) incurred throughout the entire service life (*T*) of each bus system *i*. These expenditures take into account the fact that maintenance costs increase from year to year (*t*) over the service life, as the buses age, and also the need for vehicle and engine overhauls and rebuilds over time. Figure 5.1 shows the maintenance cost assumptions,  $w_{2,i,t}$ , used in the LCC estimations for each bus system, during each year of operation, *t* over its service life; a detailed description of how Figure 5.1 was generated, along with data sources used, is presented in the Appendix B.



Figure 5.1: Annual maintenance costs

**Financial parameters and the discount rate** Financial data originally reported in Indian Rupees (INR) were, unless noted, adjusted for inflation based on the Consumer Price Index for Industrial Workers (CPI-IW) drawn from Government of India (2016), with 2010-11 as the base year. These inflation-adjusted values were converted to US dollars (USD), using the average exchange rate published by the Reserve Bank of India (2016) for the same base year. The cost projections over the service life *T* were not adjusted for inflation, since inflation impacts all financial measures uniformly and would also have to be included in the discount rate. Thus, the discount rate used is a real discount rate (that is, nominal rate minus inflation expectations). A discount rate of 12% was used, based on Zhuang et al. (2007), which contains a theoretical discussion of social discount rates, but also a review of actual rates used and recommended by multilateral agencies globally; this rate is on par with that recommended by the Asian Development Bank for India, according to this reference. We also show the LCCs at 6% and 0% discount rates, for the various cases.

# ii) Sensitivity analysis

Further to our LCC calculations, we conducted simulations to assess the sensitivity of the LCC results to hypothetical changes in fuel economy and fuel price, both of which are important parameters that influence fuel costs, which are a major operating expense for transit agencies. While bus purchase costs are significant and sometimes of the same magnitude as life-cycle fuel costs, bus purchase costs have not changed significantly over the years for CNG relative to diesel buses (Government of India, 2009a, 2009b, 2009c, 2011, 2012). On the other hand, fuel prices have varied considerably over time (Krelling & Badami, 2016; MyPetrolPrice.com, 2016). Also, fuel economy varies depending on bus type and operating conditions, as discussed below. For these reasons, two sets of simulations were conducted, to estimate the effect of, first of all, varying ratios of CNG bus fuel economy relative to diesel bus fuel economy ( $\rho = FE_{CNG bus}/FE_{Diesel bus}$ ), and secondly, varying ratios of CNG price relative to diesel price on an energy equivalent basis ( $\phi = P_{CNG}/P_{Diesel}$ ), on the LCC for CNG relative to

diesel buses ( $D = LCC_{Diesel}/LCC_{CNG}$ ); thus, when D > 1, CNG would be a better alternative for the transit agency, from a LCC perspective. Each of these two sets of simulations was carried out for Standard buses and LF/AC buses.

ρ, the ratio of CNG-to-diesel bus fuel economy, is smaller than 1 in most cases. In the case of DTC, the average fuel economy (in km/L diesel energy equivalent) of standard CNG buses was 45% lower than that of their diesel-only fleet<sup>29</sup>; in other words, ρ = 0.55. Further, ρ is about 0.62 for low-floor buses in India, based on the data for low-floor CNG buses at DTC and for comparable low-floor diesel buses in other Indian cities. However, empirical evidence on CNG and diesel buses operating under similar conditions in the USA shows ρ ranging from 0.56 to 0.92 (see Chandler, Eberts, & Melendez, 2006; Barnitt, 2008; and Clark et al., 2009). Further, in one of the cases considered in the USA (Melendez et al., 2005), CNG bus fuel economy was superior to that of diesel buses (ρ>1). In view of the foregoing, we conducted the fuel economy related simulations using a wide range of possibilities, with ρ ranging from 0.2 to 1.2.

With regard to the fuel price related simulations, note that in India, the price of CNG relative to that of diesel on an energy equivalent basis ranged from a maximum of 0.61 to a minimum of 0.40, between 2001-02 and 2008-09, after which it re-bounded to 0.60 in June-July 2012; in the USA, where fuel prices are more market-driven, this ratio ranged from a maximum of 1.00 in 2001-02 to a minimum of 0.60 during 2011-2014 (Figure 5.3; MyPetrolPrice.com, 2016; EERE, 2016). More recently in India, the average CNG-to-diesel fuel price ratio was 0.56 during 2011-2015 (MyPetrolPrice.com, 2016). In view of the foregoing, we conducted the fuel price related simulations with  $\varphi$  ranging from 0.2 to 1.2, as for the fuel economy simulations. Finally, we investigate how our sensitivity analysis is affected by the choice of discount rate.

<sup>&</sup>lt;sup>29</sup> At DTC, the average standard diesel-only fleet fuel economy was approximately 3.77 km/L; the standard CNG bus fuel economy was approximately 2.09 km/litre diesel energy equivalent.

#### 5.3 Results and discussion

Figure 5.2 and Table 5.4 show the total LCCs calculated with a break-down in terms of various cost categories, in \$/km, for the CNG-Std., CNG-LF and CNG-LF/AC buses and their diesel counterparts in the Indian context; and for CNG-LF/AC and its diesel counterpart, but with CNG and diesel prices as in the USA, in order to assess how the LCC for CNG relative to diesel might be affected if fuel prices were largely market-driven. Note that, in addition to these estimates at 12% discount rate, we also indicate the total LCCs at 6 and 0% discount rates, as well as the total LCCs on a \$/bus basis at 12, 6 and 0% discount rates, for each of these cases. Further, Table 5.5 compares CNG relative to diesel fuel systems, in terms of various key cost categories, and total LCCs, for all of these cases.

					India with US fuel prices*			
	India							
	CNG	CNG	CNG	Diesel	Diesel	Diesel	CNG	Diesel
	Std.	LF	LF/AC	Std.	LF	LF/AC	LF/AC	LF/AC
Depot infrastructure (including operation and maintenance costs)	0.028	0.028	0.028	0.022	0.022	0.022	0.028	0.022
Bopor initial dotato (including operation and maintenance occis)		4.4%	3.9%	4.4%	3.6%	3.1%	3.7%	3.1%
Due numbers	0.076	0.145	0.172	0.061	0.120	0.161	0.172	0.161
Bus purchase	14.3%	22.7%	24.3%	12.3%	19.6%	23.0%	22.9%	22.5%
Due maintenance cost (labour and name)	0.079	0.088	0.095	0.069	0.076	0.079	0.095	0.079
Bus maintenance cost (labour and parts)	15.0%	13.7%	13.4%	13.8%	12.4%	11.4%	12.7%	11.1%
Fuel cost	0.117	0.149	0.186	0.116	0.167	0.208	0.227	0.223
Fuercost	22.1%	23.3%	26.2%	23.4%	27.1%	29.7%	30.2%	31.3%
Labour (bus operators & administrative) & training (all		0.229	0.229	0.229	0.229	0.229	0.229	0.229
personnel) costs	43.3%	35.9%	32.2%	46.0%	37.3%	32.7%	30.5%	32.0%
Total life-cycle costs (LCC) for discount rate at 12%	0.528	0.639	0.710	0.497	0.613	0.699	0.751	0.714
Total life-cycle costs (LCC) for discount rate at 6% Total life-cycle costs (LCC) for discount rate at 0%		0.801	0.887	0.643	0.776	0.876	0.942	0.897
		1.067	1.176	0.881	1.044	1.167	1.255	1.197
Total LCC (1,000\$/bus) – discount rate at 12% Total LCC (1,000\$/bus) – discount rate at 6% Total LCC (1,000\$/bus) – discount rate at 0%		536	597	418	515	587	631	600
		672	745	540	652	736	791	754
		896	988	740	877	980	1,054	1,005

#### Table 5.4: Comparison of bus life-cycle costs (\$/km)

Note: \*India with US fuel prices: Fuel prices assumed to be as in the US, with all else as in the Indian context. Percentage figures show proportion of each cost component as a percentage share of total LCCs for each bus configuration, at 12% discount rate.

# 5.3.1 LCCs for CNG-Std. versus CNG-LF versus CNG-LF/AC in the Indian context

The total LCC of the CNG-LF/AC buses is 11% and 34% higher than that of the CNG-LF and CNG-Std. buses, on a per-kilometre basis (and per-bus basis), at a discount

rate of 12% (Table 5.4). Although labor and training costs are the most important component of overall LCC for all the CNG bus configurations, they are assumed to not vary across these configurations; hence, the above LCC differentials are driven, in decreasing order of importance, by the higher per-kilometre bus purchase, fuel and maintenance cost, for the CNG-LF/AC relative to the CNG-LF and CNG-Std. buses. Fuel costs per-kilometre for the CNG-LF/AC buses are 25% and 59% higher, and maintenance costs 8% and 20% higher relative to those for the CNG-LF and CNG-Std. buses (Table 5.4); the higher fuel costs reflect the significantly lower fuel economy of the low-floor buses (Table 5.3). The most important components of overall LCC, after labor and training, are fuel, bus purchase, maintenance and depot infrastructure costs, for all three CNG bus configurations, except that maintenance costs marginally exceed bus purchase costs for the CNG-Std. buses. The fuel and bus purchase costs account for a progressively higher share of total LCC, when moving from CNG-Std. to CNG-LF to CNG-LF/AC buses, again reflecting the significantly higher purchase costs and poorer fuel economy of the low-floor buses. Correspondingly, bus maintenance and labor costs account for a smaller share of total LCCs for the LF and LF/AC buses relative to the CNG-Std. buses, even though, in absolute terms, the bus maintenance costs are higher, and the labor and training costs are the same, for the CNG low-floor buses relative to the CNG-Std. buses.

		India			
Percentage difference of LCC of:	CNG Std.	CNG LF	CNG LF/AC	CNG LF/AC	
over LCC of:	vs. Diesel Std.	vs. Diesel LF	vs. Diesel LF/AC	vs. Diesel LF/AC	
Depot infrastructure (including operation and maintenance costs)	27.4%	27.4%	27.4%	27.4%	
Bus purchase	23.2%	20.6%	7.1%	7.1%	
Bus maintenance cost (labour and parts)	15.0%	15.6%	19.7%	19.7%	
Fuel cost	0.1%	-10.6%	-10.6%	1.4%	
Labour (bus operators & administrative) & training (all personnel) costs	0.1%	0.1%	0.1%	0.1%	
Total life-cycle costs (LCC) for discount rate at 12%	6.2%	4.1%	1.6%	5.1%	
Total life-cycle costs (LCC) for discount rate at 6%	5.5%	3.1%	1.2%	5.0%	
Total life-cvcle costs (LCC) for discount rate at 0%	4.7%	2.2%	0.8%	4.9%	

#### Table 5.5: Differential life-cycle costs, by bus type and cost category

**Note:** Figures here rounded based on Table 5.4 data, which reflects more precision than the three decimal digits displayed. See note in Table 5.4 and text for explanations of the India with US fuel prices case.



*Note*: This figure graphically represents the calculations shown in Table 5.4, at 12% discount rate. **Figure 5.2: Comparison of bus life-cycle costs** 

# 5.3.2 LCCs for CNG versus diesel buses for Std., LF and LF/AC bus configurations in the Indian context

Whereas the per-kilometre LCC of the CNG buses was highest for the LF/AC, and the lowest for the Std. bus configuration, as discussed, and the LCCs for diesel buses were lower than for CNG for the Std., LF and LF/AC configurations, the LCC differential for CNG relative to diesel was the lowest (1.6%) for the LF/AC configuration and the highest (6.2%) for the Std. configuration, at 12% discount rate (Table 5.5). It therefore appears that the implementation of CNG negatively affects the LCC of Std. bus fleets proportionately more than for low-floor bus fleets, for which the LCC is already high.

This result was driven by the differential bus purchase cost of CNG relative to diesel being the highest (23.2%) for the Std. buses, and the lowest (7.1%) for the LF/AC buses. More importantly, this result was due to fuel cost, which accounts for the largest

share of total LCC after labor and training, being 10.6% lower for CNG relative to diesel for the LF and LF/AC configurations, and only 0.1% higher relative to diesel for the Std. buses (Table 5.5). This effect, which is surprising, given that the fuel economy of CNG buses is significantly lower than for diesel for all three configurations on an energy equivalent basis, is mainly due to the energy equivalent CNG fuel price being only 56% of the price of diesel, on average, over the most recent five-year period (2011-2015) considered in our analysis<sup>30</sup>, and the fuel economy for CNG being 44% lower on the Std. buses relative to diesel, while it was 38% lower on the low-floor buses (Table 5.3).

A key factor driving the total LCC differential for CNG relative to diesel being the lowest (1.6%) for the LF/AC configuration and the highest (6.2%) for the Std. configuration, is the fact that the differential bus purchase and fuel costs of the CNG buses relative to diesel more than compensated for the differential maintenance costs for CNG relative to diesel being the highest for the LF/AC buses, and the lowest for the Std. buses. Also, note that maintenance costs account for a lower share of total LCC than do fuel cost and even bus purchase cost, for the largest share of overall LCCs across the board) were assumed to be the same for CNG and diesel, and so did not affect the LCC differential for CNG relative to diesel for the three bus configurations. And while the depot infrastructure cost was considerably higher (27.4%) for CNG relative to diesel (Table 5.5), this cost category accounts only for a small share of the overall LCC (Table 5.4), and so did not substantially affect the LCC differential for CNG relative to diesel for the three bus configurations.

# 5.3.3 LCCs of CNG and diesel LF/AC buses in India, with US fuel prices

Interestingly in this case, the total LCC for CNG LF/AC is as much as 5.1% higher than for diesel, just due to using US fuel prices, as opposed to only 1.6% higher, with Indian

<sup>&</sup>lt;sup>30</sup> Actually, the CNG-to-diesel price ratio was lower, so the per-kilometre fuel costs would have been even more favourable for CNG relative to diesel than in the present case, in the previous five years (MyPetrolPrice.com, 2016); this shows how sensitive the LCC results are to fuel price (and fuel economy) variations, which is the reason for our sensitivity analysis related to these factors (reported in Section 5.3.4).

fuel prices (Tables 5.4 and 5.5, and Figure 5.2). This is because, while the fuel costs, and their share of total LCC, go up because of the fuel prices being generally higher in the USA than in India during 2011-15, which is the period over which the fuel prices were used in the LCC estimations for all the cases we evaluated, the fuel costs increase more for CNG relative to diesel, because of the higher CNG-to-diesel fuel price ratio in the USA relative to India (Figure 5.3). Note that similar increases in total LCCs occur for CNG relative to diesel for the Standard and low-floor bus configurations, when US instead of Indian fuel prices are used.

Lastly, a note on the effect of using different discount rates -- as progressively lower discount rates (6% and 0%) are used, instead of 12%, the recurring costs (related to fuel, maintenance, and labor) and their shares in total LCC, progressively increase relative to the front-end fixed costs (depot and bus purchase), and so does the total LCC, for all of the cases (Table 5.4). Further, as the up-front costs, in terms of which CNG has a disadvantage relative to diesel, become less predominant relative to the recurring costs (which are only marginally higher for CNG, with the higher maintenance costs for CNG being compensated for by its lower fuel costs, relative to diesel), the differential total LCCs of the CNG bus configurations relative to their diesel counterparts reduce, or, in other words, CNG becomes less uncompetitive relative to diesel in all cases (Table 5.5). So, for example, whereas the total LCC for LF/AC buses is 1.6% higher for CNG relative to diesel at 12% discount rate, it is 1.2% and 0.8% higher respectively, at 6% and 0% discount rate. However, even at 0% discount rate, the total LCC for the CNG bus configurations are higher than for their diesel counterparts.



\* Fuel prices in the data table and for the graphs are on an energy equivalent basis, and inflation adjusted. Diesel prices shown are USD/L diesel. CNG prices are USD/L diesel energy equivalent. Source: India: MyPetrolPrice.com (2016); USA: EERE (2016).

## Figure 5.3: CNG and diesel fuel prices and ratios, India and USA, 2002-2015

#### 5.3.4 Sensitivity analysis

The previous section shows how various factors affect LCC in the different cases that we considered. In particular, note that, while fuel costs account for a significant share of total LCC across fuel systems and bus technologies, the component factors, namely, fuel economy and fuel price, are subject to large variations, owing to operating conditions, and market forces and government policy respectively. These variations strongly affect fuel costs, and total LCC. Figure 5.3 shows the variations over time in the CNG and diesel prices, and the CNG-to-diesel fuel price ratios in India, as opposed to the USA, where fuel prices are more market-driven. Indeed, fuel costs were a key contributor to the differential in total LCC for CNG relative to diesel being lower for the LF and LF/AC than for the Std. buses.

It is for these reasons that we conducted a sensitivity analysis to investigate how the fuel economy and fuel price of CNG relative to diesel would affect the LCC for these two

fuel systems. The results of these sensitivity analyses are depicted in Figures 5.4a and 5.4b, which show the ratio of the present value of LCC for diesel relative to CNG (at 12% discount rate in both cases) plotted against the ratio of fuel economy for CNG relative to diesel (in Figure 5.4a), and against the ratio of the energy equivalent fuel price of CNG relative to diesel (in Figure 5.4b), for both Std. and LF/AC buses in India.



Figure 5.4: Sensitivity of LCC to fuel economy and fuel price in India

In both Figures 5.4a and 5.4b, the horizontal dashed line represents a diesel-to-CNG LCC ratio of 1, meaning that CNG is more competitive relative to diesel above, and less so below, the line, in terms of LCC. Further, in Figure 5.4a, we show two vertical lines, representing the CNG-to-diesel fuel economy ratio of 0.56 and 0.62 for the Std. and LF/AC buses respectively (as in Table 5.3), for reference. Similarly, in Figure 5.4b, the three vertical lines represent a CNG-to-diesel fuel price ratio of 0.40, 0.56 and 0.61, being respectively the lowest level of this ratio (which occurred in Delhi in 2008), the average over the most recent five year period (2011-15), which we have used in our analysis throughout, and the highest level of this ratio (recorded in Delhi in 2002).

As is evident from Figures 5.4a and 5.4b, the current values of CNG-to-diesel fuel economy and fuel price ratios are much closer to what these ratios need to be for CNG

to become competitive with diesel, for LF/AC buses than for Std. buses. Secondly, note that the fuel economy has to be much higher for CNG -- and the energy equivalent fuel price much lower for CNG -- relative to diesel, for CNG to be competitive with diesel in terms of LCC on Std. as opposed to on LF/AC buses. At the same time, however, the diesel-to-CNG LCC ratio increases somewhat more steeply as the CNG-to-diesel fuel economy ratio increases, and decreases somewhat more steeply as the CNG-to-diesel fuel price ratio increases for LF/AC than for Std. buses. In other words, the competitiveness of CNG relative to diesel is more sensitive to the CNG-to-diesel fuel economy and fuel price ratios on LF/AC than on Std. buses. These results reflect our earlier points that a) the share of fuel costs in total LCC is higher for the LF buses than on the Std. buses (and the highest for the LF/AC buses); and b) because the total LCC is already higher for the LF buses relative to Std. buses.

Finally, Figures 5.5a and 5.5b show the sensitivity of these results to the discount rate; specifically, how the results depicted in Figures 5.4a and 5.4b are affected when using a 0% as opposed to a 12% discount rate. These figures show that the break-even fuel economy ratio was reduced, and the break-even fuel price ratio was increased, for both the Std. and LF/AC bus configurations, as a result of a 0% discount rate being chosen instead of a 12% discount rate. This means that CNG can be competitive with diesel in terms of LCC at a lower fuel economy, and at a higher fuel price, relative to diesel, the lower the discount rate that is selected. Also note that the break-even fuel economy and fuel price ratios are affected more for the Std. bus relative to the LF/AC bus, and that, correspondingly, the difference in these break-even ratios for these two bus configurations narrows, when a 0% rather than a 12% discount rate, is chosen. As discussed in the previous section, these effects are because of the much higher up-front capital costs of the LF/AC buses, which are not affected by the discount rate; meanwhile, the recurring costs, which are proportionally more predominant, as a share of the total LCCs, in the case of the Std. buses (relative to the LF/AC buses), become even more so when the discount rate is 0% as opposed to 12%.



Figure 5.5: Sensitivity of LCC to fuel economy and fuel price – effect of discount rate

#### 5.4 Conclusions, limitations, and implications

First of all, it is important to reiterate that this paper, which analyses the life-cycle (capital, operating and maintenance) costs of various CNG urban bus configurations, over their service life, relative to their diesel counterparts in India, and key factors that influence these costs, is intended to be an input into decision-making regarding the implementation of CNG buses for mitigating urban air pollution, in concert with an assessment, which this paper does not address, of the environmental and health benefits of CNG. As noted in our Introduction, the assessment of the environmental and health benefits of CNG will require a comprehensive and detailed modeling of emissions of health-critical air pollutants and greenhouse gases, related to the various CNG bus configurations, over their service life, relative to their diesel counterparts. This modeling exercise will need to consider the emission control characteristics of the various bus configurations and fuel quality standards, and real-life operating conditions and other factors. Further, a monetary valuation of the benefits associated with the emissions effects in terms of key health-critical air pollutants and greenhouse gases will be needed, for weighing against the life-cycle costs arrived at in this paper, to conduct an

analysis to inform decision-making within a cost-benefit framework (alternatively, this analysis may be conducted within a cost-effectiveness, or a multi-criteria decision-making, framework).

Our study shows that the significantly higher life-cycle cost (LCC) of the CNG-LF/AC relative to that of the CNG-LF and in particular the CNG-Std. buses is driven by higher bus purchase, followed by fuel and maintenance costs for the LF/AC relative to the LF and Std. buses. The bus purchase and fuel costs account for a progressively higher share of total LCC, when moving from CNG-Std. to CNG-LF to CNG-LF/AC buses, with the fuel costs reflecting the significantly poorer fuel economy of the low-floor buses.

The LCCs for diesel are lower than for CNG for all three bus configurations, but implementation of CNG negatively affects the LCC of Std. bus fleets proportionately more than for the low-floor bus fleets, for which the LCC is already high. At the same time, whereas replacing diesel with CNG on air-conditioned low-floor buses would increase LCC by only around 1.6%, the LCC of CNG-LF/AC buses is 16% higher than for diesel low-floor and as much as 43% higher than for standard diesel buses; and that of even (non air-conditioned) CNG-LF buses is around 4% higher than for their diesel counterparts, and 29% higher than for standard diesel buses. Finally, the LCC of diesel LF/AC buses is 14% higher than for diesel low-floor buses, and as much as 41% higher than for standard diesel buses.

While decision-making regarding CNG implementation for mitigating urban air pollution will need to consider its life-cycle costs as well as its environmental and health benefits, as already noted, the significantly higher LCCs for the low-floor and low-floor air-conditioned buses, even for diesel, raises the questions of whether these buses are justified by increased patronage, especially by those who would otherwise use personal motor vehicles, and how they affect transit supply, and affordable transit service.

Our sensitivity analysis shows the significant effect of CNG and diesel fuel price (and fuel economy), on the LCCs of CNG relative to diesel. Both diesel and CNG prices, and
the CNG-to-diesel price ratio, are lower in India relative to the USA, where these prices are more market-driven. The LCCs for CNG would be much higher relative to diesel for all the bus configurations in our study, if the CNG-to-diesel fuel price ratio were higher, as in the USA. The wide variations in CNG and diesel prices, and their critical importance for life-cycle costs and the competitiveness of CNG, demonstrate the need for careful fuel pricing policies when CNG is implemented in public bus transit.

# Chapter 6: Cost-effectiveness of CNG implementation in the public bus transit fleet in Delhi

## **Chapter overview**

In Chapter 6, I draw from the results and insights from the previous analytical chapters, and address the broader overarching question of this dissertation, regarding the desirability of CNG implementation in urban bus fleets in Delhi, and India. Specifically, (i) I investigate if CNG implementation was a cost-effective choice when it was mandated by the Supreme Court, and (ii) if CNG still makes sense today, given the vehicle and emissions technologies currently available. To address these two questions, I use cost-effectiveness (CE) analysis and scenario analysis. Data for this research was drawn from the life-cycle costs in Chapter 5, while emissions data that is applicable to Indian operating conditions, and specific to each vehicle and fuel technology, was drawn from various published and expert sources. The CE analysis was done by calculating the ratio of cost differences for CNG compared to Diesel over the emissions difference (for Diesel vs. CNG), for Standard and LF bus types. This analysis was also conducted for CNG LF versus Diesel Standard buses, which some cities may consider, and for Diesel Standard buses conforming to the most recent emissions norms. Emissions were estimated for key air pollutants affecting human health (PM and NO<sub>X</sub>) and for climate change inducing greenhouse gases (CO<sub>2</sub>, CH<sub>4</sub>, BC).

# 6.1 Introduction

In the 30-year period, from 1986 to 2016, Delhi's motor vehicle fleet grew 11x, while its road infrastructure kilometres grew only about 2x, and today the city has a fleet of more than 9 million vehicles, including motorcycles (GNCTD, 2017). As discussed in previous chapters, this rapid motorization has produced a range of negative impacts, amongst which a special cause for concern is poor air quality and related risks to human health. Given this challenge, various policy actions have been taken over the years by the Indian and Delhi governments to mitigate transport emissions. Delhi being the national

capital, and given its serious air quality problems, many of these policies were first implemented there, as discussed in Chapter 4, including the Supreme Court of India ruling in 1998 that mandated the use of compressed natural gas (CNG) in its public vehicle fleet.

Estimates of the impact in Delhi of CNG policy and retirement of diesel buses on emissions in the early 2000s, suggest that these effects were mixed, for example, with CO and SO<sub>2</sub> experiencing a significant decline, but PM<sub>10</sub> levels only marginally decreasing, while NO<sub>X</sub> levels actually increased after conversion (Kathuria, 2005; Jalihal & Reddy, 2006; Reynolds & Kandlikar, 2008; Kumar & Foster, 2009). Moreover, after 2005, PM<sub>10</sub> ambient concentrations in Delhi increased steadily and steeply (CPCB, 2012). A recent activity-based emissions inventory for Delhi shows transport sector contributions of 17% for PM<sub>2.5</sub>, 13% for PM<sub>10</sub>, 2% for SO<sub>2</sub>, 53% for NO<sub>x</sub>, 18% for CO, and 51% for VOC; with the bulk of transport PM<sub>2.5</sub> emissions (i.e., >50%) originating from heavy-duty and light-duty trucks (Guttikunda & Goel, 2013). From a public health perspective, PM<sub>2.5</sub>, NO<sub>X</sub>, SO<sub>2</sub> and O<sub>3</sub> emissions are critical since high level of exposure to these pollutants are associated with elevated risk of human respiratory and cardiovascular morbidity and mortality. PM2.5 ambient levels are significantly higher in Delhi than local ambient air quality standards stipulate (Kumar et al., 2017), exposure to which is estimated to contribute to 7,350-16,200 premature deaths every year in the city (Guttikunda & Goel, 2013). From a broader perspective, it should be recognized that a fuel mandate such as CNG, while motivated by the desire to improve local air quality, might also allow other environmental goals to be achieved, such as reductions in climate change impacts. In this regard, the combustion of fossil fuels releases criteria air pollutants, greenhouse gases, and aerosols, and some pollutants like PM have both health as well as climate effects (Reynolds & Kandlikar, 2008). Therefore, there is great importance in also evaluating the effectiveness of lowering climate-forcing emissions resulting from policies such as CNG in Delhi, in addition to air pollutants affecting human health.

From a transport perspective, the challenge with evaluating the effectiveness of a policy like CNG, as pointed out by Kathuria (2005), is that a wide range of factors, apart from vehicle and fuel technology, influences air pollution. For example, factors can include vehicle mix on roads, operating conditions, fuel quality on other vehicles, and traffic congestion, all of which change over time. Also important in evaluating emissions effectiveness is the rapidly growing motorization activity that Delhi continues to experience, which is outpacing growth in road infrastructure capacity, as noted above, and the fact that commuters have been shifting trips away from public transit and towards personal motor vehicles. So, notwithstanding improved fuel quality and CNG use, the fact is that any potential contribution to improved air quality in Delhi from these and other measures may be likely being offset by the rapid increase in motor vehicle activity (Narain & Krupnick, 2007). From this perspective, since emissions outcomes are dependent on not only characteristics of CNG technology but also a host of other factors, the evaluation of fuel policy effectiveness should, as far as possible, account for these factors that also affect emissions outcomes.

Specifically considering the evaluation of CNG policy outcomes in Delhi, as far as public bus transit is concerned, there are some interesting issues and evidence worth highlighting. For instance, Reynolds and Kandlikar (2008) argue that the use of CNG in buses in Delhi resulted in a 20% reduction in carbon dioxide (CO<sub>2</sub>) equivalent emissions relative to diesel on the same buses, if aerosols such as black carbon (BC), organic carbon (OC), and sulphur dioxide (SO<sub>2</sub>) are included in the emissions analysis. This means that there are potentially positive contributions from CNG implementation on Delhi's buses, not only in terms of mitigating air pollution and health impacts, but also in terms of mitigating GHG emissions and climate change impacts, since PM contributes to both effects.

While studies that have evaluated CNG in bus systems using CE analysis are limited, a few references are worth mentioning (Rabl, 2002; Cohen et al., 2003). Furthermore, there is a limited number of academic studies that have considered both emissions outcomes and costs of fuel policies in countries such as India. For the particular case of

the implementation of CNG as an alternative fuel on the public bus fleet in Delhi, as discussed in Chapter 4, studies have mainly focused on quantifying emissions performance, which are of course important, but have not considered financial and operational dimensions. So, there is an important contribution to knowledge that can be made by evaluating the outcomes of Delhi's CNG policy using a CE framework that integrates the LCCs of CNG implementation with the existing emissions performance research. Indeed, this was the reason for the research on operational and financial performance in Chapter 4, and for the quantification of life-cycle costs in Chapter 5.

Objective To further contribute to this important policy-analytic objective, I evaluate the CE of the implementation of CNG on past and current CNG bus technologies relative to diesel, by comparing the related life-cycle costs against their respective emissions outcomes. Overall, this study addresses two major questions. Firstly, it investigates the CE of CNG implementation in the public bus fleet in Delhi, by evaluating the costs and emissions performance of buses that were used in the conversion to CNG in the early 2000s in that city. Secondly, for Delhi, but also more broadly the entire Indian context, this study analyzes the merit of CNG implementation in a public bus fleet today considering that CNG and diesel buses must comply with stricter Indian emissions standards. Therefore, for this research I specifically analyze the CE of Standard, Low-floor and Low-floor AC CNG buses, relative to diesel. Costs are evaluated based on the life-cycle calculations for these bus types in Chapter 5, while emissions effectiveness is evaluated in terms of critical pollutants affecting human health (PM and NO<sub>X</sub> emissions), and also climate change impacts in terms of CO<sub>2</sub> equivalent emissions (the rationale for having focused on these pollutants will be discussed in Section 6.2.2, below).

## 6.2 Analytic framework, methods and data

## 6.2.1 Bus and fuel technologies evaluated

In terms of the buses and fuel technologies that were evaluated in the CE analysis<sup>31</sup>, while the focus is on comparing CNG to diesel buses, the challenge is that over time, diesel buses have had a lowered emissions profile, given more stringent vehicle and emissions standards, the improvement in diesel quality (e.g., with use of lower sulphur diesel), and the possibility of use of engine and exhaust aftertreatment devices, as discussed in Chapter 3. These issues were accounted for in the analysis by considering CNG buses, but also cleaner diesel buses, relative to the diesel buses that were replaced by CNG. Furthermore, for current bus and fuel technologies, I focus on like-to-like buses for comparison; for example, CNG-LF buses were compared to Diesel-LF buses, which comply with the most recent emissions standards in India. The different types of diesel buses to which various CNG and cleaner diesel buses are being compared to (i.e., diesel buses that are the basis of comparison), will be termed, hereafter, as "baseline" diesel buses.

The cost data for the CE analysis was based on DTC for the Standard CNG (CNG-Std.), Low-floor CNG (CNG-LF) and Low-floor AC CNG (CNG-LF/AC) buses, as in Chapter 5<sup>32</sup>. For comparable diesel buses in India, I use operational and cost performance data related to the diesel bus fleets in Bangalore, Chennai and Mumbai, as justified in Chapter 5. Thus, the CNG and diesel bus technologies analyzed reflect those operated by the largest public transit fleets in the major metropolitan areas of India; and the analysis is based on actual on-road CNG and diesel bus performance in these areas. Please see Table 5.1 (p.106) for a summary of the key vehicle and engine characteristics of buses analyzed in this chapter.

<sup>&</sup>lt;sup>31</sup> Note that a detailed discussion of the CE analytic framework is presented in Section 6.2.2, below.

<sup>&</sup>lt;sup>32</sup> While DTC is not the only operator in Delhi, as noted in Chapter 4, it operates roughly 50-60% of Delhi's public urban bus transit fleet. The DMRC (Delhi Metro Rail Corp.) and a cluster scheme of privately owned operators run the remainder of Delhi's public bus fleet, for which their technological choices largely mirror the technologies and standards currently in use at DTC. For this reason, operational data and technological choices which are the basis of the evaluation for Delhi in this chapter are based on DTC, and analysis is done on a per-bus basis.

In addition to fuel and bus technology, emissions are also highly influenced by fuel quality and standards, particularly for diesel vehicles. Over the years, India has adopted an increasingly stringent phased vehicle and fuel emissions control program starting with the Bharat Stage I (BS-I), equivalent to EURO-I (see Table C1, Appendix C). When DTC adopted CNG, Delhi's National Capital Region was transitioning from BS-I to BS-II emission standards, though most of India was only bound by the BS-I standard. Several cities in India<sup>33</sup>, including many large metropolitan areas, started to comply with the BS-IV emission standard in 2010, while the rest of the country finished the transition to the BS-IV standard only in 2017, equivalent to the EURO-IV standard. BS-IV diesel<sup>34</sup> has a lower sulphur content than its predecessor BS-III<sup>35</sup> (SIAM, 2013). Table C1 (Appendix C) details heavy-duty diesel and CNG engine emissions standards in India and the European Union for comparative purposes.

So, for the CE analysis, I grouped the buses according to the emission standards pertinent to the period when buses were in operation. In the case of DTC, CNG-Std. buses replaced Diesel-Std. in the early 2000s, when vehicles complied with Bharat Stage I or II emission standards (BS-I; BS-II). So, CNG-Std. and Diesel-Std. buses are evaluated in terms of BS-I or II compliant vehicles, and used in the CE analysis in order to address the first research objective of evaluating the effectiveness of the CNG policy as carried out in early 2000s. Also, I considered the next generation of diesel buses, which would have complied in the early 2000s with the next phase of emission standard in India, BS-III (see Table C1, Appendix C).

Finally, towards 2007-08, DTC started replacing the ageing fleet of CNG-Std. buses with low-floor CNG buses. Of these low-floor buses, approximately 80% were non-air-

<sup>&</sup>lt;sup>33</sup> Delhi (NCR), Mumbai, Kolkata, Chennai, Bengaluru, Hyderabad, Ahmedabad, Pune, Surat, Kanpur, Lucknow, Sholapur, Jamshedpur, Agra, Puducherry, Mathura, Vapi, Jamnagar, Ankaleshwar, Hissar and Bharatpur (Guttikunda & Mohan 2014).

<sup>&</sup>lt;sup>34</sup> Maximum sulphur content of 50 ppm; minimum cetane number of 51; fuel density of 820 to 845 kg/m<sup>3</sup> (SIAM, 2013).

<sup>&</sup>lt;sup>35</sup> Maximum sulphur content of 350 ppm; minimum cetane number of 51; fuel density of 820 to 845 kg/m<sup>3</sup> (SIAM, 2013).

conditioned and 20% were air-conditioned (Chapter 4). Given the period that the lowfloor buses were purchased at DTC, the buses complied with BS-III or BS-IV standards. So, BS-III and IV compliant CNG-LF and CNG-LF/AC buses -- as well as their diesel counterparts -- were assessed to address the second research question, that is, to verify if there are any significant changes in CE with adoption of more stringent fuel and vehicle standards in India and evaluate the merits of continuing to pursue a natural gas policy today.

## 6.2.2 Cost-effectiveness analysis

As pointed out by Boardman et al. (2001), a key reason for using CE analysis is when difficulty is involved in monetizing policy impacts (e.g., health impacts). For this reason, CE analysis has been widely used in the evaluation of emissions effectiveness of alternative fuel policies in transport (Kok et al., 2011; Browne & Ryan, 2011) and in evaluating health impacts of vehicle emissions (Cohen et al., 2003). This was the motivation in choosing the CE analysis framework for evaluating CNG implementation in Delhi's public buses in this study.

To analyze CNG implementation in Delhi's public bus fleet, I calculate the incremental CE ratio of CNG over diesel buses. I base the CE analysis on a dollars-per-avoided emissions basis, as described in Boardman et al. (2001), Kok et al. (2011), and Browne and Ryan (2011). In calculating the CE ratio, as shown in equation (1) below, the numerator of the CE ratio reflects the incremental life-cycle costs (LCCs) of CNG over diesel, based on life-cycle costs of CNG and diesel buses from Chapter 5. The denominator of the CE ratio reflects the incremental emissions reductions for CNG relative to diesel, based on the emissions differences of diesel buses over CNG buses in India. Emissions impacts were measured separately, in terms of health critical pollutants PM and NO<sub>x</sub>, and in terms of global warming potential, CO<sub>2</sub> equivalent emissions (CO<sub>2</sub>(e)), as justified below.

The CE ratio to evaluate different types of CNG buses, *CE*<sub>CNG</sub>, is defined as:

$$CE_{CNG} = \frac{LCC_{CNG} - LCC_{Diesel}}{EMISSIONS_{Diesel} - EMISSIONS_{CNG}}$$
(1)

where *LCCCNG or Diesel* is the present value of life-cycle costs (in \$-per-bus) of a particular type of CNG or diesel bus -- Standard or Low-floor. *EMISSIONSCNG or Diesel*, refers to the total amount of emissions (in grams or kilograms) of a particular type of pollutant (PM, NOx, or CO<sub>2</sub>(e)) over the entire service life of a particular bus type (CNG or diesel; Std. or LF). Therefore, the CE ratio reflects the incremental cost of a CNG bus over diesel per gram of avoided emissions, considering the vehicle's entire service life (i.e., \$ per gram, per bus). The smaller the CE ratio is, the more emissions reductions can be accomplished at a lower incremental cost, for a particular option; while a negative CE ratio indicates that CNG costs more and also increases emissions in relation to the diesel buses it is being compared to (positive numerator and negative denominator). Table 6.1 discusses the various possibilities and likelihood of various potential CE ratio outcomes.

Outcome of CE ratio	CE ratio breakdown	Likelihood of case
Positive	IF: $LCC_{CNG} > LCC_{Diesel}$ AND: $EMISSIONS_{Diesel} > EMISSIONS_{CNG}$	Expected, since LCC for CNG buses are greater than diesel buses (Chapter 5), and CNG is expected to show emissions reductions of critical pollutants, which is the reason for its implementation.
	IF: <i>LCC<sub>CNG</sub> &lt; LCC<sub>Diesel</sub></i> AND: <i>EMISSIONS<sub>Diesel</sub> &lt; EMISSIONS<sub>CNG</sub></i>	Unlikely, since information from previous chapters shows no cases where CNG LCC is less than diesel LCC. Though, as shown in Chapter 3, and will be shown below, there are particular pollutants that are emitted more by CNG buses than diesel buses.
Negative	IF: LCC <sub>CNG</sub> > LCC <sub>Diesel</sub> AND: EMISSIONS <sub>Diesel</sub> < EMISSIONS <sub>CNG</sub>	Possible, since it is expected that the numerator is positive, as per above; nonetheless, there are particular pollutants that are emitted more by CNG buses than diesel buses (negative denominator), as discussed.
	IF: $LCC_{CNG} < LCC_{Diesel}$ AND: $EMISSIONS_{Diesel} > EMISSIONS_{CNG}$	Unlikely, since information from previous chapters shows no cases where CNG LCC is less than diesel LCC.

Table 6.1: Potential CE ratio outcomes

In addition to the CE analysis, I also used scenario analysis to evaluate the broader impact of the CNG implementation in Delhi's public bus fleet. In the scenario approach, I considered emissions reductions due to CNG, given the potential impact of higher life-cycle costs on public bus transit supply. It may be expected that, in this case, supply constraints in the number of public buses using CNG affect the overall effectiveness of the emissions reductions, since travel demand may have to be satisfied by other types of transport modes and vehicles other than CNG buses, including conventional diesel buses -- if these are permitted to operate -- and other lower capacity vehicles.

In the sub-sections bellow I detail, discuss, and justify various aspects of these analytic methods.

Life-cycle costs To estimate CNG incremental costs over diesel in the CE numerator (Equation 1), I used bus life-cycle costs over a uniform 12-year period (70,000 km per year) based on the LCC model results in Chapter 5. However, LCCs presented in Table 6.2 are exclusively in US dollars (\$) per bus using a 12% discount rate, and includes a minor adaptation, in that, diesel and CNG low-floor bus LCCs in Table 6.2 are calculated using the LCC cost results for LF and LF/AC CNG and diesel buses from Table 5.4 (Chapter 5), and assuming a fleet that is 80% LF buses and 20% LF/AC buses; this fleet profile assumption approximately reflects DTC's current structure of low-floor buses as discussed.

Vehicle Id.						
no.	Fuel type	Bus type	Emission Standard	Capital <sup>(a)</sup>	Operating <sup>(b)</sup>	Total LCC
1	Diesel	Std.	BS-I/II	70,008	347,696	417,704
2	Diesel	Std.	BS-III/IV	73,194 <sup>(c)</sup>	358,453 <sup>(d)</sup>	431,647
3	Diesel	LF <sup>(e)</sup>	BS-III/IV	126,352	403,256	529,608
4	CNG	Std.	BS-I/II	87,037	356,735	443,771
5	CNG	LF <sup>(e)</sup>	BS-III/IV	149,960	398,454	548,414

 Table 6.2: Present value of bus LCCs, \$-per-bus, 12-yr service life, 12% disc.rate

Source: LCC model in Chapter 5 (see Table 5.4); converted from \$/km to \$/bus using 70,000 annual vehicle kilometres for 12 years of service life.

<sup>a</sup> Capital Costs = Depot infrastructure + Bus purchase.

<sup>b</sup> Operating Costs = Bus maintenance (labour and parts) + Fuel + Labour (bus operators & administrative) and training.

° Capital costs estimated from LCC model for Diesel Std. (BS-I/II), but with bus purchase being 6.18% higher; this

percentage is 2x the added purchase cost (to be conservative) of an EURO III/IV bus over an EURO II/III bus, as reported in Miller et al. (2017).

<sup>d</sup> Operating costs estimated from LCC model for Diesel Std. (BS-I/II), but with diesel price being 11% higher to account for the added refining cost of diesel with lower Sulphur content; this estimate was taken from TERI (2002).

<sup>e</sup> LCCs for LF is the average LCCs of a fleet consisting of 80% non-AC LF buses (i.e., LF bus in Table 5.4 terminology) and 20% LF/AC buses from Table 5.4.

**Choice of air pollutants** The CE analysis centers on two critical aspects of vehicle exhaust emissions: first, on human health effects, and secondly, on climate change effects. I focus on PM and NO<sub>x</sub> for their implications for health risks, while leaving out SO<sub>2</sub> and O<sub>3</sub>, since O<sub>3</sub> is a secondary pollutant and SO<sub>2</sub> from transport is relatively small (only 2% of total emissions for this pollutant according to Guttikunda and Goel, 2013) and plus, diesel sulphur has reduced drastically with more stringent emissions standards. At the same time, I also consider CO<sub>2</sub>, CH<sub>4</sub> and BC emissions, as these are GHGs with global warming potential, but also with important health impacts.

In terms of PM emissions, many studies do not use fine (PM<sub>2.5</sub>) or ultrafine (PM<sub>0.1</sub>) fraction size particulate matter, which are generally more relevant in terms of negative health outcomes. For example, Cohen et al. (2003) do not make any distinction between PM or PM<sub>2.5</sub> emissions, as they argue that the bulk of transport PM emissions is comprised of PM<sub>2.5</sub>. This argument is also presented in ARAI (2008). So, in the absence of substantial data on fine or ultrafine particulate matter emissions in the Indian context, I use PM data that was available, as a proxy for PM<sub>2.5</sub> emissions. For bus emission factors reported in Table 6.4, the data is in PM<sub>2.5</sub> terms, but this is not the case for lower occupancy vehicle PM emission factors in Table 6.5. In addition to PM and NO<sub>x</sub> tailpipe emissions, black carbon (BC) was also evaluated since it is an important component of PM and exposure to it is associated with probable human carcinogens

(Apte et al., 2011); moreover, BC is also an important climate-forcing agent (Bond & Sun, 2005; Ramanathan & Carmichael, 2008).

Air Pollutant	Diesel Buses	CNG Buses					
CH <sub>4</sub>	Methane emissions from diesel buses	Methane emissions from CNG buses are significant <sup>4</sup> . In addition, since					
	are very low <sup>1</sup> , and are not relevant for	leaked methane from CNG bus systems is an important contributor to					
	the CE analysis from a GHG	total CO <sub>2</sub> (e) emissions <sup>3</sup> , methane leakage was incorporated into total					
	perspective <sup>1,2,3</sup> .	methane emissions from CNG buses, as presented in Table 6.4 (below).					
BC and	In the Indian context, approximately	Approximately 29% of PM is estimated to be BC in the Indian context;					
OC	76% of PM emissions are estimated to	the remainder is OC <sup>7</sup> .					
	be BC, with the remainder being OC <sup>7</sup> .						
	For the purpose of estimating CO <sub>2</sub> (e) on	ly BC for diesel buses was considered, since OC from diesel buses					
	represents, in absolute terms, less than 1% of total CO <sub>2</sub> (e) emissions for buses; while BC and OC represent less						
	than 0.1% for CNG buses <sup>8</sup> .						
N <sub>2</sub> O	Emissions from diesel buses are, at best	t, not relevant to the CE analysis since the difference in emissions between					
	CNG and diesel buses are very small <sup>5</sup> ar	nd N <sub>2</sub> O emissions from conventional diesel and emissions controlled diesel					
	buses are small enough to be ignored, e	ven after adjustment to GWP <sup>2</sup> . So, N <sub>2</sub> O was not considered in the analysis					
	for either fuel system.						
CO <sub>2</sub>	Was included in the analysis given its sig	gnificant contribution to total CO <sub>2</sub> (e) emissions <sup>6</sup> .					
SO <sub>2</sub>	SO <sub>2</sub> is a climate cooling pollutant (Myhre	et al., 2013, p.684). For the purpose of the CE analysis, it was not					
	considered, since CNG buses do not em	it this pollutant <sup>8</sup> and the contribution of SO <sub>2</sub> to total CO <sub>2</sub> (e) emissions from					
	diesel buses is very small, as it represen	ts only about 1% of total $CO_2(e)$ emissions for these buses <sup>8</sup> .					
4							

<sup>1</sup> The calculated average CH<sub>4</sub> emissions for diesel buses is 1.66 g/mi compared to an average of 11.23 for CNG buses (Hesterberg et al. 2008; Table 1, p.6438).

<sup>2</sup> For example, Cohen et al. (2003) justify not using in impact analysis the emissions of CH<sub>4</sub> from diesel buses and N<sub>2</sub>O from CNG and diesel buses; they state that "none of the selected studies reported CH<sub>4</sub> or N<sub>2</sub>O emissions for either CD [Conventional Diesel] or ECD [Emissions Controlled Diesel] vehicles, although values reported by Ahlvik and Brandberg (2) indicate that they are small enough to ignore, even after adjustment for GWP." (SI, p.30).

<sup>3</sup> Exhaust methane from diesel buses (1.1 x10<sup>3</sup> tons of CO<sub>2</sub>(e)) represents less than 0.1% of total CO<sub>2</sub>(e) emissions, thus will not impact CE analysis if included or not; leaked methane from CNG buses represents about 11% of total CO<sub>2</sub>(e) emissions and therefore was included in the CE analysis (Reynolds & Kandlikar 2008: SI, Table S5, p.S8).

<sup>4</sup> Sources: Cohen et al. (2003), Reynolds & Kandlikar (2008), and Hesterberg et al. (2008).

<sup>5</sup> "Nitrous oxide (N<sub>2</sub>O), another potent greenhouse gas, is not included here because net mass emissions of this species are not appreciably different for diesel vs CNG engines" (Reynolds & Kandlikar 2008: Section 2.1, p.5861).

<sup>6</sup> Sources: Cohen et al. (2003), and Reynolds & Kandlikar (2008).

<sup>7</sup> The BC ratio is 76% for dissel buses; For CNG buses, though BC ratio is not stated in text, or SI, can be estimated to be 29% (Reynolds & Kandlikar 2008: p.5861-5863).

<sup>8</sup> Source: Reynolds & Kandlikar (2008): SI, Table S5, p.S8.

In terms of GHG emissions, the key climate forcing agents CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, SO<sub>2</sub> and aerosols, such as black carbon (BC) and organic carbon (OC) are generally considered and normalized to evaluate the climate change contribution of these different greenhouse gases using the Global Warming Potential (GWP) in the common metric of CO<sub>2</sub>-equivalent (CO<sub>2</sub>(e)) emissions (Myhre et al., 2013). For my CE analysis on mobile emissions from buses, I only evaluate CH<sub>4</sub>, CO<sub>2</sub>, and BC, based on the importance of these emissions in terms of GWP. The rationale for these choices, are summarized in Table 6.3, above, and also explained in Appendix C. The computations of CO<sub>2</sub>(e) in terms of GWP of these emissions are described in the next section.

**Emissions factor estimates** Ideally, emissions factors (EFs) would be based on measurements conducted on actual in-use public transit buses and under

operating conditions representative of Delhi and other Indian cities. While there is a vast amount of literature on emission factors, based on in-use buses for different types of diesel and CNG bus technologies internationally (e.g., WVU, 2005; Nylund et al., 2007; Hesterberg et al., 2008; Wayne et al., 2008), there are few if any reliable measurements of emissions of in-use public transit buses under operating conditions in Delhi, or India for that matter. Emission factors used in the CE model, were based on the limited literature related to vehicle exhaust emission factor measurements in India, conducted on chassis dynamometers (Bose & Sundar, 2005; ARAI, 2008; ARAI, 2009; CPCB, 2010). Note in this regard that emission measurements in traffic are less reproducible than those conducted on chassis dynamometer, as argued in Hesterberg et al. (2008).

The EFs that were used in the CE analysis, and the sources for these factors are provided in Tables 6.4 and 6.5, with details on data sources and EF calculations provided in the footnotes. Table 6.4 shows the EFs for each bus type in terms of PM, NO<sub>X</sub> and CH<sub>4</sub>, CO<sub>2</sub>, BC, and the CO<sub>2</sub>(e) that was calculated based on CH<sub>4</sub>, CO<sub>2</sub> and BC emissions using methods explained in Appendix C. Methane (CH<sub>4</sub>) emissions in Table 6.4 incorporate leakage, since this is a significant factor contributing to emissions, as described in Appendix C. Also, specific details on CO<sub>2</sub> emissions estimates based on fuel economy data, BC emission factor estimates based on PM emissions, and CO<sub>2</sub>(e) are all described in Appendix C.

For the scenario analysis, which also considered effects on emission effectiveness given mode shifts, Table 6.5 provides emission factors for cars, M2Ws (motorized two-wheeled vehicles) and M3Ws (motorized three-wheeled vehicles). EF calculations in this table are for the most prevalent and relevant fuels for each vehicle type; this approach in estimating EFs was also adopted in a study of traffic–generated emissions in Delhi (Sindhwani & Goyal, 2014).

Table 6.4: Emission	n factors for	buses, g/km.
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Vehicle			Emission	PM <sup>(a)</sup>	NO <sub>x</sub> <sup>(a)</sup>	CH <sub>4</sub> <sup>(b)</sup>	CO <sub>2</sub> <sup>(c)</sup>	FE <sup>(c)</sup>	BC <sup>(d)</sup>	CO <sub>2</sub> (e) <sup>(e)</sup>
ld. no.	Fuel	Bus	Standard	(g/km)	(g/km)	(g/km)	(g/km)	(km/l)	(g/km)	(g/km)
1	Diesel	Std.	BS-I/II	7.00	45.30	1.20	708	3.81	3.221	2666
2	Diesel	Std.	BS-III/IV	2.47	18.12	0.27	860	3.08	1.229	1602
3	Diesel	LF	BS-III/IV	2.67 <sup>f</sup>	21.38 <sup>f</sup>	0.27 <sup>f</sup>	1035	2.56	1.229 <sup>f</sup>	1777
4	CNG	Std.	BS-I/II	0.49	54.36	6.11	907	2.12	0.122	1150
5	CNG	LF	BS-III/IV	0.18	25.66	3.16	1205	1.60	0.046	1321

<sup>a</sup> Source: Goel & Guttikunda (2015), which in turn bases EFs on CPCB (2010) as primary source data, but also incorporating the effects of "emission standards applicable for newer fleet by year, corrections applicable for engine deterioration based on fleet age-mix, and changing idling conditions on the roads" (p.83). EFs for BS-I/II buses were taken from Supplementary Material and BS-III/IV buses from Table 5 in Goel & Guttikunda (2015; p.83).

<sup>b</sup> Source: <u>CNG buses</u>: EFs reflect average of EF results found in Bose & Sundar (2005), ARAI (2008, 2009), and Reynolds & Kandlikar (2008); <u>Diesel buses</u>: EFs reflect average EF results found in Bose & Sundar (2005), and ARAI (2008, 2009).

<sup>c</sup> Source: all FE data taken from Chapters 4 and 5 (with CNG on energy equivalent basis), but Diesel-Std. BS-III/IV which was taken from Goel & Guttikunda (2015). CO<sub>2</sub> EFs calculated from FE data as explained in Appendix C (Table C2).

<sup>d</sup> BC EFs calculated from PM using methods provided in Appendix C.

<sup>e</sup> CO<sub>2</sub>(e) calculated from CH4, CO<sub>2</sub> and BC using methods provided in Appendix C.

<sup>f</sup> A correction factor, as reported in IVEM (2008), Clark et al. (2002), and EEA (2013), was used to estimate PM, NO<sub>x</sub> and BC emission factors for this bus category, based on differences in bus weight of LFs to Std. Diesel buses. For CH<sub>4</sub>, no difference was observed between different classes of diesel buses (Clark et al. 2002), so the same emission factor was used for Diesel-Std. (BS-III/IV) and Diesel-LF (BS-III/IV).

Table 6.5: Emission factors for cars, M2Ws, and M3Ws, g/km.

Id. no.	Vehicle	Fuel	РМ <sup>ь</sup>	NOx <sup>b</sup>	CH₄°	$CO_{2^{d}}$	BC <sup>e</sup>	CO <sub>2</sub> (e) <sup>f</sup>
6	Car	Gasoline	0.050	0.243	0.17	223.6	0.160	324
7	M2W <sup>a</sup>	Gasoline	0.058	0.172	0.18	26.6	0.013	39
8	M3W	CNG	0.041	3.633	1.84	62	0.008	118

<sup>a</sup> M2Ws, emission factors reflect weighted average of 2-stroke (2S) and 4-stroke (4S) engines, assuming M2W fleet proportion numbers are 72:28 for 2S and 4S, respectively (Sindhwani & Goyal 2014).

<sup>b</sup> Source: Goel & Guttikunda (2015).

<sup>c</sup> Source: Sindhwani & Goyal (2014); methane includes leakage based on Reynolds & Kandlikar (2008).

<sup>d</sup> Source: Car + M2W: Ramachandra & Shwetmala (2009); M3W: Reynolds & Kandlikar (2008).

<sup>e</sup> Source: Car + M3W: Reynolds & Kandlikar (2008); M2W: Sindhwani & Goyal (2014).

<sup>f</sup> CO<sub>2</sub>(e) calculated from CH<sub>4</sub>, CO<sub>2</sub> and BC using methods provided in Appendix C.

Vehicle usage and total emissions

To calculate total bus emissions

over its service life for each air pollutant and vehicle type (Std. or LF; CNG or diesel), emission factors (g/km) were taken from Tables 6.4 and 6.5 and multiplied by estimated vehicle usage in vehicle kilometres (VKMS)<sup>36</sup>, over a 12-year cycle. DTC data shows that VKMS in the 5-year period (2006-07 to 2010-11) averaged 67,041 kilometres-per-year, per-bus, but with a slight increasing trend in recent years (CIRT, 2016; CIRT, 2017); so, the emissions calculations were based on a constant usage of 70,000 kilometres per-year, per-bus (as in Chapter 5).

<sup>&</sup>lt;sup>36</sup> For buses, vehicle-kilometres (VKMS) refer to the average effective kilometres operated by a vehicle within a fleet of vehicles. Effective kilometres are revenue earning kilometres (CIRT 2009, p.25).



Figure 6.1: Comparison of passengers carried and load factor at DTC

# 6.2.3 Scenario analysis

For the scenario analysis, because of the higher LCCs for CNG compared to Diesel, and because of operator's budgetary constraint, I investigate what happens to bus capacity, and how this then impacts ridership, and in turn affects emissions. So, modal choice and bus capacity changes were incorporated in the emissions modeling, and particularly how some trips will likely move from buses to other transport modes, all of which will have important environmental implications.

There is in fact evidence of this impact of higher costs on the supply of buses in Delhi; since DTC converted to CNG in the early 2000s, its carrying capacity-kilometres (CAPKM) on a per-population basis was actually 37% lower in the 10-year period from 2000-01 to 2010-11 than the prior 10-year period of diesel only operations, from 1990-91 to 1999-00 (CIRT, 2012; GNCTD, 2017). Of course the introduction of Delhi's metro system is also a factor, but it only accounts for a very small mode share overall during this period (Tiwari, 2017). Even with recent upgrades to Delhi's public transit bus fleet, with the introduction of low-floor buses and considering fleets operated by other operators<sup>32(p.130)</sup>, CAPKM on a per-capita basis is most likely declining, due to ageing and retirement of the old fleet of standard CNG buses, and lack of recent procurements (The Times of India, 2016; Hindustan Times, 2017). The consequence of reductions in bus capacity is that bus loading may increase and may also have the unintended

outcome of causing passenger trips to be conducted by other modes of transport, rather than by bus, thus affecting the overall effectiveness of CNG implementation. Actually, this is what Figure 6.1 shows, when we compare the number of passengers carried to the load factor at DTC from 2006-07 onwards; during this period, the load factor at DTC has been steadily increasing (see box highlight A in Figure 6.1) whereas the number of passengers carried has been relatively flat (see box highlight B in Figure 6.1).

To hypothetically keep bus loading and quality of service constant in the event that the supply of CNG buses is constrained by higher costs, it would be reasonable to assume that part of existing bus trips would have to be provided via baseline diesel buses, assuming these could continue to operate. Since this is not in the case in Delhi, as only CNG buses can operate as per policy, lower bus supply will push trips to other modes of transportation, including cars, motorized two-wheeled vehicles (M2Ws), and intermediate public transit. The evidence points to this, with Goel and Guttikunda (2015) showing a steep decrease in public transit mode shares in Delhi from 1994 to 2007, even when metro is included; in the meantime, the modal shares of private vehicles were up considerably. From the point of view of trip demand (measured in passenger-kilometres), we can assume it to be constant with respect to changes in bus technology; so, a bus supply constraint may cause part of trip demand to migrate to other modes of transport.

Given these considerations, the analysis of the effectiveness of CNG implementation was conducted using three different independent scenarios. Scenario 1 assumed there is no budget constraint on the operator's part. In this first scenario, all diesel buses are converted to CNG. Meanwhile, to evaluate the (hypothetical) cleaner diesel buses alternative, this scenario also assumed a full conversion to diesel buses that would comply with more stringent emissions standards (i.e., a fleet consisting of mix of BS-III and IV). Scenario 1 is meant to serve as a benchmark to evaluate the full benefits of using CNG, against which the effectiveness of emissions reductions, as a result of budget constraints, can be compared. Scenarios 2 and 3 assume that the higher costs of CNG will reduce the number of buses using this fuel vis-à-vis the baseline (diesel),

and that, therefore, existing bus trip demand will either be met by keeping some existing baseline diesel buses in operation (see further Scenario 2 discussion below), or via nonbus options (see further Scenario 3 discussion below).

The budget restriction is assumed to affect CNG bus fleet capacity, as follows:

CNG Fleet Capacity (CAPKM)<sub>i</sub> = 
$$\frac{LCC_{base}}{LCC_i}$$
 (2)

where *i* represents the CNG buses (Standard or LF) being compared to baseline diesel buses (Standard or LF). In Scenarios 2 and 3, CNG fleet capacity will be smaller than baseline, since diesel LCCs are smaller than CNG LCCs. Using LCCs from Table 6.2 in Equation 2, Std. CNG fleet capacity will be 0.941 of a Std. diesel bus fleet (i.e., there is a loss of 5.87% in bus supply capacity compared to a Std. diesel bus fleet), and LF CNG bus fleet capacity will be 0.966 of LF diesel buses (i.e., a 3.43% loss in supply capacity).

In Scenario 2, the smaller number of CNG buses put into operation, creates a gap in supply of bus transit services that is covered via use of baseline diesel buses that are not entirely retired, thus keeping the overall supply capacity of bus services constant from pre- to post-CNG conversion. However, note that Scenario 2 is purely hypothetical in the case of Delhi, since the Supreme Court of India ruling concerning CNG implementation did not allow diesel buses to operate within the city. Nevertheless, this second scenario may be valid for other contexts where CNG implementation is sought and restrictions do not apply. So, for the broader Indian context, a mix of Scenarios 2 and 3 are more likely relevant.

In Delhi's case, specifically, since diesel buses cannot operate in the city, Scenario 3 is the most relevant for the purpose of analyzing CNG implementation in the city's public bus fleet. The gap in bus supply resulting from higher CNG costs will change the composition of trip demand. Scenario 3 assumes that, because of the smaller number of buses, part of the existing bus users remain in the current system with an increase in bus loading, and that, part of existing bus users will migrate to lower occupancy motor vehicles for their transport needs due to the decrease in bus service quality (smaller bus supply and higher loading). In the analysis of Scenario 3, lower occupancy motor vehicles include intermediate public transit (taxies, jeeps, and auto-rickshaws), private cars, and motorized two-wheeled vehicles (M2Ws).

Estimation of the migration of bus trips to other modes, due to smaller bus supply was based on the socio-demographic profile of bus users in Delhi. A recent study showed that 34% of users cannot afford M2Ws (Suman, Bolia, & Tiwari, 2017), so, it was assumed that this is the percentage of existing bus users that will either continue to use the new CNG buses (via increase in loading) or migrate to non-motorized transport. Meanwhile, since 66% of bus users have the means of buying and using M2Ws or cars (Suman, Bolia, & Tiwari, 2017), this is the percentage of bus users assumed to change trips from buses to lower occupancy motor vehicles; this group was allocated to cars (including taxies and jeeps), M2Ws, and M3Ws, using average modal split data for Delhi as shown in the literature (Sahai, Bishop & Singh, 2009; Tiwari, 2017).

Table 6.6 shows how bus ridership migration to private modes of transport for Scenario 3 was calculated. Considering the 5.87% loss in bus supply capacity for the CNG-Std. bus fleet (Equation 2), and that 3.88% passenger-kilometres (i.e., 66% of 5.87%) will migrate to lower occupancy motor vehicles, this means that of the 3.2 million annual bus passenger-kilometres<sup>37</sup> provided by the average public transit bus in Delhi there will be an annual reduction of approximately 124,000 passenger-kilometres in bus ridership, per bus. For the LF CNG buses, since supply capacity decreases by 3.43% (Equation 2), it will result in a 2.26% loss in bus ridership, with a reduction of approximately 72,000 passenger-kilometres (PKMS) in ridership per bus, per year. Given these reductions in bus ridership, for which transport needs will be met via lower occupancy motor vehicles, PKMS in Scenario 3 will be divided according to the mentioned modal split (Table 6.6). VKMS were estimated from these PKMS needs and average vehicle occupancy ratios

<sup>&</sup>lt;sup>37</sup> 3.2 million is the average PKMS per bus on road in DTC's fleet over a 5-year period, from 2006-07 to 2010-11 (CIRT, 2008-2012).

by dividing PKMS of each mode by the average vehicle occupancy for each mode: 2.75 for cars (including taxies), 1.3 for M2Ws and 2 for M3Ws, respectively (Reddy & Balachandra, 2012). So, in Scenario 3, fleet emissions will include those from a smaller number of CNG buses, plus those from public transit ridership having switched to lower occupancy motor vehicles, as shown in Table 6.6.

	[0]	[1]	[2]	[3]	[4]				
	Motorized	Motorized modal	Average vehicle	Change	Change in				
	modal	shares in Delhi	occupancy	in	VKMS				
	shares	(assuming bus	(passengers per	PKMS	= [3] / [2]				
	in Delhi <sup>1</sup>	ridership changes) <sup>2</sup>	vehicle) <sup>3</sup>	(per-bus/yr)	(per-bus/yr)				
Impact of change in Std. CNG bus ridership: -3.88% in bus ridership = -123,765 PKMS/bus/year <sup>5</sup> with the following									
changes in moda	al shares:								
Cars <sup>4</sup>	19.7%	20.4%	2.75	+49,533	+18,004				
M2Ws	26.2%	27.1%	1.3	+65,854	+50,657				
M3Ws	3.3%	3.4%	2	+8,377	+4,189				
Buses	45.1%	43.4%	50	-123,765	-2,475				
Metro/Train	5.7%	5.7%							
Total	100%	100%							
Impact of change	e in LF CNG bus ri	dership: -2.26% in bus ride	ership = -72,249 PKM	'S/bus/year <sup>6</sup> with	the following				
changes in moda	al shares:								
Cars <sup>4</sup>	19.7%	20.1%	2.75	28,916	10,510				
M2Ws	26.2%	26.7%	1.3	38,443	29,572				
M3Ws	3.3%	3.4%	2	4,890	2,445				
Buses	45.1%	44.1%	50	-72,249	-1,445				
Metro/Train	5.7%	5.7%							
Total	100%	100%							

Table 6.6: Impact of bus ridership migration to private modes of transport (Scenario 3)

<sup>1</sup> Source: Sahai, Bishop & Singh (2009), Reddy & Balachandra (2012), Tiwari (2017).

<sup>2</sup> Source: These estimates take into account the net loss in bus ridership, as noted in text, and redistribution of these passenger kilometres to other private modes of motorized transport (assuming modal shares in column [0]).

<sup>3</sup> Source: for Cars 2Ws and 3Ws from Reddy & Balachandra (2012); for buses, based on average bus load% and bus capacity at DTC, over a 5-yr period, from 2006-07 to 2010-11 (CIRT, 2008-2012).

<sup>4</sup> Car also includes Jeeps, and Taxies.

<sup>5</sup> Bus PKMS ridership decrease estimate = 3.87682% (x) 3,192,430 PKMS per bus, per year; the PKMS per bus on road was based on DTC's PKMS 5-year fleet average divided by DTC's 5-year average buses on road from 2006-07 to 2010-11 (CIRT, 2008-2012).

<sup>6</sup> Bus PKMS ridership decrease estimate = 2.26314% (x) 3,192,430 PKMS per bus, per year; the PKMS per bus on road was based on DTC's PKMS 5-year fleet average divided by DTC's 5-year average buses on road from 2006-07 to 2010-11 (CIRT, 2008-2012).

			EMISSIONS <sup>1</sup>						
	Annual passenger	Annual vehicle	PM	% change relative	NO <sub>X</sub>	% change relative	CO <sub>2</sub> (e)	% change relative	
	kilometres	kilometres	(g/PKMS)	to baseline	(g/PKMS)	to baseline	(g/PKMS)	to baseline	
Scenario 1 - No Capital Restriction:	emissions of bus	s fleet reflects full ben	efit of conversion	on to CNG					
Diesel-Std.(BS-I/II), baseline	$\wedge$	$\wedge$	0.15353		0.9933		58.47		
vs. CNG-Std.(BS-I/II)			0.01068	-93.0%	1.1919	+20.0%	25.22	-56.9%	
vs. Improved diesel-Std.(BS-III/IV)	3,192,430	70,000	0.05425	-64.7%	0.3973	-60.0%	33.94	-42.0%	
Diesel-LF(BS-III/IV)			0.05859		0.4688		38.96		
vs. CNG-LF(BS-III/IV)	$\vee$	$\vee$	0.00403	-93.1%	0.5626	+20.0%	28.97	-25.7%	
Scenario 2 - Capital Restriction: em	issions reflect pa	artial conversion to CN	VG buses; rema	ining ridership c	omplemented	by Diesel buse	S		
Diesel-Std.(BS-I/II), baseline	3,192,430	70,000	0.15353		0.9933		58.47		
VS.									
CNG-Std. (BS-I/II)	3,004,908	65,888							
and Diesel-Std. (BS-I/II)	<u>187,522</u>	<u>4,112</u>							
Total	3,192,430	70,000	0.01576	-89.7%	1.1849	+19.3%	26.40	-54.8%	
Diesel-LF(BS-III/IV), baseline	3,192,430	70,000	0.05859		0.4688		38.96		
VS.									
CNG-LF(BS-III/IV)	3,082,962	67,600							
<u>Diesel-LF(BS-III/IV)</u>	<u>109,468</u>	<u>2,400</u>							
Total	3,192,430	70,000	0.00517	-91.2%	0.5607	+19.6%	29.17	-25.1%	
Scenario 3 - Capital Restriction: Fle	et emissions refl	ect partial fleet conve	rsion to CNG bu	uses; part of bus	ridership mig	rates to LWVs	(Cars*, M2Ws	s, & M3Ws)	
Diesel-Std.(BS-I/II), baseline	3,192,430	70,000	0.15353		0.9933		58.47		
VS.									
CNG-Std. (BS-I/II)	3,068,666	67,525							
and M2W	65,854	50,657							
M3W	8,377	4,189							
<u>Cars<sup>2</sup></u>	<u>49,533</u>	<u>18,004</u>							
Total	3,192,430	140,374	0.01156	-92.5%	1.1584	+16.6%	26.93	-53.9%	
Diesel-LF(BS-III/IV), baseline	3,192,430	70,000	0.05859		0.4688		38.96		
VS.									
CNG-LF(BS-III/IV)	3,120,181	68,555							
and M2W	38,443	29,572							
M3W	4,890	2,445							
<u>Cars<sup>2</sup></u>	<u>28,916</u>	<u>10,510</u>							
Total	3,192,430	111,082	0.00469	-92.0%	0.5561	+18.6%	29.89	-23.3%	

# Table 6.7: Annual vehicles emissions in Delhi, g/pass-km

<sup>1</sup> Emissions on a g/PKMS basis were calculated by estimating total emissions; total emissions were estimated by multiplying VKMS (above) by the emission factors in Tables 6.4 and 6.5 for each vehicle type. Total vehicle emissions were then converted to g/PKMS based on the average passenger-kilometres (PKMS), per-bus, per-year, of approximately 3.2 million PKMS; this is the average PKMS over the 5-year period of 2006-07 to 2010-11 at DTC (CIRT, 2008-2012).
<sup>2</sup> Includes Cars, Jeeps, and Taxies.

#### 6.3 Results and analysis

I present in Table 6.7 (above) total emissions from buses and vehicles, for the three scenarios, and in Table 6.8 and Figure 6.2 (below) the calculated CE ratios for Std. and LF buses. In Table 6.8, each CE ratio is expressed in life-cycle cost (US dollars) to achieve service-life emissions reduction for PM, NO<sub>X</sub> and CO<sub>2</sub>(e), respectively. For each pollutant, CE ratios are compared for CNG-Std.(BS-I/II) versus Diesel-Std.(BS-I/II), CNG-LF(BS-III/IV) vs. Diesel-LF(BS-III/IV), and Diesel-Std.(BS-III/IV) vs. Diesel-Std.(BS-I/II), since some Indian cities might contemplate this substitution. Calculations in Table 6.8 and Figure 6.2 assume there is full conversion of the bus fleet to CNG, and thus, these results are compatible with Scenario 1.

#### 6.3.1 Analysis of emissions results for the three scenarios

Scenario 1 From Table 6.7, Scenario 1, we see that the change in bus emissions due to CNG implementation are -93% for PM, -57% for CO<sub>2</sub>(e), and +20% for NO<sub>X</sub>, relative to Diesel-Std.(BS-I/II) bus emissions, respectively. For CNG-LF (vs. Diesel-LF), results show PM emissions reduction is still guite significant at 93%, while NO<sub>x</sub> emissions are up 20%. In terms of CO<sub>2</sub>(e), while CNG-LFs lowered emissions by 26%, compared to Diesel-LFs, this reduction is only half the emissions reductions benefit seen with CNG-Std. (vs. Diesel-Std.). Actually, CO<sub>2</sub>(e) emissions from CNG-LF buses are higher compared to emissions from CNG-Std. buses (Table 6.4) since there is a significant impact of the more powerful engines and low-floor bus technology, including the use of AC systems, on fuel economy performance, and thus CO<sub>2</sub> emissions performance. These CO<sub>2</sub>(e) results are despite BC emissions from CNG-Std. buses being nearly 3x higher than for CNG-LFs, and CH<sub>4</sub> emissions being 2x higher (Table 6.4). Lastly, emissions reductions for Diesel-Std.(BS-III/IV) (vs. Diesel.BS-I/II) are also substantial, including for NOx. However, if we compare emissions of CNG-Std.(BS-I/II) to Diesel-Std.(BS-III/IV), which is not shown in Table 6.7, PM would be

down by 80%, NO<sub>X</sub> +200%, and CO<sub>2</sub>(e) -26% for the CNG buses, so CNG implementation is better than using cleaner diesel (BS-III/IV) considering PM emissions reduction goal (Table 6.7).

Scenarios 2 and 3 From Table 6.7, we see in Scenarios 2 and 3 that PM, NO<sub>X</sub> and CO<sub>2</sub>(e) emissions reductions percentages from CNG implementation, relative to the Diesel (BS-I/II) baseline buses, are similar to the results in Scenario 1. That is, substantial reductions for PM emissions for all CNG technologies, increase in NO<sub>X</sub> emissions, and moderate reductions in CO<sub>2</sub>(e) emissions (all relative to CNG buses corresponding baseline diesel counterparts). Moreover, these results show that the emissions reductions benefits due to a full substitution of diesel by CNG buses in Scenario 1 weakened relative to baseline diesel buses, but only very marginally, as higher costs limited the number of CNG buses in operation in Scenarios 2 and 3. Particularly, PM emissions reductions weakened more in Scenario 2, as opposed to Scenario 3, due to the use of a small number of diesel buses that remained in operation in Scenario 2, and the much higher PM emissions profile of these diesel buses. Meanwhile, in terms of CO<sub>2</sub>(e), it was Scenario 3 where emissions reductions weakened slightly more, relative to Scenario 1, than Scenario 2 (vs. Scenario 1), due to the poorer CO<sub>2</sub>(e) emissions performance of lower occupancy vehicles, on a PKMS basis.

However, when comparing emissions results of CNG buses in Scenarios 2 and 3 to CNG emissions in Scenario 1 in Table 6.7, some interesting results also emerge. For instance, PM emissions for a mixed fleet of (mostly) CNG-Std., but also some Diesel-Std. buses, in Scenario 2 is 48% higher than PM emissions from only CNG-Std. buses in Scenario 1. For a mixed fleet of (mostly) CNG-LF, but also some Diesel-LF buses, in Scenario 2 (vs. only CNG-LF buses in Scenario 1), PM emissions are 28% higher. Meanwhile, PM emissions for a mixed fleet of CNG buses in Scenario 3, that includes lower occupancy vehicles, are closer to Scenario 1 results, being +8% for CNG-Std. buses and +16% for CNG-LFs, versus only CNG-Std. and CNG-LF buses in Scenario 1, respectively. Since PM emissions from CNG implementation in Scenario 2 are higher than emissions from CNG in Scenario 3, from a PM emissions perspective, removing as

many diesel buses is likely better, even if lower occupancy vehicles make up for a potential shortfall in CNG bus supply.

These same types of comparisons from the previous paragraph show for NO<sub>x</sub> emissions that across the three CNG implementation scenarios, results are very close to each other. That is, there are only marginal comparable CNG-to-CNG emissions differences between the three scenarios, which vary at the most by 3%. So, from a NO<sub>x</sub> perspective, due to these small CNG emissions differences between the three scenarios, there is likely no significant impact in using a residual number of diesel buses versus using lower private occupancy vehicles to make up for the shortfall in CNG bus transit supply after implementation.

For CO<sub>2</sub>(e), a CNG-to-CNG comparison across the three scenarios shows emissions generally close to each other. There is a marginal increase of about 4.7% in CO<sub>2</sub>(e) emissions for a mixed fleet of (mostly) CNG-Std., but also some Diesel-Std. buses, in Scenario 2 (vs. only CNG-Std. buses in Scenario 1), and 1% increase for mixed fleet of (mostly) CNG-LF, but also some Diesel-LF buses, in Scenario 2 (vs. only CNG-LF buses in Scenario 1). In Scenario 3, CO<sub>2</sub>(e) emissions went up 7% for a mixed CNG-Std. fleet (vs. only CNG-Std. buses in Scenario 1) and by 3% for a mixed CNG-LF fleet (vs. only CNG-LFs in Scenario 1). This increase in Scenario 3 is due to a small proportion of passengers migrating to lower occupancy vehicles to fill the gap in bus supply with an overall increase in vehicle-kilometres needed to transport the same amount of passengers and for the same commuting distances (Table 6.7). So, for CO<sub>2</sub>(e), the above comparisons show that emissions differences across the 3 scenarios are only marginal, with CO<sub>2</sub>(e) emissions from CNG buses being lower in Scenario 2 than in Scenario 3. So, from a  $CO_2(e)$  emissions perspective, it is likely better to keep diesel buses in operations, rather than allowing for a shortfall in CNG bus supply being met by lower occupancy vehicles.

A last note on the scenario analysis is that I assumed <u>total</u> LCCs as the basis for the budgetary constraint, which in turn limited bus supply. An alternate approach, not

considered in this chapter, is that the adoption of CNG may be limited strictly by <u>capital</u> costs rather than <u>total</u> costs. Capital costs, being up-front costs, have to be financed either through the operator's cash reserves (usually very limited), or through government financing and private financing. If this were the case, we would see greater decrease in bus supply, since operating costs for diesel buses are only marginally different to CNG, while diesel capital costs are much lower than comparable CNG buses (Table 6.2). Thus, the impact of this change on supply of CNG buses would be greater, and the deterioration in emissions performance seen in Scenarios 2 and 3 (Table 6.7) will likely be worse.

# 6.3.2 CE ratio analysis

Cost-effectiveness ratio calculations, based on Equation 1 (Section 6.2.2), are presented in detail in Table 6.8, which also includes a breakdown of CE ratio calculations, given incremental LCCs and service-life emissions differences. A summary of the CE ratios from Table 6.8 is also presented in Figure 6.2.

Interestingly, whereas Table 6.7, Scenario 1, showed lower PM and CO<sub>2</sub>(e) emissions for CNG buses, than for Diesel buses, CE ratio results show that CNG buses may not be the optimal choice (Table 6.8). In terms of PM emissions effectiveness, Diesel-Std.(BS-III/IV) (relative to Diesel-Std. BS-I/II) has the largest emissions reductions at the lowest incremental cost, and therefore, the lowest CE ratio, at 3.7 \$/kg, lower than for CNG-Std. buses (relative to Diesel-Std. BS-I/II), at 4.8 \$/kg. So, it appears that, from a CE perspective, Diesel-Std.BS-III/IV would likely have been a better choice for tackling PM emissions reductions at a lower cost, rather than having implemented CNG-Std. buses. The breakdown data in Table 6.8, shows that while bus service-life emissions reductions of PM for CNG-Std. (vs. Diesel Std.) of 5,473 kg is greater than the emissions reductions for Diesel-Std.BS-III/IV (vs. Diesel-Std.BS-I/II), of 3,803 kg, the incremental LCCs for CNG-Std. (vs. Diesel-Std.) are 2x the incremental LCCs for Diesel-Std.BS-I/II), and so the lower CE ratio in the latter case.

(1) Incremental life-cycle costs:	\$/bus					
CNG-Std.BS-I/II (-) Diesel-Std. BS-I/II	26,067					
CNG-LF BS-III/IV (-) Diesel-LF BS-III/IV	23,608					·>
CNG-LF BS-III/IV (-) Diesel-Std. BS-III/IIV	121,569					
Diesel-Std. BS-III/IV (-) Diesel-Std. BS-I/II	13,942					
(2) Bus service-life emissions differences:	РМ	NOx	CO <sub>2</sub> (e)	CH₄	CO <sub>2</sub>	вс
	(kg)	(kg)	(tonnes)	(kg)	(tonnes)	(kg)
Diesel-Std. BS-I/II (-) CNG-Std.BS-I/II	5,473	-7,610	1,274	-4,119	-166	2,603
Diesel-LF BS-III/IV (-) CNG-LF BS-III/IV	2,090	-3,595	383	-2,425	-143	994
Diesel-Std. BS-III/IIV (-) CNG-LF BS-III/IV	1,924	-6,335	191	-2,425	-290	917
Diesel-Std. BS-I/II (-) Diesel-Std. BS-III/IV	3,803	22,832	940	780	-128	1,750
(3) CE RATIO = [ (1) ÷ (2) ]	PM	NOx	CO <sub>2</sub> (e)	CH₄	CO <sub>2</sub>	вс
	(\$/kg)	(\$/kg)	(\$/tonne)	(\$/kg)	(\$/tonne)	(\$/kg)
CNG-Std. (BS-I/II) vs. Diesel-Std. (BS-I/II)	4.8	-3.4	20	-6.3	-157	10.0
CNG-LF (BS-III/IV) vs. Diesel-LF (BS-III/IV)	11.3	-6.6	62	-9.7	-165	23.8
CNG-LF (BS-III/IV) vs. Diesel-Std. (BS-III/IIV)	63.2	-19.2	638	-50.1	-420	132.5
Diesel-Std.(BS-III/IV) vs. Diesel-Std.(BS-I/II)	3.7	0.6	15	17.9	-109	8.0

## Table 6.8: Cost-effectiveness ratio calculation breakdown, Scenario 1

In terms of the low-floor buses, despite lower PM emissions from CNG-LFs relative to Diesel-LFs, as show in Table 6.7, the CE ratio of 11.3 \$/kg in Table 6.8 for this pollutant is more than double compared to the CE ratios for CNG-Std. (vs. Diesel-Std.) and for Diesel-Std.BS-III/IV (vs. Diesel-Std.BS-I/II) previously discussed. The reason for this substantially higher CE ratio for PM emissions for CNG-LFs (vs. Diesel-LFs) can be seen in Table 6.8. Despite a lower incremental LCC of CNG-LFs (vs. Diesel-LFs), compared to LCC difference for CNG-Std. (vs. Diesel-Std.), the service-life emissions reductions of 2,090 kg for the CNG-LF buses (vs. Diesel-LFs) are not as substantial as the reductions reported for CNG-Std.(BS-I/II) (vs. Diesel-Std.BS-I/II) of 5,473 kg; and so, the CE ratio is considerably higher. Also, when CNG-LF (BS-III/IV) is compared to Diesel-Std. (BS-III/IV), the PM emissions CE ratio of 63.2 \$/kg is higher than all the other bus comparisons already discussed; this is mostly due to the high incremental LCC of CNG-LFs over Diesel-Std. buses, as show in in Table 6.8.

This last bus comparison, of CNG-LF (BS-III/IV) vs. Diesel-Std. (BS-III/IV), is important as it shows that implementation of low-floor CNG buses as a substitute for conventional standard diesel buses, which, as noted earlier, some Indian cities might contemplate, comes at substantially higher costs but no significant marginal reductions in emissions. Thus, the use of low-floor buses in the Indian context needs to be carefully considered, since its total life-cycle costs are significantly higher than for standard buses, and since this vehicle technology is increasingly being used across the country (MoUD, 2012). This raises a dilemma for policymakers and planners, since the choice of low-floor buses is based on the assumption that increasing incomes and behavioral changes of transit users are leading them to private modes of transportation, and an increased quality of bus service is needed to counter this effect. However, these benefits (environmental and quality of the bus service) come at significantly higher costs of purchase and operation, all of which reduce the environmental CE of low-floor buses, when compared to less costly standard buses. Furthermore, these higher LCCs most likely limit the capacity of operators to expand their fleets, thus potentially adversely affecting the capacity of bus service provision, which results in possibly higher loading, lower frequency of service, and thus, uncertain outcomes in terms of quality of service desired, and likely lower ridership, particularly among choice bus transit riders.

The NO<sub>x</sub> CE ratio reported in Figure 6.2a are negative for all CNG-to-diesel comparisons. The reason is that, while the numerator in equation 1 is positive, as expected, since CNG bus LCCs are higher than the diesel buses they are being compared to (Table 6.8), emissions are also higher for all CNG bus types relative to diesel buses (and thus a negative denominator in equation 1). So, as far as NO<sub>x</sub> emissions is concerned, CNG is not a desirable choice. Nonetheless, the NO<sub>x</sub> CE ratio for Diesel-Std.(BS-III/IV) relative to older generation diesel buses (BS-I/II) is positive and close to zero, showing that the cleaner diesel buses are a better choice for reducing NO<sub>x</sub> emissions, that is, the incremental costs of the improved diesel system are low and show good emissions reductions results for this pollutant (Table 6.8).



Figure 6.2: Cost-effectiveness ratio, \$-per-emissions reduction, Scenario 1

For the analysis of  $CO_2(e)$  cost-effectiveness ratios in Figure 6.2b, it is interesting to first look at the breakdown of each individual pollutant influencing its value, that is CH<sub>4</sub>, CO<sub>2</sub> and BC. For CH<sub>4</sub>, all CNG buses have negative CE ratios, since, as already noted, emissions of this pollutant are generally expected to be higher, and CNG is mostly methane and especially when considering the fuel leakage problem. For CO<sub>2</sub>, also, all the CE ratios are negative, since all alternative buses, including Diesel-Std.BS-III/IV, emit more than baseline diesel buses (Table 6.8); in this case, the "least worst" choice, is Diesel-Std. (BS-III/IV) as a substitute for Diesel-Std. (BS-I/II). For CNG buses, CO2 emissions are higher due to the fuel economy performance of CNG buses used in Delhi being much poorer than comparable diesel buses (Table 6.4). As discussed in Chapters 3 and 4, CNG's fuel economy penalty relative to the diesel buses they replaced amounted to 45%, both on Std. and LF buses, on an energy equivalent basis rather than the expected fuel economy loss of 20-25% usually reported in the literature. Lastly, for BC, since all CNG buses have substantially lower emissions than the baseline diesel buses (Table 6.4), and since BC emissions are related to PM emissions, the differences in the CE ratios for BC in Figure 6.2b closely follow the differences seen in Figure 6.2a for PM, as discussed above.

In view of the foregoing, in terms of  $CO_2(e)$ , Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II) is the best choice, given the CE ratio of 15 \$/tonne, which is lower than CE ratio of 20 \$/tonne for CNG-Std. (vs. Diesel-Std.), and significantly lower than the ratio for CNG-LF (vs. Diesel-LF) of 62 \$/tonne, and for CNG-LF (vs. Diesel-Std. BS-III/IV) of 638 \$/tonne. The  $CO_2(e)$  CE ratio for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II) is lower than the ratio for CNG-Std. (vs. Diesel-Std. BS-I/II), since the incremental LCC for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II) is much smaller than for CNG-Std. (vs. Diesel-Std. BS-I/II), respectively (Table 6.8); and despite bus service-life emissions reduction of 940 tonnes of  $CO_2(e)$  for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II), being smaller than the emissions reduction for CNG-Std (vs. Diesel-Std. BS-I/II) of 1,274 tonnes (Table 6.8). Meanwhile, from Table 6.8, it can be seen that the CE ratio for  $CO_2(e)$  for both CNG-LF comparisons are much greater than for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II) or for CNG-Std. (vs. Diesel-Std.), since emissions reductions benefits are not as high, and particularly for the CNG-LF (vs. Diesel-Std. BS-III/IV) comparison, the incremental LCC is also much greater than the other comparisons in question, respectively.

**Changes to the discount rate** In Chapter 5, I considered how changes in the discount rate affected the LCC results, so here I include a brief discussion of how a change to this parameter will affect the CE ratio results. Overall, even if the discount rate is progressively lowered to zero, Diesel-Std. BS-III/IV buses (vs. Diesel-Std. BS-I/II) continue to have a lower CE ratio for PM, NO<sub>X</sub> and CO<sub>2</sub>(e), than for all the other CNG bus comparisons, respectively. However, the incremental LCC for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II), as reported in Table 6.8, almost doubles when the discount rate is zero, whereas the increase in LCC differences for the CNG buses comparisons (vs. their diesel counterparts) is not as significant at lower discount rates. So, as the discount rate decreases, the CE ratio advantage for Diesel-Std. BS-III/IV (vs. Diesel-Std. BS-I/II) for all the mentioned pollutants diminishes compared to the ratios of the other CNG bus comparisons (relative to the diesel counterparts), respectively.

## 6.4 Conclusions and implications

In this chapter, I assessed the cost-effectiveness of CNG implementation in Delhi's public bus fleet by comparing emissions of health-critical pollutants and GHGs against the life-cycle costs of buses. This analytic framework is an important tool for evaluating policy and planning decisions, and to my knowledge, this quantification and analysis has not been systematically done for the specific context of Delhi -- or even India -- concerning CNG buses, and therefore, fulfils an important research need.

## 6.4.1 Summary of key results

**Scenario emissions** In Scenario 1, I showed that a full substitution of diesel by CNG buses brings substantial PM emissions reductions, for both the early 2000 CNG buses (relative to comparable diesel buses at the time), as well in terms of newer LF-CNG buses (relative to modern LF-Diesel buses). Lower CO<sub>2</sub>(e) emissions are also possible for all CNG bus types; but, NO<sub>X</sub> emissions were higher for all CNG buses. However, this scenario is not consistent with the reality in Delhi.

Since LCCs are higher for CNG than diesel, which likely constrained bus capacity as discussed, I analyzed Scenario 2 (where some Diesel buses would continue to operate to account for reduced supply of CNG buses) and Scenario 3 (where the operator retires all diesel buses and bus transit capacity effectively is reduced on CNG, which was the case in Delhi). In terms of emissions differences between CNG and corresponding diesel buses in Scenarios 2 and 3, compared to the differences between CNG and diesel in Scenario 1, I showed that emissions for CNG in Scenarios 2 and 3 were only marginally worse than in Scenario 1 for all air pollutants analyzed. However, when I compared emissions changes exclusively between CNG buses across the three scenarios, that is, in CNG-to-CNG comparisons, other interesting results also emerged. Particularly, PM emissions from CNG in Scenario 2 were higher than PM emissions from CNG in Scenario 3, for both Standard and low-floor buses. Meanwhile, in the

CNG-to-CNG comparisons for  $CO_2(e)$ , it showed that  $CO_2(e)$  emissions from CNG buses in Scenario 2 were lower than  $CO_2(e)$  emissions from CNG buses in Scenario 3. These results have important implications, which will be discussed below.

**CE ratios** Whereas the analysis of total emission results (above) favored CNG buses (and particularly Standard buses) over improved diesel buses that complied with stricter emissions standards, the opposite was true in the analysis of CE ratio across all the air pollutants considered. Particularly, in terms of PM emissions effectiveness, Diesel-Std.(BS-III/IV) had the lowest CE ratio relative to all the other CNG bus comparisons, due to lower incremental LCCs. Also, the highest CE ratios for all air pollutants analyzed was for the comparison of CNG-LF (BS-III/IV) to Diesel-Std. (BS-III/IV).

# 6.4.2 Implications

In light of the above results, it is interesting to re-visit the two central questions I initially raised. That is, (i) what were the merits of CNG policy, from a public bus transit perspective based on CNG technology of the early 2000s, and (ii) what is the merit in continuing to pursue this policy in Delhi, or for other cities of India, given current technologies and alternatives?

For the first question, the analysis showed that the CNG implementation policy produced benefits, since CNG-Std. buses offered the most potential in bringing down emissions of health-critical PM and CO<sub>2</sub>(e) pollutants in the early 2000s. This emissions advantage was in relation to comparable diesel buses at the time, and even relative to cleaner diesel buses that would comply in the early 2000s with the subsequent phases of Indian fuel emissions standards (BS-III emission standard was only adopted in Delhi from 2005 and BS-IV from 2010). However, CNG's higher capital and operating costs adversely impacted the CE of emissions reductions in comparison to comparable cleaner diesel buses (BS-III/IV), and the cost of emissions reductions was higher with CNG-Std. than it would have been with these cleaner diesel buses. So, in the early

2000s, the bus fuel and technological choice that would have resulted in the largest emissions reduction of air pollutants affecting human health and climate change inducing GHGs, at the lowest incremental cost, would likely have been to upgrade to a new fleet of diesel buses running on lower sulphur fuel and using improved exhaust aftertreatment systems. This approach would also likely have produced other advantages, such as lower overall implementation costs, a potentially larger bus fleet, and other positive externalities (such as cleaner diesel that would be used nation-wide).

This is not to say that the CNG implementation decision taken in Delhi was ineffective or unjustifiable, since this discussion has to be put into a broader context. In the late 1990s and early 2000s, the air quality in Delhi was in a critically poor state, the institutional realities of upgrading oil refining to meet higher diesel standards perhaps were a major constraint, and the possibility of diesel adulteration was a point of concern for many stakeholders, which could negate any benefits of cleaner diesel technology. Thus, all of these contextual factors may actually justify the higher costs paid for achieving emissions benefits produced by CNG, when these -- and other -- risk factors and institutional realities are accounted for.

Furthermore, the scenario analyses showed that the emissions reductions benefits, from a full substitution of diesel buses by CNG, marginally weakened when higher costs limited the number of CNG buses put into operation. While a full substitution of diesel by CNG buses is still the best option for minimizing PM and  $CO_2(e)$  emissions, the alternate scenarios also produced emissions reductions, which were marginally worse than a full substitution. Also very important is that analysis of Scenarios 2 and 3 showed an important trade-off between urban air pollutant and greenhouse gas emissions reductions. For PM emissions reduction in Delhi, Scenario 3 showed for both standard and low-floor CNG buses, that restricting the use of diesel buses -- despite lower public bus transit supply and the use of lower occupancy vehicles -- was more beneficial than keeping a residual number of diesel buses running, since PM emissions from CNG buses in Scenario 2 were higher than in Scenario 3. However, the opposite was true in terms of  $CO_2(e)$  emissions reductions, that is, when public bus transit supply is

constrained, keeping a residual number of diesel buses on the road minimizes  $CO_2(e)$  emissions (as shown by Scenario 2 results) than allowing transit ridership to migrate towards lower occupancy vehicles (as shown by Scenario 3 results). So, from the point of view of local health critical pollution (i.e., PM), but not in the case of  $CO_2(e)$  emissions, the restriction of only allowing CNG buses on the road, as was required by the Supreme Court of India CNG directive, was an effective choice, rather than allowing diesel buses to continue to operate in the city, and despite increase in traffic of lower occupancy motor vehicles. Of course, the Supreme Court directive was not motivated, at the time, by GHG or climate change concerns, but by the air pollution problem.

In terms of the second question, the analysis showed that, under the BS-IV fuel emissions standard in India, low-floor CNG buses continue to offer substantial emissions reductions in terms of PM and CO<sub>2</sub>(e) when compared to low-floor diesel buses. However, CNG-LFs cost-effectiveness ratios, relative to Diesel-LFs, are much higher than the ratios for CNG-Std. relative to Diesel-Std. since the emissions gap between the CNG and diesel buses narrowed, which shows the increased incremental cost of CNG implementation today (and the well known fact that, the higher the level of pollution control, the higher the marginal cost of pollution abatement). The difference is even more significant when CNG-LF buses are compared to Diesel-Std. buses of similar emissions and fuel technology (BS-III/IV). The problem is that the low-floor buses implemented in Delhi towards 2007/08 had much higher LCCs than the prior generation of standard CNG buses. These substantial cost differences between modern low-floor and standard buses still exist today for both CNG and diesel buses.

Looking at the future of CNG in Delhi, the challenge is that the city has an immense stake in this fuel, which is to say that changing the fuel system back to diesel would come at higher costs than in other Indian cities that already operate diesel bus fleets. In the case of Delhi and its surroundings, all these facts justify considering more carefully the use of CNG-Std. buses and low-sulphur diesel buses with advanced exhaust aftertreatment systems, given their lower costs, relative to CNG-LFs. The implementation of CNG-Std. buses would have lower impact on funding than the more

costly low-floor variants, allowing for a more significant expansion of the public bus transit fleet, thus enabling greater environmental benefits due to CNG use, while also being less climate damaging. For other cities that have fleets currently running on diesel fuel, the analysis shows that the higher costs of CNG do not justify the environmental benefits compared to cleaner diesel bus technologies, unless, as stated above, other contextual factors may justify its use.

# **Chapter 7: Conclusion and policy implications**

# 7.1 Introduction

In this last chapter, I present an overview of the key findings from my dissertation research and discuss the implications of these findings from a broader policy perspective, especially given the significance of my research for CNG public bus transit policy-making in India and for similar contexts. Finally, I outline important issues that were raised throughout the dissertation and how these issues can be further investigated and analyzed, helping guide possible future research on fuel and transportation policy.

As I showed in the first chapters, the critical problem motivating my research is the substantial contribution of the transport sector to air pollution. Remarkably, and despite various interventions, the contribution of road transport to air pollution continues to grow in fast expanding economies like India and China. As result, exposure to air pollution in India and China have led to a substantial rise in mortality and morbidity of growing urban populations. As I have shown, heavy-duty diesel vehicles, disproportionately contribute to this problem, and therefore merit particular policy attention to address urban air pollution.

In terms of road transport, there have been various approaches to address the urban air pollution problem, including widespread implementation of stricter fuel and emissions standards, and the use of alternative fuel and propulsion technologies in buses. Worldwide, diesel-fueled buses continue to be the predominant choice in public transit systems, given their superior performance in terms of reliability and fuel-efficiency. Today, diesel buses comply with stricter emissions standards and emit considerably less health critical pollutants compared to buses used 10 to 15 years ago. For example, United States (US) and European Union (EU) standards in 2016 limit emissions of NO<sub>x</sub> and PM<sub>2.5</sub> to less than a tenth of the levels allowed 15 years earlier (Posada et al., 2016). However, to achieve such results, buses must use lower sulphur diesel, exhaust

gas recirculation systems, and exhaust aftertreatment systems<sup>38</sup>. The use of alternative fuel and propulsion technologies is varied, with countries like the US, already having substantial experience and sizable number of bus operators using CNG and diesel-hybrid buses, and with increasing interest in other alternatives, such as battery electric, fuel-cell, and hydrogen buses.

The challenge for a country like India is that, only in 2017 were EURO-IV equivalent standards implemented countrywide, and so, emission standards still lag those seen in the US or the EU. In such contexts, the use of CNG buses can still substantially reduce emissions of harmful pollutants like PM, relative to existing diesel buses, without need in CNG buses for more advanced fuel, engine and aftertreatment technologies.

Given the above, the overall objective of my PhD research was to assess the operational and financial performance consequences of, and the cost-effectiveness of emissions reductions due to, replacing diesel bus fleets with CNG in Delhi, India. The broader goal of this work is to support decision- and policy-making related to the use of alternative transport fuel systems in low- and middle-income countries, like India, and help inform whether this alternative transport fuel system can cost-effectively mitigate urban air quality problems. For this reason, in my first analytical chapter (Chapter 3) I conducted a comprehensive critical review of CNG in urban bus transit fleets based on the experiences of fleet operators in the US and Latin America. Drawing on the key lessons from this review for CNG evaluation and the important issues raised, I turned my attention to India, by analyzing in Chapter 4 the operational and financial performance CNG buses in Delhi. Both these research projects, in turn, helped provide the basis for modelling the LCCs of CNG buses in the Indian context (Chapter 5). Lastly, based on these LCCs and coupled with my estimation of the emissions outcomes of CNG implementation in Delhi, I evaluated of the cost-effectiveness of reducing key pollutants in this city (Chapter 6). Below I present key results, followed by

<sup>&</sup>lt;sup>38</sup> Such as Diesel Oxidation Catalysts (DOC), Diesel Particulate Filters (DPF) and Selective Catalytic Reduction (SCR) catalysts.

a brief policy relevant discussion, in terms of my dissertation research done in Chapter's 3 to 6.

# 7.2 Key research findings

**Chapter 3** The objective of Chapter 3 was to investigate key lessons from the experiences of five urban bus transit fleet operators in the US, Mexico and Chile with CNG and to explore the relevance of those lessons for addressing urban air quality concerns. It is important to stress that these experiences are relevant to the Indian context in 2000 and beyond, given the bus technologies that were evaluated and emission norms present in India since 2000. To this end, the chapter comprised of a critical review of the performance of CNG based on these experiences to see how fuel implementation can be more effectively evaluated with respect to a bus operator's financial, emissions, and operational goals, and what is required for this implementation to be successful. Key lessons drawn from the analysis of the five cases were:

1) Despite other important benefits for CNG, such as improving energy security and lowering GHG emissions, the primary rationale for fleet operators in considering natural gas implementation were the desire curb air pollution and to comply with more stringent emission standards.

2) CNG buses required simpler and less expensive control measures to achieve more stringent emissions standards than diesel, but at higher purchase costs.

3) Fuel economy (FE) of CNG buses was generally 25% lower, on an energy equivalent basis, relative to diesel in city operations. Interestingly, bus operations in higher average speeds improved performance for both diesel and CNG buses, but much more for CNG. Also important was the fact that CNG's FE penalty disappeared when exhaust control measures like DPF and EGR were used to reduce PM and NO<sub>x</sub> emissions on diesel buses.
4) CNG buses have significant potential for reducing PM and NO<sub>x</sub> emissions, relative to diesel, but that exhaust control technology choice, maintenance practices, and operating conditions on CNG were key in achieving favorable results. For diesel buses, PM emissions were not strongly correlated with diesel sulphur content. Retrofitting older buses with advanced emission controls engines can cost-effectively achieve significant emission reductions of key pollutants, such as PM, with a performance that is comparable to CNG and newer diesel buses.

5) Most of the variations in total operating cost across fuel systems was primarily due to changes in fuel cost, since maintenance cost varied considerably less; in turn, the variation in fuel costs was due to differences in fuel economy -- where CNG performs poorly -- and fuel unit price (which depended greatly on context; even between cities within the US). Furthermore, analysis showed that, from a financial perspective, a life-cycle cost modeling approach to account for the total cost of ownership of buses would be more appropriate to compare various fuel and bus systems, given the differences in operational performance and considering the different composition of capital and operating costs over time.

The critical problem raised in this chapter is the need to incorporate operational and financial issues that are important to vehicle operators in the broader set of transport and environmental policy goals in policy evaluation. It is important and useful to critically evaluate CNG policy from the perspective of vehicle users and operators, because it is the policy responses of these actors that crucially determine the extent to which implementation, and the associated emissions reductions, actually occur and are successful. More particularly, analyzing the operational and financial performance of bus fleets on CNG or any other alternative fuel is important because it is this performance that critically determines the bus operator's policy responses. Chapter 3 showed that for CNG, environmental benefits can be attained at competitive incremental costs in comparison to diesel and diesel-hybrids, making this fuel an important option in controlling emissions in bus fleets without the need for additional exhaust control measures or improved fuel quality, which is very important in contexts such as India. At

the same time, the challenge for CNG is that "cleaner" diesel and more advanced diesel engine technologies are making headway in order to comply with stricter emission standards set by environmental authorities, particularly by using low sulphur diesel in combination with exhaust emissions control measures.

**Chapter 4** The objective of Chapter 4 was to critically evaluate the operational and financial performance of DTC's CNG-fueled bus fleet in order to assess how the conversion from diesel to CNG in Delhi had affected fleet operations and financial performance, and thus addressing the important research need of conducting a post-implementation assessment of CNG. From the analysis of DTC's experience, key lessons were drawn related to the long-term viability of large-scale conversions of bus fleets to CNG, for comparison with other CNG bus transit fleets, and for informing techno-economic and environmental analyses of CNG bus transit operations, as follows:

1) Given various technical, logistical and institutional challenges, the implementation of CNG at DTC showed the importance of co-ordination between various actors, including vehicle manufacturers, fuel suppliers, bus and other public motor vehicle operators, and different levels of government.

2) The initial drastic capacity reduction in bus services at DTC, due to bus scrappage and CNG mandates, demonstrated the importance of more careful planning and of carrying out the implementation in a phased manner.

3) CNG required significant investments in buses at considerable added costs relative to diesel buses, in addition to the investments in depot modifications and fuel infrastructure. Also, operating costs, per kilometre, grew considerably with the use of CNG, firstly due to significantly increased fuel expenses per kilometre -- despite low and declining CNG price -- because of the lower fuel economy of CNG buses; and secondly, due to increased maintenance costs (and breakdowns) per kilometre, despite declining effective kilometres per bus on road. Moreover, both these costs were further

exacerbated by the introduction of CNG low-floor buses. As a whole, these factors severely affected DTC's financial situation for the worse.

The key question raised from this research is how has the significant investment in CNG contributed to, or took away from, the objective of providing convenient, affordable and viable public transit service in Delhi, which is crucially important first and foremost to cater for the accessibility and mobility needs of the masses, and also to minimize the need for personal motor vehicle activity and its associated impacts, including air pollutant emissions. It may be argued that the debt burden, and the adversely affected financial situation of DTC, due to the increased capital and operating costs after CNG implementation, likely detracted from this objective, as well as the ability to enhance public transit capacity and provide widespread coverage region-wide.

**Chapter 5** The objective of Chapter 5 was to evaluate the lifecycle costs (LCCs) of CNG bus configurations in Delhi, and compare them with those associated with their diesel counterparts in the Indian context. The evaluations were based on actual (on-road) bus performance data for DTC and other public bus transit agencies, in order to closely reflect fuel and bus technologies, operating conditions, and costs that are prevalent in urban India. I also analyzed the sensitivity of life-cycle costs relative to fuel economy and fuel price variations to show how these key factors affect the viability of CNG bus systems relative to their diesel counterparts. Key lessons drawn from these analyses were:

1) The LCCs for CNG are higher than for diesel, but CNG negatively affects the LCC of Standard buses proportionately more than for the low-floor buses, for which the LCC is already high. The LCC is significantly higher for the low-floor air-conditioned, relative to that of the low-floor, and in particular the Standard, CNG buses.

2) The significantly higher LCC of the CNG-LF/AC relative to CNG-LF and in particular the CNG-Std. buses is driven by higher bus purchase, followed by fuel and maintenance costs.

3) The sensitivity analysis showed the significant effect of the fuel price (and fuel economy) of CNG relative to diesel, on the competitiveness of CNG relative to diesel. The CNG-to-diesel price ratio, a key factor in this regard, has varied widely, owing to market forces and government policy in India. Both diesel and CNG prices, and the CNG-to-diesel price ratio, are lower in India relative to the US, where these prices appear to be more market-driven.

Overall, the analysis showed the critical importance of the fuel price and fuel economy of CNG, for the competitiveness of CNG relative to diesel buses. This, along with the wide variation in CNG and diesel prices, demonstrates the need for careful fuel pricing policies when CNG is implemented in bus transit. But also, the need for CNG price to properly reflect the costs of production and delivery. A key question for policy that this research raises is whether the significantly higher LCCs for low-floor and low-floor airconditioned buses, even for diesel, are justified by increased patronage, especially by those who might otherwise use personal motor vehicles. As well, what are the opportunity costs of the low-floor and AC buses, in terms of transit supply under budget constraints, and providing affordable transit service to commuters?

**Chapter 6** The objective of Chapter 6 was to evaluate the costeffectiveness (CE) of CNG implementation on past and current bus technologies relative to diesel, by comparing the related LCCs versus their respective emissions outcomes. Overall, I addressed two major questions. Firstly, I investigated the CE of CNG implementation in the public bus fleet in Delhi, by evaluating the incremental costs and emissions performance of buses that were used in the conversion to CNG in the early 2000s in that city. Secondly, for Delhi, but also more broadly the Indian context, I analyzed CNG implementation in a public bus fleet today considering that CNG and diesel buses must comply with stricter Indian emissions standards. Costs were based on LCCs from Chapter 5, while emissions were evaluated in terms of PM, NOx, and CO<sub>2</sub>(e). Key lessons drawn from the analyses were: 1) A fully implemented fleet of CNG-Std. buses in the early 2000s would minimize emissions of PM and CO<sub>2</sub>(e) pollutants relative to existing diesel buses and even in relation to cleaner diesel technologies available at the time. However, the choice that resulted in the largest emissions reduction, at the lowest cost, in tackling these emissions challenges would likely have been to upgrade to a new fleet of diesel buses running on lower sulphur fuel and using improved exhaust aftertreatment systems, given the higher incremental LCCs of CNG-Std. buses. Modern low-floor CNG buses also offer considerable reductions benefits of PM and CO<sub>2</sub>(e) emissions relative to comparable low-floor diesel buses; however, CNG-LF's costs of lowering these emissions relative to low-floor diesel buses is much higher than for CNG-Std. relative to comparable Diesel-Std. buses, and substantially more so if CNG-LF buses are compared to Diesel-Std. buses of similar emissions and fuel technology, thus immensely increasing its cost-effectiveness ratio relative to all other options.

2) The results of the scenario analysis showed that, in Scenario 1, with a full substitution of diesel buses by CNG, which would only happen if there is no budget constraint, there would be substantial PM reductions, lower CO<sub>2</sub>(e) emissions, but higher NO<sub>X</sub> emissions, for all CNG bus types. However, this scenario is not consistent with the reality in Delhi. Since LCCs are higher for CNG than diesel, which likely constrained bus capacity, I analyzed Scenario 2 (where some diesel buses would continue to operate to account for reduced supply of CNG buses) and Scenario 3 (where the operator retires all diesel buses and bus transit capacity effectively is reduced on CNG, which was the case in Delhi). The analysis of these two scenarios shows an important trade-off between urban air pollutant and greenhouse gas emissions. Whereas PM emissions were lower in Scenario 3, CO<sub>2</sub>(e) emissions were lower in Scenario 2, for CNG vs. Diesel, on both Std. and LF buses.

So, in relation to the policy decision that required CNG implementation in Delhi, I showed that CNG was not necessarily the most effective choice in the early 2000s, as cleaner diesel buses would have a lower CE ratio, and thus resulted in the largest emissions reductions, at a lower incremental cost, in addressing both health-critical (PM

and NO<sub>x</sub>) and climate change (CO<sub>2</sub>(e)) challenges. However, if we consider that CNG implementation was a given, following the Supreme Court of India decision, then, the decision of restricting buses to use only CNG was an optimal choice from the point-of-view of health-critical air pollution (PM emissions), and despite the extent to which a lower supply of CNG buses may have increased traffic of lower occupancy motor vehicles. From the point-of-view of climate change objectives, I showed that keeping a residual number of diesel buses would have likely been a better choice, but then again, the primary motivation of the policy instituted in early 2000s was to address the air pollution challenge (specifically, PM emissions) and not climate change. Furthermore, as I argued in Chapter 6, this policy decision should also be analyzed from a broader context of air quality that was critically poor at the time in Delhi, the institutional challenges of upgrading oil refining capacity, and concerns around the enforcement of actually using cleaner diesel in buses. The CNG policy decision, when taken with these contextual factors into consideration, may actually justify the higher costs that were incurred for achieving emissions benefits.

### 7.3 Policy implications of research

In the early 2000s, after CNG started being implemented in Delhi, PM emissions decreased due to significant reductions coming from the public bus transit system (Guttikunda, 2012). However, these emissions benefits were accompanied by major disruptions to bus supply during the 2-year transition period from diesel to CNG. Further, the disruptions had serious adverse consequences to bus transit users at the time (Bell et al., 2004), least of which were frustrated commuters due to the lack of buses and extremely high load factors, and related disruption of economic activity due to this lack of public transit availability, all of which was not quantified in my research. As discussed, the problem was that when the natural gas policy came into effect, there were various resource, logistical and institutional challenges encountered by DTC in establishing a CNG infrastructure for its bus fleet operations and in the initial transition to CNG that prevented the technology from expanding. In this sense, a sensible policy approach at the time in Delhi would have been a gradual implementation of CNG, with

diesel buses still operating during the transition to the new fuel system, as stated in the conclusion of Chapter 4.

Even after CNG bus numbers in Delhi increased, the emissions benefits of CNG were diluted, as shown in Chapter 6, since bus service provision capacity, on a perpopulation basis, never reached the levels seen before CNG. The problem, as I showed in Chapter 4, was that the added costs of CNG negatively affected DTC's financial situation over time, and this continues to reflect in the operator's extremely high debt service load and low fare recovery rates even today<sup>39</sup>. In fact, I analyzed in Chapter 5 the incremental life-cycle costs of CNG over diesel, for which the key contributor to these differences was higher fuel expenditures for CNG; and fuel expenditures were higher, despite a low CNG fuel price, mostly due to the significantly poorer fuel economy of CNG buses in Delhi, which was almost double the worst expected performance result seen in the literature for like-to-like buses, as shown in Chapter 3. Furthermore, in capital expenditures terms, the added costs of CNG for a standard busbased fleet is 24% higher, while for LF/AC buses is +10%, relative to their corresponding diesel buses.

So, given the above considerations, a crucial question that my research raises is, apart from its emissions outcomes, how has the significant investment in CNG contributed to, or taken away from, the objective of providing convenient, affordable and viable public transit service in Delhi, which is, as argued earlier, crucially important. This question gains particular importance because of the decline in passenger-kilometres at DTC, and the equally significant reduction in public transit modal shares in Delhi among other metropolitan cities in India, over the past couple of decades.

It may be argued that the debt burden, and the overall financial situation due to the increased capital and operating costs per passenger-kilometre due to CNG

<sup>&</sup>lt;sup>39</sup> The problem is persistent, as shown by the latest financial report for DTC -- for the 2014-15 period -- where interest payments corresponded to 73% of all expenses (less depreciation), and operating expenses equaled to 121% of all traffic receipts (CIRT, 2017).

implementation has in fact detracted from this objective, as well as the ability to enhance public transit capacity and provide widespread coverage region-wide. After all, if the same investment had been made in regular diesel buses, a larger capacity to deliver bus transit services might have resulted at lower cost, thereby better achieving equity, and as I showed in Chapter 6, better environmental results, by helping avoid a larger number of personal motor vehicle trips. An additional problem in this regard is of course that the investments are being made without any accompanying transit demand management measures.

All of the foregoing points have important implications for policy-making and implementation, as well as for policy analysis. They demonstrate the need to analyse and formulate environmental policies such as the implementation of CNG on Delhi's buses broadly, in terms of a wide range of impacts for different groups in society, rather than narrowly in terms of only environmental (in this case emissions) outcomes, and to explicitly consider and address conflicts and trade-offs between environmental, and other (transit operation, socio-economic and equity) objectives.

### 7.3.1 Other issues

**Parameter estimation** Another important issue that was raised relates to the choices for performance data collection methods and the high degree of variability of operational measures that resulted from different choices in performance evaluation methodology. In Chapter 3, the case of WMATA showed significantly different fuel economy performance results between chassis dynamometer tests and on-road data for an identical sample of buses, being much more favorable towards CNG buses in the dynamometer tests than in the on-road results. The chassis dynamometer tests were conducted using driving test cycle specifically designed to mimic road conditions of WMATA's fleet, but showed a fuel economy penalty for CNG buses, vis-à-vis diesel, being nine times smaller as recorded by on-road data collected during a 12-month evaluation cycle. The differences in this example are striking, and have significant implications for analysis, given that fuel economy data typically feeds into

multiple dimensions of various analytic frameworks, such as in assessing vehicle range (fuel autonomy) requirements given operating efficiency goals, in estimating fuel expenditures in terms of financial feasibility, or even in terms of evaluating environmental outcomes, such as with CO<sub>2</sub> emissions estimation.

I do not want to argue on the merits of chassis dynamometer tests versus on-road tests. But, in WMATA's example in Chapter 3, the longer-term on-road statistics provided a more robust sample of performance estimates for use in the analytic framework (as did on-road data from DTC's CNG buses), and also illustrated how limited sampling (be it via chassis dynamometer or other methods) adds much uncertainty to the analytic and decision-making process. The implication of this for policymaking is that feasibility studies should ideally account for uncertainty in the use of parameter assumptions, whenever data limitation is present, such as by use of simulation or stochastic methods, by considering expected variations of these input parameters.

For this dissertation, I collected data in the Indian context based on long-term and actual on-road conditions, in order to increase the validity of analysis findings. Even so, I incorporated uncertainty in parameter estimation into the analytic framework via the use of a sensitivity analysis in Chapter 5 and the scenario analysis in Chapter 6. With this framework, key stakeholders and decision-makers can evaluate choices by incorporating this uncertainty into the planning process, to set goals and objectives, realistically budget resource needs and financial requirements, and set expectations for operators, users, or other stakeholders.

**Reliability measures** In Chapter 4, I explained how DTC outsourced maintenance of newly procured low-floor buses to vehicle manufacturers, for the entire duration of the expected vehicle service life of 12 years. Interestingly, reliability performance of the newly purchased low-floor buses was not significantly different from the ageing fleet of standard CNG buses that still operated at DTC when data was gathered; actually, low-floor bus reliability was much worse during a certain period. This fact raises the issue of why new CNG-LF buses were apparently less reliable than much

older CNG-Std. buses. While the new low-floor CNG technology could have been influenced by short-term teething problems, it also cannot be dismissed that the choice of performance metric to hold outsourced firms accountable to contractual maintenance quality standards may have influenced the overall reliability performance of CNG technology. In DTC's case, rather than holding low-floor bus maintenance providers accountable to a reliability measure, such as number of breakdowns-per-kilometre, vehicle manufactures were only accountable to meet a bus availability target<sup>40</sup>. So, it should be quite obvious that, if key performance measures related to both reliability and availability are not integrated into pricing mechanism of maintenance contracts (e.g., in the form of penalties in payment transfers to providers for not meeting breakdown targets), there can be an incentive for maintenance providers in having buses starting its trips, even though buses may be at high risk of not completing the journey due to mechanical or other technology related problems, since it would increase the availability percentage. Whether or not this was a motivating factor on the part of maintenance providers driving the poor reliability performance statistics of new low-floor CNG buses at DTC, the fact remains that there is intense public attention on the issue (negative at best), as cited in previous chapters. This also shows the critical importance of carefully selecting performance measures.

From this problem, I raise a few key policy lessons. First, that outsourcing fleet maintenance was a very interesting (and valid) strategy, as it -- theoretically -- allowed the transit operator to transfer the technological performance risk of a new fuel system to the vehicle manufacturer. The problem is that, to effectively transfer this risk, the transit operator must be capable of integrating the other important aspects that influence operational efficiency of its fleet. For example, in Chapter 3, I showed that NYCT has set internal performance metrics for in-house maintained buses in terms of not only availability, but also reliability and recovery time of buses submitted to maintenance. Perhaps, all these operational performance goals should be taken into account in the contractual obligations of maintenance service providers in DTC's case,

<sup>&</sup>lt;sup>40</sup> Availability at DTC, is measured by dividing the completed distance of trips by scheduled distance of trips (Chapter 5).

to fully allow for a more effective transfer (and thus management) of technological operational risks, and thus aim for more reliable fleet operations.

# 7.4 Scope for future research

### 7.4.1 Trade-off between costs and public transit capacity and its impacts

Since the early 2000s when CNG was implemented in Delhi, there has been a significant change in mode shares in the city, with substantial increase in the size and share of low occupancy motor vehicle usage, as I stated in Chapter 6. I showed, through that research, to what extent bus supply constraints and modal shifts reduced the effectiveness of environmental objectives. To do this, I quantified the effects of hypothetical modal shifts on the cost-effectiveness of CNG measures using a scenario analysis. The latest data shows that from 2001 to 2011 the urban part of Delhi experienced an increase of 3.5 million people (+27%) with growing urban density rates (GNCTD, 2017), an environment that favors the feasibility of any public transit system; during this same period, however, public bus transit fleet numbers have remained constant, at best. Meanwhile, I have to recognize the difficulty and uncertainty in attributing causality of the modal changes specifically to reductions in bus capacity; which in turn, I assumed were caused by the operator's capital constraints. Certainly bus supply restrictions are a contributing factor to these modal changes, however, in this respect, I believe there is much room for research in India and countries that face similar challenges.

So, given the above context, I believe that more research is needed in quantifying precisely how (and to what extent) higher bus life-cycle costs may negatively affect the capacity of service provision of transit operators. In addition, a related question is the need to investigate how possible bus supply capacity reductions may impact commuter mode choices and transit demand. Relative to these two questions, there are many aspects that would have to be explored in terms of the decisions that are made regarding bus fleet numbers in Indian cities. For instance, I would include the need of a

deep understanding of the bus operator's rationale and priorities, financial and operational constraints, and how all of this is factored into fleet number and deployment decisions; the political economy of how fleet funding and decision-making is taken across various layers governmental jurisdictions; the socio-economic context and other institutional factors that may affect decision-making; and, the motivations and influences of various stakeholders in this process. For both these research questions, the case of Delhi, and perhaps even of other large Indian cities that have exhibited similar trends of a public bus transit system that has not kept pace with population growth, will be a fertile ground to consider in an integrated manner policy-relevant, contextually-grounded research on these complex issues. This proposed line of research would certainly enhance cost-effectiveness analysis framework not only for evaluating alternative transport fuel outcomes, as I did in this dissertation, but also in evaluating other public transit and urban transport objectives, considering an integrated transit perspective.

# 7.4.2 CNG feasibility

For context, consider that Delhi, Mumbai, Kolkata, Chennai and Bangalore in total benefited from 41% of buses funded by the Indian national program called Jawaharlal Nehru National Urban Renewal Mission, JNNURM (MoUD, 2012), but that these cities, combined, represented only about 16% of India's urban population. This may suggest that larger bus operators, in larger metropolitan cities of India, have access to more financial resources, and are thus able to count on technological options, such as CNG, that would otherwise not be available to smaller operators in smaller Indian cities that face greater budget constraints. In fact, CNG had its use expanded beyond Delhi to other large cities as well; from 2010-11 to 2014-15, 1210 additional CNG buses (+32%) were used in other urban areas like Mumbai, Pune, and Ahmedabad (CIRT, 2012; CIRT, 2017). Also, there is a considerable domestic installed capacity with logistical support for expanding natural gas use in bus transport in other cities (Government of India, 2014).

So, it is reasonable to expect that a gradual expansion of CNG to greater scale across the country may not encounter the same challenges faced in Delhi. However, these larger cities have established formal public transport systems and thus are more likely to receive funding for bus purchases for fleet renewal or fleet augmentation, whereas these formal public transport systems are not typically present in smaller Indian cities that normally rely on informal para-transit systems (Tiwari, 2011). Given this context, it is very unlikely that smaller Indian cities may leapfrog from higher polluting diesel buses to more costly technologies such as CNG. Furthermore, in terms of infrastructure and logistical capabilities, other cities in India that do not have local CNG installed logistical and feedstock supply networks, as in Delhi. These facts seriously raise the question of whether or not such systems, as CNG, should even be considered in these contexts.

An additional issue related to the preceding challenge is the influence on alternative transport fuel feasibility and viability for buses, vis-à-vis diesel systems, given variations in scale in the size of a bus fleet. In other words, is there an ideal (or minimum) bus fleet size that optimizes the financial efficiency of a CNG-based bus fleet, given technological infrastructure and operations requirements of this fuel system? Furthermore, how would CNG system feasibility be impacted, in India, if adopted across a range of city sizes? My dissertation findings were based mostly on performance data from "megacities" of India, that is, cities with populations greater than 8 million people, where travel distances are longer and there is a substantial critical mass in terms of public transport demand (Tiwari, 2011), as highlighted above, and thus the need for large fleets. However, such may not be the case in smaller cities, where a smaller critical mass exists in terms of public transit demand, and if CNG cannot be scaled down costefficiently to a smaller fleet size, this may complicate the financial and operational feasibility of CNG bus systems. Investigating these issues with in-depth research will allow for a more comprehensive analytic framework; that is capable of handling varying degrees of public transit realities, and that therefore is more effective in evaluating the viability of bus alternative transport fuel systems across these varying contexts.

#### 7.4.3 Fuel supply constraints

I indicated in the dissertation that there were various substantial barriers to implementing CNG in Delhi. A specific problem was the logistical constraint, that is, in terms of natural gas supply network needs (from upstream to distribution to bus depots), especially since this set-up was not in place at first. If this is the case in other contexts, this raises the question of how CNG fuel prices may be affected.

Determining the makeup of fuel prices can be quite complex, as prices are influenced by a variety of factors, such as natural gas upstream and downstream costs, retail costs, market conditions, in addition to which there are public policy influences through taxes and subsidies. For example, in the US, where prices tend to be more market driven, from 2010 to 2013 natural gas feedstock costs, as represented by the average city-gate price (EIA, 2013) represented only about a quarter of the final retail price of CNG (EERE, 2016). Other key cost components of fuels include capital infrastructure for the refuelling stations, energy used in gas compression (which tends to be significant), and fuel retailers' margins, among others. So, natural gas feedstock prices can represent only a small proportion of overall CNG prices, and thus, there might be large variations in CNG costs from city to city, depending on the challenges and costs of establishing a CNG distribution network available to the bus transit operator in India.

In Delhi, there is a well-established natural gas distribution network provided by the local utility company, which is a business that is not dependent exclusively on gas demand from the bus system. According to Delhi's natural gas distributor, piped natural gas sales for domestic, commercial and industrial usage, amounted to 1.3 billion standard cubic meters in 2012-13, representing 25% of all natural gas sales, with the remainder 75% of natural gas being sold in the form of CNG for use as vehicle fuel (IGL, 2017). So, given the complexity of determining a viable CNG bus fleet operation, which in turn is dependent on infrastructure requirements and fuel availability, and a competitive fuel price vis-à-vis diesel, more in-depth review of these issues is warranted, in case adoption is required at varying level of city sizes, in terms of varying

levels of public transit demand, and given possible natural gas feedstock supply constraints.

# 7.4.4 Natural gas supply scalability

Lastly, I would like to raise the importance of the scalability issue relating to CNG use in road transport, that is, of the consequences of expanding natural gas use in transportation in India, in terms of fuel supply availability, resource use challenges, and effects on the price of natural gas and CNG. To put this discussion into perspective, consider that road transport in India uses 10% of total primary energy supplied, which is much smaller than the share of road transport to total primary energy supply observed among higher-income OECD countries (21%) or even the shares observed globally (15%) (IEA, 2018). Assuming road transport in India increases proportionally to resemble the shares of this sector in the rest of the world, this means that energy consumed in road transport will grow at a much faster pace than energy consumed in other sectors of the economy; a trend that the country has witnessed over the past decades, and that will likely continue into the foreseeable future. Furthermore, the share of natural gas in India, compared to other energy sources, represents only 6% of total energy sources (Figure 7.1a); of this 6%, road transport consumes only a small fraction (5%) (Figure 7.1b); in 2016, natural gas consumption in road transport amounted to 2.15 million tonnes of oil equivalent (Mtoe) in India, while oil consumption was 79.6 Mtoe (IEA, 2018).

So, consider the case in which natural gas energy consumption in road transport doubled in India. Using data from IEA (2018) in Figure 7.1b, this case would require an increase in domestic natural gas supply to road transport to 4.15 Mtoe, increasing overall natural gas use by an extra 5%, assuming natural gas use in all other sectors remain the same. A four-fold increase in natural gas consumption in road transport energy to 8.6 Mtoe in India, would require natural gas supply to increase by an extra 14%. The bottom-line from this simple exercise is that any of these scenarios -- however unlikely -- do not represent unreasonable logistical supply challenges for a

country like India, considering current installed natural gas upstream capacity, and assuming a gradual increase in this capacity with more investment and resources; also consider that India's share of global natural gas consumption is only 1.5% (IEA, 2018), and thus, would not be a major global disruptor to demand for this energy feedstock, assuming most of the increase in natural gas consumption is supplied from imports.



Figure 7.1: Natural gas supply and demand, India, 2016

Though India may have the capacity to scale up CNG adoption, given the above discussion, it is also interesting to consider if this increase in natural gas use will lead to changes in fuel costs or even in fuel availability; in my research I showed that changes in these factors have direct consequences for the bus operator's financial and operational viability.

The life-cycle cost analysis and the cost-effectiveness study conducted in Chapters 5 and 6, assumed supply and prices for natural gas and CNG as exist currently, that is, considering road transportation sector natural gas demand that is currently 2 to 3% of overall road transport energy demand in India today (IEA, 2018). While the sensitivity analysis in Chapter 5 did consider variations in price of CNG relative to diesel, these scenarios were not based on an in-depth analysis of a possible expansion of CNG use in transport sector, and the impacts that this change could have of fuel prices.

Furthermore, the dissertation also did not analyze the possible consequences of increasing CNG adoption rates within the road transport sector, and related requirements from upstream to downstream supply chain needs, in terms of feedstock exploration and production (or import), logistics of doing this, feedstock processing (purification), transportation, storage to consumption hubs, as well as local distribution of natural gas by utility companies, and setting up of CNG refuelling infrastructure. This is quite a complex supply system, for which scaling-up would require massive investments, time, and resources. These issues are worth investigating.

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## Appendix A: List of sites visited in Delhi for data collection

Table A1 lists DTC sites visited in 2010 for data collection; the sites at which data was collected were chosen to ensure representativeness of the geographical regions served by DTC's urban transit routes, and accounted for the fact that DTC operated two types of CNG bus technologies in 2010, for which bus attributes varied significantly.

Site Name	Location	DTC Region	Activities
1. Strategic Business Unit	Hauz Khas	South	DTC's management office for the LF bus fleet
2. Central Workshop I	Banda Bahdur Marg	North	planning for the standard buses, and re-conditioning, re- treading and bus body repairs
3. Sukhdev Vihar Bus Depot	Sukhdev Vihar	South	Bus depot (operating the LF)
4. Hari Nagar Depot I	Hari Nagar	West	Bus depot (operating the LF)
5. Hari Nagar Depot II	Hari Nagar	West	Bus depot (operating the Standard)
6. Rohini Depot I	Rohini	North	Bus depot (operating the LF)
7. Kalka Ji Bus Depot	Sanjay Colony	South	Bus depot (operating the Standard)

Table A1: Sites visited for da	a collection at DTC in Delhi, India
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Source: Google Earth, 2015 Figure: Delhi Map with location of DTC sites visited in 2010.

In 2010-11, operation, storage and maintenance of DTC's bus fleet were conducted from 49 depots across Delhi; based on fleet average statistics for this period, a typical DTC depot facility has an area of about 20,000 m<sup>2</sup> with an average capacity of servicing about 110-130 buses (CIRT, 2012). The majority of these depots have dedicated CNG refueling infrastructure on site, which includes diesel-powered compressors, CNG buffer storage tanks, and usually two fast-fill dispensers per depot (Figure A1).



*Source*: Map (Google Earth) and pictures (author).

### Figure A1: DTC bus depot layout, Sukhdev Vihar depot, 2010-11

# Appendix B: Supplementary material for Chapter 5

Model Notation	Parameter	Description
Т	Vehicle service Life	Vehicle service life was assumed to be 12 years for LF buses, and 10 years for standard buses, reflecting the average lifespan of these buses in the Indian context (CIRT, 2012).
k	Annual kilometres per bus	From 2001-02 to 2010-11, the average annual distance operated for CNG buses at DTC was approximately 74,000 kilometers, as opposed to 86,000 kilometers for buses in Mumbai, Chennai and Bangalore (CIRT, 2012). The average annual kilometers, weighted by fleet size, for the above three cities plus Delhi, was approximately 80,000 kilometers in 2010-11, but has been declining for the past 5 years for DTC, BEST and BMTC. Since the average annual kilometers per bus is just under 70,000 kilometers in Delhi and Mumbai, which is likely due to the higher traffic congestion in these cities than in Chennai and Bangalore, as shown by average speeds in WSA (2008a), the LCC model assumed an annual average of 70,000 kilometers operated per bus (k), to account for this possible trend.
Fleet size		In order to compare the various bus systems on an uniform basis, and because of the variable size of the fleets representing these systems in our study, we calculated the LCCs for 100 buses for all of the bus systems in the Indian context, because this is the approximate number of buses serviced in each depot at DTC in 2010-11; 100-bus depots are also typical of US transit operations (Lowell et al., 2007).
d	No. of depots	Because the LCCs were calculated for 100 buses, which is the approximate number of buses per depot at DTC, the number of depots, which is important for determining depot infrastructure and staffing requirements and costs, is 1 in our model.
у	Annual vehicle- kilometres	Calculated by multiplying the number of buses (in the case of our model, 100) by the annual operated kilometers for each bus (k); this measure is assumed to be constant for all years of operation and common to all of the bus systems that were evaluated.
e	l'otal Training Hours per depot	Vector e was estimated by multiplying training hours by staffing numbers for each staff function. Staffing numbers, per depot were assumed to be 540 operators, 110 mechanics, and 51 administrators, for a total staff ratio per bus of approximately 6, based on staffing number breakdowns observed at DTC in 2010-11. For training hours, no specific data was available within the Indian context, so training hours reported in Lowell et al. (2007) were used for CNG and diesel buses; that is, a vector of initial training hours of [3,25,2] for [bus operators, mechanics, administrators], respectively, for CNG, and [2,20,0] for diesel buses. Yearly ongoing refresher training requirements of [1,7,1] hours for CNG buses, and [0,5,0] for diesel buses, for [operators, mechanics, administrators], respectively, were assumed, also based on Lowell et al. (2007).Vector <i>e</i> was estimated by multiplying training hours by staffing numbers for each staff function. Staffing numbers, per depot were assumed to be 540 operators, 110 mechanics, and 51 administrators, for a total staff ratio per bus of approximately 6, based on staffing number breakdowns observed at DTC in 2010-11 (in the US, these numbers were assumed to be 300 operators, 20 mechanics and 30 administrators per depot based on Lowell et al., 2007). For training hours, no specific data was available within the Indian context, so training hours reported in Lowell et al. (2007) were used for CNG and diesel buses; that is, a vector of initial training hours of [3,25,2] for [bus operators, mechanics, administrators], respectively, for CNG, and [2,20,0] for diesel buses. Yearly ongoing refresher training requirements of [1,7,1] hours for CNG buses, and [0,5,0] for diesel buses, for [operators, mechanics, administrators], respectively, were assumed on Lowell et al. (2007)

# Table B1: Description of parameter assumptions used in the LCC model

Model Notation	Parameter	Description
FE	Fuel economy	The fuel economy performance of standard CNG buses at DTC from 2001-02 to 2010-11 was around 45 percent lower than that of DTC's diesel-only fleet, on an energy-equivalent basis, in the 10 years prior to the switch to CNG, and also relative to that of standard diesel buses operated by BEST, BMTC and MTC-CNI during 2001-02 to 2010-11. Our LCC model assumed 3.81 km/L as the fuel economy for standard diesel buses based on this 10-year average performance from CIRT reports for BEST, BMTC and MTC-CNI, and 2.12 km/L as the energy equivalent fuel economy of standard CNG buses, based on the data for DTC during the same 10-year period (CIRT, 2002-2012). Fuel economy for LF/AC CNG buses was assumed to be 1.33 km per liter diesel energy equivalent, based on data collected on a sample of 25 of these buses operating out of the same depot at DTC for the 12-month period from Jan 2009 to Jan.2010. Meanwhile, the fuel economy of diesel LF/AC buses was assumed to 2.14 km/L, based on the average fuel economy performance of "Volvo" type buses used in BEST, BMTC and MTC-CNI, and reported in CIRT reports. The fuel economy for diesel low-floor buses had to be estimated since there was no disaggregated data for this type of bus. Estimation was done using a linear interpolation of LF performance in CNG buses (i.e., between LF and LF/AC CNG buses), and adjusted for diesel performance given LF/AC diesel buses as observed in India. The following formula was used: $FE_{diesel LF} = \frac{FE_{CNG LF}}{FE_{CNG LF/AC}} \cdot FE_{diesel LF/AC}$ The ratios of our estimated fuel economy for low-floor diesel buses to that for the Standard and LF/AC diesel buses are very close to these ratios based on data in the ESMAP (2011) report for Hyderabad and Mysore.
r P	Discount rate Fuel price	The discount rate was based on Zhuang, Liang, Lin, & De Guzman (2007) for the Indian context, as discussed in the paper. The LCC model assumed a CNG price of \$0.480/liter diesel energy equivalent and diesel price of \$0.860/liter, based on the five-year average of CNG and diesel prices reported in Delhi, from 2011 to 2015 (MyPetrolPrice.com, 2016). Delhi was selected for this purpose, because the city has an important and sizable market for both fuels, aside from natural gas demand from DTC. Since fuel prices have fluctuated much over the years, on an inflation-adjusted basis, we used the most recent five-year averages for the CNG and diesel fuel prices for the Indian context, and also for the USA, based on EERE (2016).
W0	Depot infrastructure, operation and maintenance costs	Based on the available data in the Indian context, depot infrastructure cost in the model was set to 2.2 million USD for CNG buses and 1.7 million USD for diesel buses. The depot service life was set to 25 years, based on Lowell et al. (2007) and Clark et al. (2009). Note that the CNG depot infrastructure includes only equipment and building costs related to the operation and maintenance of buses, but excludes the refueling infrastructure, which, as discussed in the paper, was indirectly accounted for via the fuel price. The costs related to operation and maintenance of depot infrastructure (including those pertaining to depot up-keep due to physical depreciation, as well as costs related to utilities, and other depot infrastructure operational costs) are also important to consider. However, since no data was available in relation to the annual depot operating and maintenance costs in the Indian context, we assumed these costs to be 5.5% of total capital expenditures, based on Lowell et al. (2007) and Clark et al. (2009).

Model	Parameter	Description
Notation		
w <sub>1</sub>	Bus purchase cost	CNG bus costs were based on inflation adjusted procurement information collected at DTC. Data for Indian transit operators that purchased both CNG and diesel buses (Government of India, 2009a; 2009b; 2009c; 2011; 2012) shows that, taking all of their bus purchases as a whole, the average cost premium for CNG buses was 7-25%, with lower premiums for the more expensive LF CNG buses, and higher premiums for the less expensive standard CNG buses, relative to their diesel counterparts. These figures are similar to the 8-20% premiums for CNG buses in the USA and Europe (Lowell et al., 2007; Clark et. al., 2009; Posada, 2009). For our LCC calculations in the Indian context, CNG-Std., CNG-LF and CNG-LF/AC bus costs were based on data collected on recent purchases at DTC, and inflation-adjusted. In order to ensure the best possible comparative purchase costs for their diesel counterparts in the Indian context, we used data for other large urban bus operators in India for the Std. buses, and for the Municipal Corporation Ludhiana (2013) for the LF and LF/AC buses, because their diesel buses are of the same type and model as the CNG buses at DTC.
		At the end of the fleet's useful service life, bus costs, w1, are subtracted by a residual value, based on the fact that buses typically operate beyond, and thus have value at the end of, their service life, T. For example, Laver et al. (2007) show that buses in the US operate 15 years on average, even though the target service life is 12 years. Schubert and Fable (2005), estimate the average salvage value of a transit bus in the 12th year of operation to be about 15% of its initial capital cost for diesel buses and 13% for CNG buses, which is what our LCC model assumed, in the absence of specific data in the Indian context. The residual value is treated in the LCC model as a negative cost (i.e., a revenue to the operator) received at year T+1.
W2	Maintenance cost	Maintenance expenditures, which incorporate labor and parts costs, were assumed to vary year-on-year throughout T, incorporating the ageing of vehicles and the need for vehicle and engine overhauls and rebuilds. So, the data reported in Table 3 reflect the average annual costs throughout T for each bus system. A detailed breakdown of maintenance cost assumptions over the service life of each bus system is provided in Figure 1. Data for standard CNG buses in Figure 1, were based on the observed maintenance costs (labor and materials) of DTC's standard CNG buses from 2001-02 to 2010-11, since this 10-year period captures the aging of these buses, from the time they were introduced in 2001-2003. The cost profile for standard diesel buses was estimated based on the average fleet maintenance costs from 2001-02 to 2010-11 for BEST, BMTC and MTC-CNI. For the CNG LF/AC bus costs in Figure 1, data was based on the contractual obligations between DTC and vehicle manufacturers, which includes all maintenance costs (labor and consumable parts) as well as systems overhaul throughout the 12-year targeted service life of vehicles. Diesel LF/AC costs were estimated based on average maintenance costs observed in the past 12 years for standard buses in India for BEST, BMTC and MTC-CNI, but adding an average maintenance cost premium of 20%, which is the added cost observed in DTC for LF/AC CNG buses over standard CNG buses.

Model Notation	Parameter	Description
W3	Fuel cost	Fuel expenditures were assumed to be constant throughout the forecast period $T$ and calculated as a function of fuel price ( $P$ ) and fuel economy ( $FE$ ).
W4	Labour cost	Operator labor costs, measured in \$-per-kilometer for various bus fleet types. Operator labor represents all labor, <u>other than those involved in</u> <u>maintenance of buses</u> (e.g., workshop and maintenance personnel), and includes drivers, conductors, traffic supervisors, administrative personnel and others (see personnel costs in CIRT 2012 for details). Total operator labor costs = $[w_4]$ x [annual vehicle-kilometers ( <i>y</i> )]. Note that we have assumed operator labor costs to be constant throughout <i>T</i> , and the same for all bus systems within each country.
W5	Training cost	Training costs were calculated based on the hours of training for each bus system, as described in relation to training hours (e) above, and average salaries estimated on an hourly basis from CIRT data for DTC, BEST, BMTC and MTC-CNI in 2010-11, for staff involved in bus operations, maintenance, and administration (CIRT, 2012). On this basis, the average hourly salaries we used for operators, mechanics and administrators respectively in our calculations were USD 2.84, 4.92 and 6.94.

## **Appendix C: Supplementary material for Chapter 6**

#### C.1: Methane (CH<sub>4</sub>): tailpipe and fugitive emissions

Many studies do not provide methane emissions estimates, with the exception of Hesterberg et al. (2008); though, even within this study, the number of data points was particularly small. In Nylund et al. (2007), it is argued that methane emissions can also be accurately estimated from hydrocarbon emissions, since HC emissions from CNG buses are predominantly methane (Nylund et al., 2007; p.61). This same approach was used in this study, when direct methane emissions were not available, that is, for emission factors that were taken from Bose and Sundar (2005) and ARAI (2008, 2009). Furthermore, since methane leakage is a significant contributor to total methane emissions, this study used, for the different classes of CNG buses, a multiplication factor of 1.3062 to account for leaked methane (Reynolds & Kandlikar, 2008: Table 1, p.5861). Therefore, methane emission factors reported in Table 6.4 incorporate methane leakage in the estimates.

#### C.2: Black Carbon (BC)

BC is a product of incomplete combustion of fossil fuels, and is a strong light-absorbing component of particulate matter (EPA, 2012). In the US, mobile sources were responsible for 52% of BC emissions, but 93% of these came from diesel vehicles or engines (EPA, 2012), which is critically relevant within the Indian context, where -- nationally -- the proportion of diesel to gasoline use is 5:1 (TERI & ICCT, 2011). In estimating emissions from mobile sources, Bond et al. (2004) calculate the fraction of BC in PM to be approximately 66±16%, with no significant variation in this ratio between normal vehicles (i.e., well-maintained) or super-emitting ones. They estimate BC emissions to be a function of PM<sub>10</sub>, on a g-per-kg of fuel basis, and the proportion of BC in PM<sub>10</sub>, emphasizing that no differentiation between light-duty and heavy-duty vehicles is necessary, as emission factors per mass of fuel (not per distance) are very similar among the various vehicle classes. Note that in the absence of PM<sub>10</sub> emissions, they

take total particulate matter, as they are similar, with Cohen et al. (2003), also making no distinction between PM or  $PM_{2.5}$  (ultra-fine particulate matter) emissions, as they argue that the bulk of PM emissions is comprised of  $PM_{2.5}$ . In the absence of specific ultra-fine particulate matter or even  $PM_{10}$  data, we also took PM data that was available.

Contextually, the challenge in measuring BC is presented by the effect of super-emitting vehicles (i.e., "smokers"; Faiz et al., 1996) on PM emissions, as, even though the ratio of BC in PM will not change, total PM emission outcomes will (Bond et al., 2004). Taking into account this factor, Reynolds and Kandlikar (2008) estimate the proportion of BC in PM, within the context of Delhi, to be 76% for diesel buses and 29% for CNG buses, based on the assumption that 40% of all buses in Delhi were super-emitting vehicles, and based on Bond et al. (2004) methodology, that estimate BC from a PM<sub>1</sub> (i.e., fine particulate matter with diameter smaller than one micrometer), in which they assume that approximately 86% of diesel PM<sub>10</sub> is composed of emissions with diameters smaller than one micrometer; so, actually 65.4% of PM<sub>10</sub> is BC for diesel emissions. Also, recent studies for Delhi, based on chassis dynamometer mass emission tests on vehicles and using a composite vehicle fleet profile that accounts for vehicles of different vintages and categories, the fraction of BC in PM<sub>10</sub> for diesel buses was estimated to be around 27±2%, while for CNG buses it was 20±1% (ARAI, 2009; Gargava et al., 2014). The difference in estimates between Reynolds and Kandlikar (2008) and the last set of references in Delhi may be accounted by, in part, the source and vintage of PM emission factor data, for which Gargava et al. (2014) use data based on recent chassis dynamometer test and for a more up-to-date bus technology, but do not account for vehicles that could have a super-emitting profile, as done in Reynolds and Kandlikar (2008); furthermore, the sample of vehicles used in Gargava et al. (2014) was small for meaningful statistical significance. In view of all of the foregoing, in estimating BC emissions the average of the two estimates (Reynolds & Kandlikar, 2008; Gargava et al., 2014) was used for the purpose of this investigation, using the average proportion of BC in PM of 46% for diesel buses and 25% for CNG buses.

#### C.3: Carbon Dioxide (CO<sub>2</sub>)

CO<sub>2</sub> emissions were estimated from fuel economy data for Delhi and India, which reflects up-to-date performance of standard and newer fleet of low-floor buses currently in operation in Delhi and other major Indian cities, as well as the performance of early 2000 diesel and CNG buses (Chapter 4). This will allow for more accurate representation of engine and vehicle types in India, since fuel economy is an accurate estimator for CO<sub>2</sub> emissions based on the fraction of carbon content in the fuel, and assuming that all fuel oxidizes (Lipman & Delucchi, 2002; MJB, 2007). Table C2 details data and sources for CO<sub>2</sub> emission factor estimation, which is based on the following equation, noting that CO<sub>2</sub> emissions for CNG were calculated based on energy equivalence to diesel, given fuel characteristics in India:

$$CO_2 \left(\frac{g}{km}\right) = \frac{CO_2 - to - Carbon Ratio\left(\frac{CO_2 \frac{B}{mol}}{C\frac{B}{mol}}\right) \times Carbon \text{ percentage in Fuel (Wt.%)}}{Fuel Economy\left(\frac{km}{L}\right) \times Fuel Liquid Density\left(\frac{L}{kg}\right)} \times 1000 \text{ (g/kg)}$$

#### C.4: CO<sub>2</sub> equivalent (GWP for BC, CH<sub>4</sub>, CO<sub>2</sub>)

As per IPCC methodology, the combined effect of GHG emissions from buses is expressed in terms of CO<sub>2</sub> equivalent emissions, based on the global warming potential (GWP) of BC, CH<sub>4</sub> and CO<sub>2</sub> using the concept of radiative forcing (RF) as means of comparing the strength of various mechanisms that are causing climate change over a 100 year life-cycle, that is GWP<sub>100</sub> (Myhre et al., 2013: Table 8.7, p.714). I should also point out the uncertainties in estimating RF from different polluting gases. Well-mixed greenhouse gases, including CO<sub>2</sub> and CH<sub>4</sub>, have very robust evidence and high confidence level in their estimates, while aerosols, such as BC, less so (Myhre et al., 2013: Section 8.5.1). The challenge in estimating the effect of BC on global warming is that, given its physical properties, it is not well mixed in the atmosphere and with different and uncertain indirect and snow/albedo effects, resulting in estimates for GWP<sub>100</sub> ranging from 330 to 2240 CO<sub>2</sub>(e), if this metric has to be used (EPA, 2012: Section 2.7.3.1, p.61). The extent and nature of these uncertainties is beyond the scope

of this study, and so central GWP<sub>100</sub> estimates were taken, based on literature that is available (Table C3).

## Tables:

# Table C1: Heavy-duty vehicle emission norms, India vs. European UnionIndia:

	Effective date	Categor	ry	Test cycle	CO (g/kWh)	THC (g/kWh)	NOx (g/kWh)	NMHC (g/kWh)	CH4 (g/kWh)	PM (g/kWh) <sup>*</sup>	ELR smoke (m <sup>-1</sup> )*	Free accl smoke (m <sup>-1</sup> ) <sup>a</sup>
Nagal	1.04.05	Diesel, CNG or LF with GVW >	PG vehicles >3500	Engine Steady state cycle (ESC)	2.10	0.66	5.00	NA	NA	0.10 / 0.13 <sup>b</sup>	0.80	2.45
CNG or	BS-III	Diesel , CNG or L with GVW >3500 exhaust after treat	PG vehicles with advanced tment system	Engine Transient Cycle (ETC)	5.45	0.78	5.00	NA	NA	0.16/ 0.21 <sup>b</sup>	0.80	2.45
LPG Ingines	1.04.10	Only Diesel with GVW	vehicles >3500	Engine Steady state cycle (ESC)	1.50	0.46	3.50	NA	NA	0.02	0.50	2.45
	BS-IV	Diesel, CNG or L with GVW	PG vehicles >3500	Engine Transient Cycle (ETC)	4.00	-	3.50	0.55 <sup>°</sup>	1.10 <sup>d</sup>	0.03	NA	NA
оре	an U	nion:										
		EU Emission	Standard	<b>Ta</b> s for Heavy-Dւ	ble 1 ity Die	sel Eng	ines: St	eady-S	tate Te	sting		
tage		EU Emission Date	Standard	Ta s for Heavy-Du t <u>CO</u>	ble 1 ity Die	sel Eng	ines: St NOx	eady-S	tate Te	sting PN	<u>s</u>	Smoke
tage ro I	1992.	EU Emission Date ≤ 85 kW	Standard Tes	Ta s for Heavy-Du t CO	ble 1 ity Die	sel Eng HC g/kV	ines: St NOx Vh 8.0	eady-S	tate Te PM .612	sting PN 1/kW	th	Smoke 1/m
tage ro I	1992, 1992,	EU Emission Date ≤ 85 kW > 85 kW	Standard Tes	ta s for Heavy-Du t 9 4.5 4.5	ble 1 ity Die I	sel Eng HC g/kV	ines: St NOX Vh 8.0 8.0	eady-S	tate Te PM .612 0.36	sting PN 1/kW	h	Smoke 1/m
ro I	1992, 1992, 1996.	EU Emission Date ≤ 85 kW > 85 kW 10	Standard Tes	Ta s for Heavy-Du t 29 4.5 4.5 4.0	ble 1 ity Die	sel Eng HC g/kV 1.1 1.1	NOx NOX Wh 8.0 8.0 7.0	eady-S	.612 0.36	PN 1/kW	'h	Smoke 1/m
tage ro I ro II	1992, 1992, 1996. 1998.	EU Emission Date ≤ 85 kW > 85 kW 10	Standard Tes	Ta s for Heavy-Du 29 4.5 4.5 4.0 4.0	ble 1 ity Die i i i i i i i i i i i i i i i i i i	sel Eng HC g/kV I.1 I.1 I.1 I.1	NOx NOX Wh 8.0 8.0 7.0 7.0	eady-S	.612 0.36 0.25 0.15	sting PN 1/kW	'h	Smoke 1/m
tage ro I ro II ro III	1992, 1992, 1996. 1998. 1998.	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard Tes ECE R-4	Ta s for Heavy-Du 29 4.5 4.5 4.0 4.0 4.0 5LR 1.5	ble 1 Ity Die	sel Eng HC g/kv 1.1 1.1 1.1 1.1 1.1 1.2 5	NOX NOX Wh 8.0 7.0 7.0 2.0	eady-S	.612 0.36 0.25 0.15 0.02	PN 1/kW	<u>'h</u>	Smoke 1/m 0.15
iro I Iro II	1992, 1992, 1996. 1998. <i>1999.</i> 2000.	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 EEV only 10	Standard Tes ECE R-4	Ta s for Heavy-Du 29 4.5 4.5 4.0 4.0 4.0 6LR 1.5 2.1	ble 1 Ity Die 1 1 1 1 1 1 1 1 1 0 0 0	HC g/kv 1.1 1.1 1.1 1.1 1.1 1.25 .66	NOX NOX 8.0 8.0 7.0 7.0 2.0 5.0	eady-S	.612 0.36 0.25 0.15 0.02 0.10 <sup>a</sup>	sting PN 1/kW	<u>t</u> h	Smoke 1/m 0.15 0.8
iro II iro II iro III	1992, 1992, 1996. 1998. 1999. 2000. 2005.	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10	Standard Tes ECE R-4	Ta s for Heavy-Du 29 4.5 4.5 4.0 4.0 4.0 4.0 6LR 2.1 1.5	ble 1 ity Die 1 1 1 1 1 1 1 1 1 1 0 0 0 0 0	HC g/kv I.1 I.1 I.1 I.1 I.1 I.1 I.225 I.666 I.46	NOX NOX Wh 8.0 7.0 7.0 2.0 5.0 3.5	eady-S	tate Te PM .612 .0.36 .0.25 .0.15 .0.02 .10 <sup>a</sup> .0.02	Sting PN 1/kW	h	Smoke 1/m 0.15 0.8 0.5
ro II ro II ro III ro IV ro V	1992, 1992, 1996, 1998, 1998, 2000, 2005, 2008,	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10	Standard Tes ECE R-4 ESC & E	Ta       colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2">colspan="2"       colspan="2">colspan="2">colspan="2"       4.5     4.5       4.0     4.0       4.0     4.0       4.5     2.1       1.5     1.5	ble 1 ity Die 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	sel Eng 1C g/kV 1.1 1.1 1.1 1.1 1.25 .666 .46 .46	NOX NOX Wh 8.0 7.0 7.0 2.0 5.0 3.5 2.0	eady-S	tate Te PM .612 .0.36 .0.25 .0.15 .0.02 .10 <sup>a</sup> .0.02 .0.02	sting PN 1/kW	<u>h</u>	Smoke 1/m 0.15 0.8 0.5 0.5
ro I ro II ro III ro IV ro V ro V ro VI PM = 0	1992, 1992, 1996, 1998, 1998, 2000, 2005, 2008, 2013,0 .13 g/kWl	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard Tes ECE R-4 ESC & E Standard WHSC 0.75 dm <sup>3</sup> s	Ta s for Heavy-Du 29 4.5 4.5 4.5 4.0 4.0 4.0 4.0 5.1 5.1 5.5 1.5 weet-ume per	ble 1 ity Die 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 0	HC         g/kv           1.1         1.1           1.1         1.1           1.1         1.1           1.25         .66           .46         .46           .13         er and a provider of the second	NOX Wh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 rated por	eady-S	tate Te PM .612 0.36 0.25 0.15 0.02 0.02 0.02 0.02 0.02 0.02 0.02 0.02 0.02 0.02	8.0×10	5011	0.15 0.8 0.5 0.5
Stage Suro I Suro II Suro IV Suro IV Suro V Suro VI - PM = 0	1992, 1992, 1996, 1998, 7999, 2000, 2005, 2008, 2013, 2013, 2013, 2013, 2013, 2013, 2013, 2013, 2013, 2013, 2014,	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard Tes ECE R-4 ESC & E ESC & E WHSC 0.75 dm <sup>3</sup> s	Ta s for Heavy-Du 4.5 4.5 4.5 4.0 4.0 4.0 4.0 4.0 5 1.5 1.5 1.5 1.5 wept volume per transported to the set to the tary of the set to the	ble 1 ity Die i i i i i i i i i i i i i i i i i i	AC 8/4/2014	NOX Vh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 	eady-S	tate Te PM 	8.0×10 min <sup>-1</sup>	5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5 5	Smoke 1/m 0.15 0.8 0.5 0.5
Stage	1992, 1992, 1996, 1998, 1998, 2000, 2005, 2005, 2008, 2013, 0,13 g/kWI	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E WHSC 0.75 dm <sup>3</sup> s	Ta s for Heavy-Du 4.5 4.5 4.0 4.0 4.0 4.0 4.0 4.0 1.5 1.5 1.5 1.5 wept volume per Ta or Heavy-Duty est <u>CO</u>	ble 1 ity Die i i i i i i i i i i i i i i i i i i	AC 8////	NOX V/ 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 cated pool as Engir	eady-S	tate Te PM 	8.0×10 8.0×10 Testing	5 77 5 11 5 5	Smoke 1/m 0.15 0.8 0.5 0.5 0.5 PN <sup>e</sup>
iro I       iro II       iro III       iro IV       iro V       iro VI       PM = 0	1992, 1996, 1998, 1998, 1999, 2000, 2005, 2008, 2013, 2013, 2013, 2013, 2013, 2013, 2013, 2014,	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E WHSC 0.75 dm <sup>3</sup> s	Ta       s for Heavy-Du       t       (1)       (2)       (4)       (5)       (4)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (5)       (6)       (7)       (7)       (7)       (7)       (7)       (7)       (7) </td <td>ble 1 ity Die i i i i i i i i i i i i i i i i i i</td> <td>AC 8/// 8/// 8/// 8/// 8/// 8/// 8/// 8/</td> <td>NOX Wh 8.0 7.0 7.0 2.0 5.0 3.5 0.40 rated pool as Englin CH4<sup>a</sup> g/kW/</td> <td>eady-S 0 ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( (</td> <td>tate Te PM .612 .0.36 .0.25 .0.02 .0.02 .0.02 .0.02 .0.01 ed &gt; 300 ansient NOX</td> <td>sting PN 1/kW 8.0×10 00 min<sup>-1</sup> Testing PM<sup>1</sup></td> <td>5 5 5</td> <td>Smoke 1/m 0.15 0.8 0.5 0.5 0.5 PN<sup>e</sup> 1/kWh</td>	ble 1 ity Die i i i i i i i i i i i i i i i i i i	AC 8/// 8/// 8/// 8/// 8/// 8/// 8/// 8/	NOX Wh 8.0 7.0 7.0 2.0 5.0 3.5 0.40 rated pool as Englin CH4 <sup>a</sup> g/kW/	eady-S 0 ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( (	tate Te PM .612 .0.36 .0.25 .0.02 .0.02 .0.02 .0.02 .0.01 ed > 300 ansient NOX	sting PN 1/kW 8.0×10 00 min <sup>-1</sup> Testing PM <sup>1</sup>	5 5 5	Smoke 1/m 0.15 0.8 0.5 0.5 0.5 PN <sup>e</sup> 1/kWh
age 0   0    0    0    0    0    0    0    1 1 1 1 1 1 1 1 1 1 1 1 1	1992, 1996, 1998, 1998, 2000, 2005, 2008, 2013, 0, 13 g/kW	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E ESC & E WHSC 0.75 dm <sup>3</sup> s andards fr ETC	Ta       s for Heavy-Du       t       (1)       (2)       (4,5)       (4,6)       (4,0)	ble 1 ity Die i i i i i i i i i i i i i i i i i i	AC 8/40 AC 8/4	NOX NOX Vh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 rated poor CH <sub>4</sub> a g/kWl 0.65 1.6	eady-S 0 ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( ( (	tate Te PM .612 .0.36 .0.25 .0.02 .0.02 .0.02 .0.02 .0.01 ed > 300 ansient NOX	8.0×10 8.0×10 0 min <sup>-1</sup> Testing 0.02 0.16	5 77 5 5 7 7 7 7 7 7 7 7 7 7 7 7 7 7 7	Smoke 1/m 0.15 0.8 0.5 0.5 0.5 1/kWh
age 1 1 1 1 1 1 1 1 1 1 1 1 1	1992, 1992, 1996, 1998, 1998, 2000, 2005, 2003, 2013, 0,13 g/kWl EU	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E WHSC 0.75 dm <sup>3</sup> s andards for ETC	Ta       s for Heavy-Du       t       20       4.5       4.5       4.0       4.0       2.1       1.5       1.5       1.5       wept volume per       abreak       co       co       abreak       co       abreak       abreak       abreak       co       abreak	ble 1 ity Die i i i i i i i i i i i i i i i i i i	AC 8/// 8/// 8/// 8/// 8/// 8/// 8/// 8/	NOX NOX Vh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 0.40 0.40 cH <sub>4</sub> <sup>a</sup> g/kWH 0.65 1.6 1.1	eady-S	tate Te PM .612 	8.0×10 8.0×10 00 min <sup>-1</sup> Testing 0.02 0.16 0.02	5 77 5 5 5 5 6 6 7 7 7 7 7 7 7 7 7 7 7 7 7	Smoke 1/m 0.15 0.8 0.5 0.5 0.5 1/kWh
tage       ro I       ro II       ro IV       ro V       ro VI       PM = 0       tage       ro III       ro III       ro IV	1992, 1992, 1996, 1998, 1998, 2000, 2005, 2008, 2013, 0,13 g/kWI EU EU 2000 2005 2008	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E WHSC 0.75 dm <sup>3</sup> s andards fr ETC	Ta s for Heavy-Du 4.5 4.5 4.5 4.0 4.0 4.0 4.0 4.0 1.5 1.5 1.5 1.5 1.5 1.5 1.5 1.5 1.5 4.0 4.0 4.0 5.45 4.0 4.0 5.45 4.0 4.0	ble 1 ity Die ity Die i i i i i i i i i i i i i i i i i i	AC 8/40 AC 8/40 AC 8/40 AC 8/55 AC	NOX NOX V/I 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 0.40 rated pool as Englin CH <sub>4</sub> a g/kW/I 0.65 1.6 1.1	eady-S 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	tate Te PM .612 	8.0×10 8.0×10 00 min <sup>-1</sup> Testing 0.02 0.16 0.03 0.03	2011 0 000 000 0 000 000000	0.15 0.8 0.5 0.5 1/kWh
ge 1 11 11 11 11 11 11 11 11 11 11 11 11	1992, 1992, 1996, 1998, 1999, 2000, 2005, 2008, 2013, 2013, 2013, 2013, 2014, 2000, 2005, 2008, 2013, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 2000, 2005, 200, 200	EU Emission Date ≤ 85 kW > 85 kW 10 10 10 10 10 10 10 10 10 10	Standard ECE R-4 ESC & E ESC & E WHSC 0.75 dm <sup>3</sup> s andards fr ETC WH	Ta       s for Heavy-Du       t       20       4.5       4.5       4.0       4.0       4.0       4.0       4.0       4.0       4.0       4.0       4.0       1.5       1.5       1.5       1.5       asymptotic boundary       asymptotic boundary <td>ble 1 ity Die ity Die i i i i i i i i i i i i i i i i i i</td> <td>AC 8/40 AC 8/40 AC 8/40 AC 78 AC 78</td> <td>NOX NOX Vh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 cated poor as Engir <u>CH4<sup>a</sup></u> <i>g/kWh</i> 0.65 1.6 1.1 1.1</td> <td>eady-S 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0</td> <td>tate Te PM </td> <td>8.0×10 8.0×10 0 min<sup>-1</sup> Testing 0.02 0.16 0.03 0.03 0.01</td> <td>2 5 2 7 3 7 5 7 6 7 6 7 6 7 7 7 7 7 7 7 7 7 7</td> <td>Smoke 1/m 0.15 0.8 0.5 0.5 1/kWh 1/kWh .0×10<sup>11</sup></td>	ble 1 ity Die ity Die i i i i i i i i i i i i i i i i i i	AC 8/40 AC 8/40 AC 8/40 AC 78 AC 78	NOX NOX Vh 8.0 7.0 7.0 2.0 5.0 3.5 2.0 0.40 cated poor as Engir <u>CH4<sup>a</sup></u> <i>g/kWh</i> 0.65 1.6 1.1 1.1	eady-S 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0	tate Te PM 	8.0×10 8.0×10 0 min <sup>-1</sup> Testing 0.02 0.16 0.03 0.03 0.01	2 5 2 7 3 7 5 7 6 7 6 7 6 7 7 7 7 7 7 7 7 7 7	Smoke 1/m 0.15 0.8 0.5 0.5 1/kWh 1/kWh .0×10 <sup>11</sup>

		Emissions			Mult.	Carbon in	CO <sub>2</sub> emission
ld.	Bus type/fuel	cert.	FE <sup>1</sup>	FE <sup>2</sup>	Factor <sup>3</sup>	fuel mass <sup>4</sup>	factors
	Unit of measure		km/L	kg/km	factor	% wt.	g/km
1	Std. Diesel (MY00+)	BS-I/II	3.814	0.219	3.664	88.30	708
2	LF Diesel (D50) <sup>5</sup>	BS-III/IV	2.686	0.311	3.664	86.60	986
3	Std. CNG (MY00+)	BS-I/II	2.124	0.334	3.664	74.04	907
4	LF CNG⁵	BS-III/IV	1.598	0.444	3.664	74.04	1205

#### Table C2: CO<sub>2</sub> emission factor estimation for CE analysis. g/km

<sup>1</sup> CNG Fuel Economy (FE) converted to diesel litre equivalent based on energy content of 50 MJ/kg for CNG and 42.5 MJ/kg for diesel (Faiz et al., 1996) and liquid density of diesel of 0.835 kg/l (SIAM, 2013). Bus FE data from LCC model assumptions (Chapter 5).

<sup>2</sup> FE in kg/km converted from FE in km/l; for this, it was assumed a liquid density of 0.835 kg/l for diesel (SIAM 2013), and 0.422 kg/l for CNG (Faiz et al., 1996).

<sup>3</sup> Multiplication factor for fuel mass that is carbon, assuming all fuel oxidizes; ratio calculated using the molar mass of CO<sub>2</sub> (44 q/mol) divided by molar mass of Carbon (12 g/mol).

<sup>4</sup> Carbon percentages, with respect to weight, for D350 and D50 from Bose & Sundar (2005); for CNG buses estimate assumes 92.6% of natural gas is methane (% of molar fraction), based on WVU (2005), and very close to reported methane fraction in natural gas in Delhi (Reynolds & Kandlikar 2008).

FE reflects weighted average of AC (20%) and non-AC (80%) buses in fleet.

<sup>6</sup> DTC doesn't operate modern standard CNG buses, only LF. FE estimated from CO<sub>2</sub> emissions in ARAI (2009) for OEM CNG buses, MY05+.

Compound	GWP <sub>100</sub>	Source
CO <sub>2</sub>	1	Myhre et al. (2013): Table 8.A.1 (p.731)
CH <sub>4</sub>	28 (34) <sup>1</sup>	Myhre et al. (2013): Table 8.7(p.714)
BC	455 (± 57%)	Reynolds & Kandlikar (2008)
	740 <sup>2</sup>	Myhre et al. (2013)
	(-61% to +72%) <sup>3</sup>	

#### Table C3: CO<sub>2</sub>(e) emission factor estimation

<sup>1</sup> Figure in parenthesis refers to GWP with climate-carbon feedbacks, which is here only for reference.

<sup>2</sup> We take the average range value for  $CO_2(e)$ . GWP<sub>100</sub> for BC is 740 for transport in Asia, as used in the latest IPCC report supporting the average GWP value of 659 (Myhre et al. 2013: Table 8.SM.16); For reference, Bond and Sun (2005) reported GWP100 of 680.

<sup>3</sup> Table 8.SM.7 (Myhre et al. 2013), for uncertainty in RF by BC.