

Agent-Based Modeling for use in wildlife diseases:

A case study on how environmental variables can affect host-parasite interactions in the red
colobus monkey

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March, 2010

A thesis submitted to McGill University in partial fulfillment of the requirements of the degree
of Master of Science.

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Thesis Format and Style

This is a manuscript-based thesis composed of one paper, which has been submitted to the journal of Ecological Modelling. This paper represents the research I have conducted for my Masters degree in science. To create a cohesive text, and to give additional background on this research, a general introduction and conclusion are provided.

Contributions of Co-Authors

I am the lead author on the paper included in this thesis; I was responsible for conceptualization of the research problem, designing and building the agent-based model, running simulations and interpreting the results, and writing the manuscript. Colin Chapman, my thesis co-advisor, provided support and advice throughout my MSc and contributed data, as well as made various intellectual contributions; as such, he is co-authored on this chapter. Raja Sengupta, my other thesis co-advisor, provided support and advice throughout my MSc and contributed resources, as well as made various intellectual contributions; as such, he is co-authored on this chapter. Tony Goldberg provided support and advice on writing the manuscript as well as providing intellectual contributions; as such, he is co-authored on this chapter.

Acknowledgements

For continued support and encouragement, I would like to thank my two supervisors Colin Chapman and Raja Sengupta, as well as Tony Goldberg. Funding the Ugandan research was provided by Canada Research Chairs Program, Wildlife Conservation Society, Natural Science and Engineering Research Council of Canada, National Geographic, and the National Science Foundation to CC, and to the generous support of the Morris Animal Foundation through its support of the Kibale EcoHealth Project (award #D07ZO-024 to T.G.). Support for the modeling work was provided by Natural Science and Engineering Research Council of Canada (RS, CC). Permission to conduct the research in Uganda was given by the National Council for Science and Technology and the Uganda Wildlife Authority. I would also like extend a special thanks to Mel Lefebvre, Lauren Chapman and Julian Zhunping for helpful comments on this research, as well as to Margaret Kalácska for providing access to computer resources.

Abstract

The effect of anthropogenic landscape change on disease in wildlife populations represents a growing conservation and public health concern. Red colobus monkeys (*Procolobus rufomitratus*), an endangered primate species, are particularly susceptible to habitat alteration and have been the focus of a great deal of disease and ecological research as a result. To infer how landscape changes can affect host and parasite dynamics, a spatially explicit agent-based model is created to simulate movement and foraging of this primate, based on a resource landscape estimated from extensive plot-derived tree population data from Kibale National Park, Uganda. Changes to this resource landscape are used to simulate effects of anthropogenic forest change. With each change in the landscape, outcomes of disease within the simulated red colobus population are monitored using a hypothetical microparasite with a directly-transmitted life cycle. The model predicts an optimal distribution of resources which facilitates the spread of an infectious agent through the simulated population; the density of resource rich sites and the overall heterogeneity of the landscape both being important factors. We find that the characteristics of this optimal distribution are similar to those of logged sections of forest adjacent to our study area.

Résumé

La manière dont le changement du paysage anthropique modifie la dynamique des populations d'espèces sauvages est une préoccupation croissante pour la conservation et pour les praticiens de la santé publique. Les singes colobus rouge (*Procolobus rufomitratus*), une espèce primate en danger, ont démontré une sensibilité à l'altération de leurs habitats, et ont été au centre de beaucoup recherche écologique. Pour déduire comment les modifications paysagées peuvent affecter les dynamiques hôtes et parasites chez le colobus rouge, un modèle spatialement explicite basé sur des agents a été créé pour simuler le mouvement de ce primate. Ce modèle est basé sur un paysage dérivé de données venant de populations arborescentes du Parc National Kibale en Ouganda, recueillies entre 1989 et 2006. Des modifications apportées à ce paysage de ressources ont été faites pour simuler les changements potentiels causés par les modifications de la forêt anthropique et naturelle. Avec chaque modification dans le paysage, les résultats des maladies émergentes au sein de la population simulée du colobus rouge ont été surveillés pour la présence d'un micro-parasite hypothétique ayant un cycle de vie directement transmis. Un modèle compartimentée a été utilisé (SEIR) où les primates étaient considérés soit susceptibles (S), exposés au parasite (E), infectieux (I), ou rétabli d'une infection (R). Nous constatons qu'il existe une répartition optimale des ressources qui facilite la propagation d'une maladie infectieuse par le biais de la population simulée. Nous avons également trouvé que les caractéristiques de cette distribution optimale suggèrent que la propagation de maladies infectieuses serait facilitée dans les régions perturbées.

Introduction

Environmental degradation and habitat loss worldwide is contributing to changes in host parasite dynamics on a scale that is threatening to conservation goals and public health (Daszak et al., 2000; Patz et al., 2000; Jones et al., 2008). Conservation efforts in and around Kibale National Park, Uganda, have examined the role of parasitism in the decline of the red colobus monkey (*Procolobus rufomitratu*s) (Chapman et al., 2005a; Gillespie and Chapman, 2006). These studies have highlighted the need to further explore the anthropogenic effects on primate and pathogen interactions. The following research aims to better understand the role that landscape structure plays in influencing the dynamics of host-parasite interactions. Our goal will then be to infer how changes to landscape structure will affect the spread of infectious diseases among red colobus within our study area.

To link landscape structure to disease outcomes, I draw upon theoretical developments in the fields of landscape ecology and epidemiology. The field of landscape ecology can be described as the study of how the composition and spatial structure of habitats affect ecological patterns and processes (Pickett and Cadenasso, 1995; Wu and Hobbs, 2002; Donovan and Strong, 2003). In this particular case, the ecological process of interest is the relationship between red colobus and their parasites, and the spatial structure influencing this ecological process is the distribution of red colobus resources.

Epidemiology can be described as the study of the factors which determine the frequency and distribution of disease, disability and death within populations (Rothman, 2002; Timmreck, 2002). Within epidemiology, the question of how landscape alters host-parasite dynamics is receiving increased interest, especially in the newly emerging (or re-emerging) field of spatial

epidemiology (Ostfeld et al., 2005). The need to develop a better understanding and theory of how space and spatial structure affects host parasite dynamics in wildlife is driven from an increased awareness of the role that anthropocentric changes, such as habitat fragmentation, are having in increasing the rate of emerging infectious diseases (Daszak et al., 2001; Patz et al., 2004). Typically, in classical epidemiological models, spatial structure is vastly simplified, which in certain circumstances cannot always be assumed or is not in of itself desirable (Keeling and Eames, 2005). Meta-population, network and agent-based (or individual based) models can be seen as recent attempts to include more spatial structure within epidemiology to produce a better understanding of natural systems (Hess, 1996; Keeling, 1999; Bian, 2004; Nunn et al., 2008). In some human and animal systems, the need for complex detailed environmental structure has proven to be unessential for understanding the macro level phenomenon of host-parasite dynamics (Bansal et al., 2007). However, in the case of red colobus, because movement and behavior is tied with environmental structure (Snaith and Chapman, 2008), there is an opportunity to gain more insight by including a more detailed description of the landscape.

To incorporate a landscape ecology framework within epidemiological theory, a theoretical model was built. This model was designed with the ability to input a landscape, with a defined structure, which then plays a role influencing the behavior of the red colobus host. This approach, suggested by Brown et al (2006) and described as a generative landscape science, uses bottom up modeling tools, such as ABM, to specifically to examine ecological patterns and processes produced from landscape patterns and structures.

An agent-based model framework was used to create the simulated behavioral patterns of individual red colobus. The agent-based modeling approach has been used successfully in modeling systems where system level dynamics is driven from individual level behavior. These

models often operate in complex ways that exhibit non-linear behavior, mimicking patterns seen in the natural world (Grimm and Railsback, 2005; Sengupta and Sieber, 2007). Within our model, the behavior of individual red colobus hosts determined the overall population rates of contact and is the main driver for the spread of an infectious disease throughout the simulated population of red colobus. As the behavior of the hosts is influenced by the landscape in our model, spatially dependant outcomes are generated and the effects of landscape structure can be tested.

Recent use of similar spatially explicit agent-based models for applications in infectious disease modeling has added an additional factor to epidemiological models, allowing for interactions between individual hosts and their environments (Bian, 2004; Bian and Liebner, 2007; Roche et al., 2008; Perez and Dragicevic, 2009). With advances in geographical information systems (GIS) and in remote sensing (RS), these models are able to develop more detailed spatial models that can incorporate complex spatial structures of the environment within ABMs (Brown et al., 2005). For example, the uses of GIS/RS in combination with ABM has been used successfully in the epidemiological context by looking at the spread of malaria within human populations (Linard et al., 2009). The strength of this approach combines bottom-up modeling, focused on the effects of the environment and individual behaviors, with advanced mapping techniques allowing for detailed environmental structures to have influence on disease outcomes.

Within the following chapter, I aim to incorporate the theoretical framework from landscape ecology and epidemiology with technical approaches of ABM and GIS to better understand the role that the distribution of resources plays on host-parasite dynamics for the red colobus. With this knowledge, I hope to add to a better understand of how changes to the

landscape will affect the red colobus in Kibale National Park, Uganda, as well as provide insight to other host parasite systems in which mobile hosts are reliant on spatially distributed resources.

Chapter 1: Landscape resource distribution identifies hot spots of disease transmission: An agent-based model of red colobus resources and disease dynamics

Introduction

Increasing environmental change, driven by anthropogenic causes, is recognized as a major challenge to global health (Daszak et al., 2001; Patz et al., 2004; Jones et al., 2008). In the past 50 years, the size of the human population grew by 3.7 billion people (Potts, 2007) and in the next 50 years the global population is expected to grow to surpass 9 billion people, with most of this growth occurring in the tropics (United Nations, 2009). This exponential growth of the human population forebodes many challenges to ecological systems, and continued growth is predicted to exacerbate the increasing demands for environmental products and services (Houghton, 1994). Consequently, demand for resources has increased, bringing about large-scale alterations of environmental conditions for wildlife populations. While some effects are obvious (e.g., deforestation, fragmentation, and habitat loss), others are likely to be subtle and difficult to quantify (e.g., future ecosystem composition and host/vector interactions). These effects, although complicated and difficult to generalize, are likely important drivers of emerging infectious diseases (Daszak and Cunningham, 2003).

Species and populations in the order Primates are particularly suitable for investigating such complex issues because of the large amount of information already available on their ecology, and because primate species are in decline throughout the world as a result of anthropogenic pressures (Chapman and Peres, 2001; Struhsaker, 2005; Mittermeier et al., 2007). Habitat destruction and hunting are thought to be the main factors driving the declines in primate populations (Mittermeier et al., 2007). However, these same anthropogenic changes can also dramatically alter host parasite interactions and are thus a concern for primate health (Formenty

et al., 1999; Hahn et al., 2000; Graczyk et al., 2002; Chapman et al., 2005a; Goldberg et al., 2008b). The health of non-human primates is, in turn, of specific concern for human public health, due to the high risk that primates pose as reservoirs of zoonotic pathogens (Wolfe et al., 2007; Davies and Pedersen, 2008). This is illustrated by such pathogens as the human immunodeficiency virus (HIV; the cause of AIDS) (Hahn et al., 2000) and *Plasmodium falciparum* (the cause of virulent malaria) (Rich et al., 2009), both of which trace their origins to chimpanzees. In a recent study on the origins of major human infectious diseases, Wolfe et al. (2007) found that even though non-human primates constitute 0.5% of all vertebrates, their zoonotic transmission has contributed about 20% of the major human infectious diseases.

Red colobus (*Procolobus rufomitratus*) of Kibale National Park, Uganda offer an ideal modelling system for examining the effects of habitat change on primate disease dynamics. There are 40 years of long-term data on the ecology, demography and responses to anthropogenic change for the red colobus of Kibale (Struhsaker, 1997; Chapman et al., 2005b). Additionally, the health of the red colobus has received considerable recent attention (Chapman et al., 2007; Goldberg et al., 2008a; Goldberg et al., 2009). This research has revealed evidence of transmission of disease agents between red colobus and humans (Goldberg et al., 2008b) and has highlighted the important role that fragmented landscapes play in augmenting parasitism within the red colobus population (Gillespie and Chapman, 2006).

Our research is specifically designed to examine how the spatial distribution of resources in a forest habitat might affect the transmission of parasites within a red colobus population. We suspect resource distribution might be a influencing factor in transmission of parasites from observations of how red colobus use their habitats; where they choose to forage, how social groups organise, and rates of contact between social groups. Red colobus live in social groups in

which movement appear to be largely driven by foraging for food. Red colobus are folivorous primates, whose foods vary in spatial aggregation based on individual tree species and on tree size (Chapman and Chapman, 2002). The characteristics of forests habitats are therefore important in determining the amount and distribution of resources for these primates, and hence, movement behaviour of groups. To test ideas about how forest compositions, and therefore variation in the distribution of these resources, could affect disease transmission within a population of red colobus we created a simulation model. This simulation model would allow us to input a resource landscape upon which a simulated red colobus population, built with detailed behavioural data, would forage, allowing us to test landscape effects on parasite transmissions at a population level.

To construct this simulation model, an agent-based modeling (ABM) framework was used (Grimm and Railsback, 2005; Sengupta and Sieber, 2007). This approach has been used as an effective tool in simulating primate group behaviour (Te Boekhorst and Hogeweg, 1994; Hemelrijk, 2002; Bryson et al., 2007; Sellers et al., 2007), as well as infectious disease spread among primates (Nunn et al., 2008; Nunn, 2009). The overall construct of our model follows previous theoretical constructs of agent-based models which were focused on spatially explicit epidemiological simulation (Bian, 2004; Roche et al., 2008). We also make use of recent software advances linking ABM models with GIS (geographic information system) capabilities, allowing more detailed geographies to be incorporated implicitly, which allows these simulations to link spatial and temporal processes (Brown et al., 2005); for examples see, Bennett and Tang (2006), Perez and Dragicevic (2009) and Kramer-Schadt (2009). Linking spatial and temporal processes has been recognised as crucial for addressing questions regarding landscape effects on disease (Ostfeld et al., 2005). In this case it allows us to link long-term forest data from our

study area to a wealth of behavioural and epidemiological data for red colobus.

1.1 Study Site and Data Collection

We focus on the red colobus living inside Kibale National Park, located in western Uganda.

Kibale is composed of moist-evergreen forest ($\sim 795 \text{ km}^2$; $0^\circ 13' - 0^\circ 41' \text{ N}$ and $30^\circ 19' - 30^\circ 32' \text{ E}$), that receives approximately 1697 mm (1990-2009) of rain each year distributed among two rainy seasons (C. A. Chapman and L. J. Chapman unpublished data). We focus on a subsection of the park (Kanyawara K-30 $\sim 250\text{ha}$) where detailed studies of red colobus group movement in relation to habitat quality have been conducted (Chapman et al., 2001; Chapman and Chapman, 2002; Snaith and Chapman, 2005; 2008), as well as studies on forest properties for the past 40 years (Struhsaker, 1997; Chapman et al., 2010). We use tree data from three time periods (1989, 1999, and 2006), where 26 plots (200 x 10 m) were used to record identity and diameter at breast height (DBH) values of trees larger than 10cm DBH.

2. The Model

The simulation model was developed using Repast Simphony 1.2.0 software (<http://repast.sourceforge.net/>) combined with open source GIS software from JTS (Vivid solutions Inc.). The model description follows the ODD (Overview, Design concepts, Details) protocol for describing individual- and agent-based models (Grimm et al., 2006).

2.1 Purpose

The purpose of the model is to test the effects that resource distribution has on parasite transmission within a population of red colobus.

2.2 State variables and scales

The model is composed of three agents: the landscape, the host, and the microparasite. The model's base is the landscape where individual resource polygons (30x30 m) make up a grid surface. Each polygon contains state variables; current amount of resources and maximum resource levels.

The host, red colobus, forages on this surface, represented as points in a continuous space, moving from resource polygon to resource polygon and contains state variables; energy level and the number of neighbours needed to be considered safe from predation. Each red colobus agent also contains spatial memory and builds its own list of remembered sites and estimates of resources at these sites, in real time. Additionally, the red colobus agent is considered to be in one of four states representing the progression of disease, following a SEIR model. Specifically, the red colobus is either (S) susceptible to infection, (E) exposed but not yet infectious, (I) infectious to other hosts, and (R) recovered with immunity developed against this disease.

The microparasite agent, with a maximum of one within a host agent, can successfully infect another host, as determined by a probability function based on distance to the next host. The microparasite agent contains state variables; host primate and maturity level of infection. The microparasite characteristics were modeled after a general class of parasites that are directly transmitted between hosts, through contact and proximity, in which their transmission probabilities are relatively high. We aim here to represent the general behaviour of multiple classes of infectious agents that are transmitted by direct contact, such as filoviruses, poxviruses, or retroviruses (the latter two of which have been documented in Kibale red colobus; (Goldberg et al., 2008a; Goldberg et al., 2009) or close interactions (e.g., respiratory aerosols), such as metapneumoviruses or respiratory syncytial viruses, which have recently been shown to impact

ape conservation across Africa (Kaur et al., 2008; Kondgen et al., 2008). We emphasize that we do not model any specific agent, but rather create a generalized framework that can be adapted to future such efforts.

The scale of the model is established so that one time step in the model represents one half hour, in which red colobus agents, 263 agents as part of 5 distinct social groups, forage on a 225ha grid surface. The model is then run, starting with the infection of one red colobus agent, until the microparasite agents goes extinct or the model reaches 6 months. The 6 months limit was chosen as the model was found to become relatively stable after this time period.

2.3 Process overview and scheduling

The model is intended to simulate a novel parasite for which the primate hosts have no previous exposure. Time is modeled in discrete time steps, where each step is preformed in the same sequence. Each day is considered to be 26 half hour steps for a total of 13 hours, as red colobus are diurnal and are generally only active from 07:00-20:00 (Struhsaker, 1975). Within each time step, each primate agent will, in turn, go through its steps, followed by the parasite agents. At the end of the step the resource agents perform a re-grow step (Fig. 1).

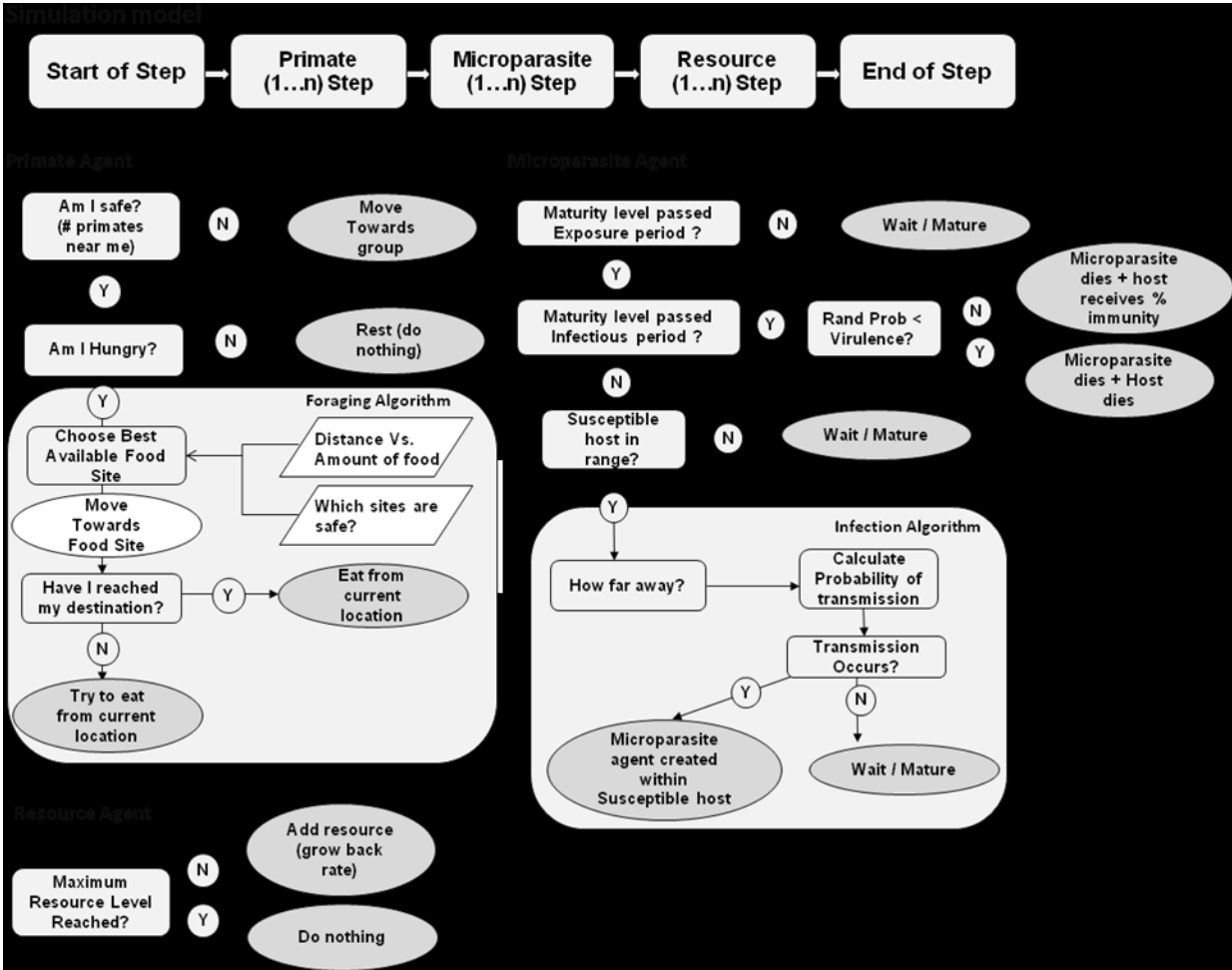


Fig. 1: Flow diagram outlining model processes and general scheduling of the models algorithms. The model processes each type of agent separately; with each agents decision tree described below (details on the decision tree are discussed in section 2.6 submodels). Shaded boxes represent an end point is reached in the decision tree.

2.4 Design Concepts

The model is driven by the interaction between the landscape, the red colobus, and the micro-parasite. The landscape affects the way in which the red colobus forage, including movement and contact rates between groups; movement and contact rates in turn affect transmission rates of the microparasite through the population.

2.4.1. Adaption

The red colobus agent is the only agent considered that displays adaptive behaviour. It can adapt to balance its needs to gain safety within a group against the increased food competition experienced by being a part of the group (Snaith and Chapman, 2008). It accomplishes this, within our model, by measuring its food intake. When there is sufficient food intake, the red colobus agent will value safety within the group; predation pressures are the assumed driver behind the red colobus' need for safety in numbers, however, there is no specific predation within the model. When food intake is less than the ideal, the primate is assumed to be more prone to take risks and values feeding more than security within the group. This assumption is derived from the observation that red colobus groups tend to spread out during periods of food scarcity and contract during periods of food abundance (Snaith and Chapman 2008).

2.4.2. Emergence

Red colobus agents in the simulation model form part of a specific social group, of which there are numerous groups in the system. Each group is only tied by mutual safety requirements, there are no other interactions (e.g., territoriality, intergroup dominance), which is consistent with our current understanding of red colobus socio-ecology (Struhsaker, 1975; Snaith and Chapman, 2008). The lack of strict territorial behaviour is critical for our assumptions of how transmission will occur, as groups of red colobus often forage within close proximity of each other, occasionally within the same tree, allowing for close contact and similar range use (Struhsaker, 2010). Foraging movements by an individual in the group affect others by limiting locations that are considered safe or desirable. Through combined individual decisions, group movement patterns are produced (Fig. 2).

Simulation Model

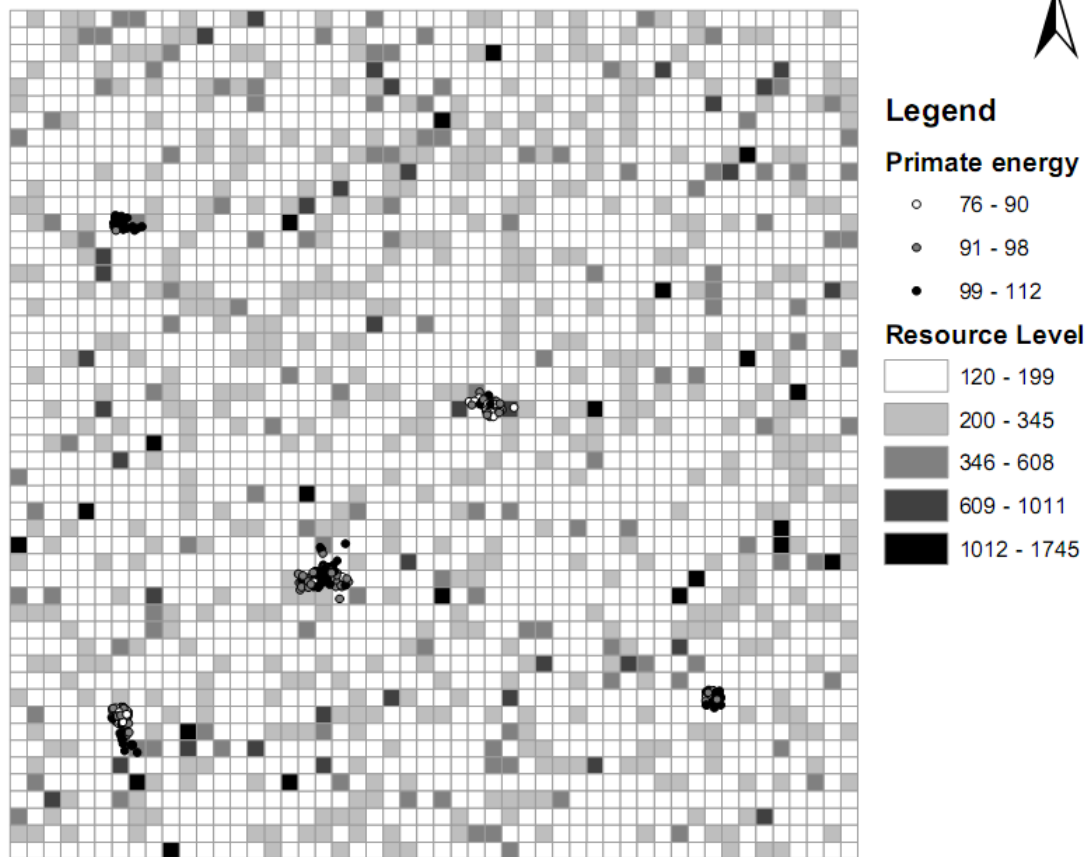


Fig. 2: Simulation environment: individual red colobus agents are represented by circles, resources are represented by grid squares.

Disease spreads through the system through proximity of susceptible host to infectious individuals. Within a specific social group the speed of the infection will spread based on the dynamics of the social group, where as inter group transmission is driven through shared use of the forest landscape. Within the simulation model the driver behind both these transmission routes is the characteristics of the landscape and how they affect red colobus social groups (i.e. group spread, shared resource sites, daily movement patterns). Through movement and foraging choices of these social groups transmission between and within groups of the microparasite agent

leads to the emergence of host-parasite dynamics.

2.4.3 Sensing

The red colobus agent is assumed to know the values of the resources and distances to all resource sites within their search radius. They are also able to remember the location and quantity of past resource sites that contained a significantly higher amount of resources (i.e., spatial memory); it is assumed that the amount remembered at these resource sites will increase by the same re-grow factor as the resource agent, allowing red colobus agents to estimate resource levels at these sites while not within their search radius. The red colobus agent is also assumed to know the location of other primates within its search radius. The microparasite and resource agents do not have any sensory capabilities.

2.4.4 Interaction

The interaction between the host red colobus agent and a microparasite agent is defined by the microparasite. The microparasite will change the status of the red colobus agent based on its stage of development. It will also, at the end of its life cycle either remove the host from the simulation (death) or give it a certain amount of immunity (modeled as a resistance to subsequent infections). Red colobus agents will also remove resources from resource agents (polygons) lowering total resources available in that resource agent for other red colobus agents.

2.4.5 Stochasticity

The transmission of the microparasite is modeled as a stochastic process in which proximity is a factor. This represents our uncertainty in the events leading to a successful transmission from an infected host to that of a susceptible host. Stochasticity is also included in determining the outcome of the infection, resulting in the death of the host or the host acquiring a certain amount of immunity. Again this represents the ignorance of the field with respect to factors leading to

the death of a host. The placement of a red colobus agent within a resource polygon, once selected by the agent as a desired polygon to move towards, is randomly chosen; representing the limited information concerning fine scale within-group movements.

2.5 Initialization

The primary input into the model is the distribution of red colobus resources, in which we aim to reproduce distribution patterns observed at our study site. To accomplish this we use forest structural data. We fit a pareto distribution, a statistical distribution used often to fit data showing a inverse J-shape, to the age class distribution of trees within our study site (Pareto shape 1.5185, location 10; Kolmogorov-Smirnov test; $P < 0.05$). A forest age class distribution is essentially a histogram classing individual trees by their sizes, measured here by their diameter at breast height (DBH). From these distributions we can estimate the distribution of resources available to red colobus, assuming a direct relationship between DBH of trees and food availability. The relationship between DBH and the availability of red colobus resources is one that has been found at multiple sites (Rovero and Struhsaker, 2007; Snaith and Chapman, 2008). Our forest plot data showed a linear relationship between average DBH of a plot and the amount of red colobus food found within that plot (Adjusted R square = 0.627 , $P < 0.01$), collected from the 26 forest plots. We were able to then populate our model environment, a 225ha grid with 30 m x 30 m grid squares, with DBH values selected from this fitted distribution. Each grid cell, now containing DBH values, is then used as an estimate for the amount of resources available to the red colobus. By including a variable “DBHtoFood” to multiply the value of DBH within all sites, we are able to reproduce the linear relationship between red colobus food and DBH. By using a fitted distribution to simulate forest stand structure (age class distribution of forest trees) we are able to simply change the shape of the distribution to redistribute resources in our simulated

environment, creating heterogeneous or homogeneous resource landscapes, mimicking patterns seen at our study site. A final step is added where a common conversion factor is multiplied to each resource cell to maintain equal amounts of resources between simulation runs.

This resource landscape is then populated with 5 distinct social groups, of different sizes, representing average density of red colobus in our study area, as measured by Snaith and Chapman (2008). The starting locations of these groups are randomly chosen, but fixed between runs. Each group is assigned with an initial memory of significant resource sites within an expected home range, based on data on group and home range size (Snaith and Chapman, 2008). One individual from the same group is randomly chosen on the first step and becomes exposed at $t=1$, initiating the disease transmission process.

Table 1: Model input parameters, both those calibrated and non-calibrated, included within the simulations.

Table 1 – Summary of input parameters				
Agent	Parameter	Value	Units	References
<i>Calibrated Parameters (POM approach)</i>				
Red Colobus	Energy gain per feeding	100/11	energy	(Snaith and Chapman, 2008)
	Energy loss per step	100/26	energy	(Snaith and Chapman, 2008)
	Target energy level	100	energy	(Snaith and Chapman, 2008)
	Search Radius	100	meters	-
	Safe Radius	Group Size	meters	-
	Safe Neighbour size+/-	5	agents	-
	Significant DBH size	50	cm	-
	Weight of known sites	2	-	-
Environment	DBH to Food	6	energy/DBH	-
	Grow back rate	5	energy/step	-
	Pareto Location parameter	10	-	(Chapman unpublished data)
Microparasite	Transmission: Contact	2.5	%	-
	Transmission: Droplet	1.25	%	-
	Transmission: Airborne	0.625	%	-
	Incubation Period	19	days	(Hutin et al., 2001)
	Infectious Period	6	days	(Hutin et al., 2001)
<i>Non-Calibrated Parameters</i>				
Environment	Pareto shape parameter	0.5-4.0	-	(Chapman unpublished data)
Microparasite	Virulence	0-100	%	-
	Immunity developed	0,100	%	-

2.6. Submodels

2.6.1. Primate agent energy balance

A simple energy balance was created based on observations of the time spent feeding by red colobus monkeys in the field. Within our study area, red colobus spend on average 43% of their time feeding (Snaith and Chapman, 2008). Thus, the assumption was made that this represents their required food effort per day. Therefore, out of 26 steps, which represents one day in the simulation, approximately 11 of these steps should be feeding events. By arbitrarily choosing 100 as a target energy level, the energy gain per feeding must be $100/11$. The red colobus agent will also lose $100/26$ units of energy each time step, mimicking metabolic requirements and will result in a total loss of 100 units of energy each day, and are considered to be “hungry” when below the target energy level.

2.6.2. Primate agent safety rules

The balance between safety and feeding competition, for the red colobus agent, is accomplished by allowing the agent to modify its parameter “safe neighbour size.” The number of neighbours that the red colobus agent needs to be considered safe will increase when feeding is successful and decrease when feeding targets are not being met. A primate agent considers itself safe when a sufficient number of its group members are nearby; where “nearby” is defined by a circle of radius X around the agent in question. The value of X is determined by a function of the red colobus’ social group size ($\text{Safe Radius} = (1 \text{ meter/individual}) * \text{group size}$), allowing individuals within larger groups to be aware of more group mates.

Each primate will measure its success in feeding by counting the number of feeding events per day. If the primate is feeding on average 11 steps per day, i.e., 43% of the day as observed in our study area, then the number of neighbours to be considered safe is increased. If

it is averaging less than 10 feeding steps per day the number of neighbours is lowered. The primate will also decrease safe neighbour size when energy levels are lower than half the target energy level, indicating a stress period, and increase safe neighbour size when energy targets are exceeded indicating very favourable conditions.

If the primate agent does not consider itself in a safe position, i.e., not enough neighbours nearby, it will move towards its social group. It does this by locating its closest group members, considering only the number of neighbours it needs to be considered safe, and moves to the center of that group. If the center of these nearby members is farther than 50 meters it the agent will move 50 m towards that center. This value of 50 m is based on observations that one rarely sees animals further than this from other group members and this value is thought to be approximately the limit with which an individual could visually see another group member through the canopy (Chapman unpublished data).

2.6.3. Primate agent foraging rules

Primates forage by choosing the best food site based on a simple rule comparing the distance to the food site and the amount of resource at the site. A food site index is created, using this rule, to allow a red colobus agent to choose the best site from sites considered to be safe (Boyer et al., 2006; Ramos-Fernandez et al., 2006). Food sites are considered safe if, once moved, the agent would have enough neighbours nearby.

$$\text{Food Site Index} = \text{distance (meters)} / \text{food (energy)}$$

$$\text{Best Food Site} = \text{Min (Food Site Index)}$$

Once a food site is chosen, the red colobus agent will then move towards it. If the food site is contained within the agent's search radius, the agent is assumed to move directly to the site. If the desired food source is beyond this threshold (i.e., from a remembered site), then the

agent will move towards it, choosing a safe site by again comparing food sites with the Food Site Index, using the distance to the remembered site as a factor ($FSI = \text{distance to target site} / \text{food}$).

We also included a variable, “weight of known sites”, to allow remembered sites to potentially have more draw when choosing possible food sites; representing our assumptions that red colobus use spatial memory in making foraging decisions and will favour past sites in which they have had successful feeding. The value of “weight of known sites” is divided from the food site index of a remembered site, thereby increasing its attraction to a red colobus agent.

The amount of spatial memory used by a red colobus agent was controlled with the variable “significant DBH”. If any resources site was above this value the red colobus considered it to be a significant site and would add it to its list of remembered sites. By increasing and decreasing the significant DBH value we could then increase and decrease the amount of memory used by the red colobus agent in making foraging decisions.

2.6.4. Microparasite transmission

The microparasite agent within our model is able to replicate itself in nearby hosts, where the success of transmission is a function of the distance to other hosts. This function is represented here as a discrete function representing probability of transmission based on stages in hosts proximity; modeling transmission routes of contact, droplet range (from coughing/sneezing), and finally airborne exposure.

Probability of Transmission (distance=x) = {	If $x \leq 2.5\text{m}$	$p_{\text{contact}} = 0.025\%$
	If $2.5\text{m} < x < 5\text{m}$	$p_{\text{droplet}} = 0.0125\%$
	If $5\text{m} < x < 10\text{m}$	$p_{\text{airborne}} = 0.00625\%$

Once transmission occurs, the new microparasite agent will age during its incubation period only becoming infectious after the incubation period has elapsed. At the end of the

infectious period the microparasite's virulence parameter will determine the probability that the microparasite will kill the host, in which case both the host and microparasite are removed from the simulation. We assume here that virulence is independent of the mode of transmission and that the infection and incubation times are relatively static (see Kramer-Schadt et al., 2009 for discussion).

2.6.5. Resource agent re-grow rate

The resource agent (grid square) holds a certain amount of resources, which the red colobus can “eat”. A resource agent, if reduced by foraging, is able to then re-grow at a constant rate (“grow back rate”) every time step, until it reaches its set maximum resource level; which is its initial value at the start of the simulation.

2.7 Model predictions and Observations

To parameterize and build the model we employed “model cycling” with a pattern orientated modeling (POM) approach (Grimm and Railsback, 2005; Kramer-Schadt et al., 2007). With this, we compared patterns of behaviour from our model to selected performance criteria that we have defined. These selected performance criteria were geared towards aspects of the system which we thought were essential in determining transmission rates, consisting of: monthly home range size of red colobus groups, average daily movement of red colobus groups, and the average spread of red colobus group.

Simulated groups of red colobus that followed simple rules balancing safety requirements against increased food competition, were able to reproduce expected trends observed in the field: larger groups foraged over larger home ranges, traveled longer distances per day, had a larger group spread and needed to spend more time moving per day than smaller groups (Snaith and

Chapman 2008; Chapman and Chapman 2000). In lower quality habitats, groups increased their home ranges, travel distance per day, group spread, and time spent moving per day. These trends are consistent with empirical data on group movement patterns when habitat quality has been taken into account (Chapman et al., 2006; Snaith and Chapman, 2008). These patterns held for a wide variety of resource landscapes, as well as different parameterizations of our assumed decision rules. This suggests that our assumptions have captured important components of primate group movement and behaviour. To verify that our simulation could adequately reproduce movement behaviours, we attempted to parameterize the simulation to reproduce the performance criteria for a group of 70 individuals. Compared to our observed data (simulated value / observed value), we were able to test the fit of our models; 86% for average monthly home range, 75% for average day range and 67% for average group spread. As a further test of our model, we varied group size without changing parameterization and compared the simulated results to our observed field data (Fig. 3). The model was able to reproduce trends in our observed data, suggesting that our model has captured many elements of group movement behaviour in red colobus groups.

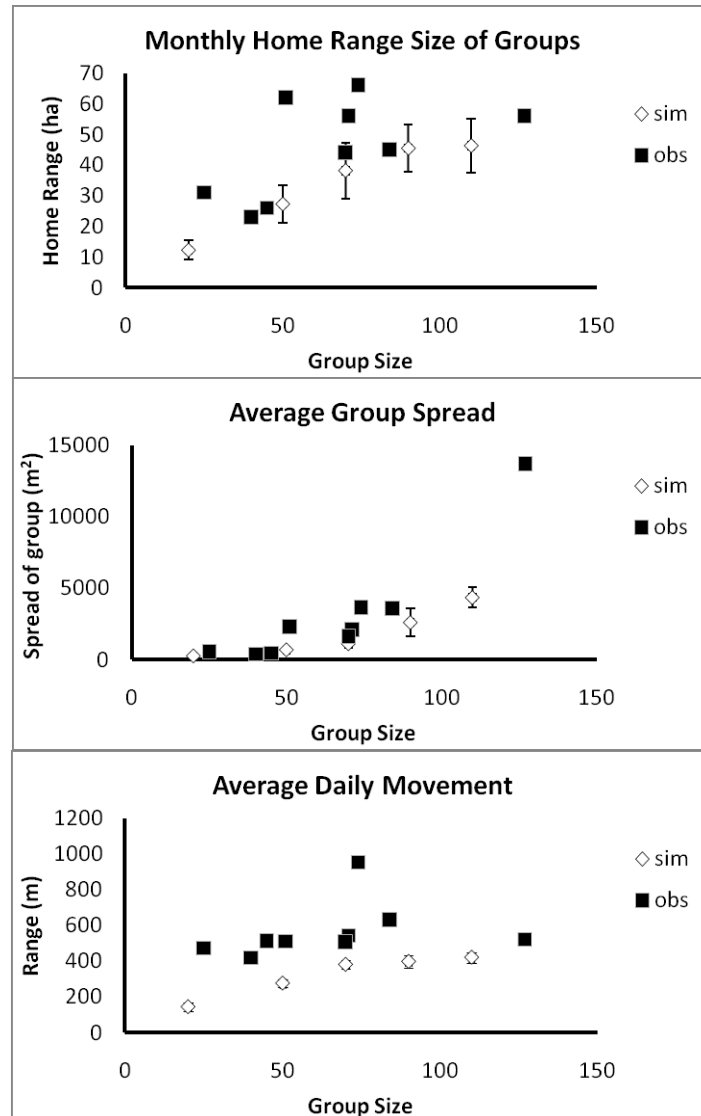


Fig. 3: Model predictions of performance criteria (group monthly home range, average spread of group and average daily movements) while holding environmental variables constant and only varying group size. Observed values are represented by squares, and diamonds represent simulated values. Simulations results follow trends in observed data.

If we examine our model, overall the values for our performance criteria were consistently underestimated. We attribute this to two potential factors. First, during our calibration and construction of the model, we focused on one group of 70 individuals, and did not take increased movement due to food

competition from multiple group scenarios into account. We also attribute some of the underestimations to be due to our relatively simple rules governing social interactions between individuals of a group. The omission of ranking or demographics in our social groups somewhat complicates group movement behaviour within the model, as every primate agent has the same weight in group movement decisions, causing what seems to be, in some cases, indecision in group movement. With the inclusion of higher ranked individuals or group demographics, we might be able to better fit model predictions to observations in average daily movements and home range sizes.

We also found patterns that we did not expect in the simulated groups, but are seen in wild populations of red colobus. Observations in the field have been made of groups, mostly large groups, breaking into smaller subgroups temporarily to forage separately, and then re-group. This is termed fission-fusion of groups. With our simple assumptions regarding the trade off of safety and food competition within a group, we see fission-fusion dynamics in simulated groups undergoing food stress, in such cases as; large groups (~100+ individuals) or with groups in low resource landscapes.

3. Results

When the simulation was run the microparasites spread within social groups, showing a continuous cycling when immunity was not developed and only once when immunity was developed. We also observe the microparasite spread between social groups. By varying forest composition and hence the distribution of resources within the simulation, we were able to test how changes in resource distribution could affect transmission rates in red colobus monkey populations.

Spatial memory, “SigDBH”, was found to be an important factor, influencing the direction of the effects of the landscape. When the spatial memory of groups was not considered

in the simulations, creating more homogeneous landscapes increased the ranging behaviour of primate groups, consistently increasing the probability that groups would contact each other. However, when spatial memory was considered in the model, transmission rates were overall consistently higher, and revealed contrary results. When spatial memory was considered in the simulations, highly homogeneous landscape had relatively lower contact rates; with the rate of contact between groups increasing with increasing resource heterogeneity. However at a point, in highly heterogeneous landscapes, the probability of contact between groups decreased. This shift from an increase to a decrease in the rate of contact can be seen as a reflection of the number of remembered sites held in common between groups. In landscapes with higher number of significant sites, a large number of resources are shared by multiple groups, such that two or more groups will choose the same site at a given time only rarely. This probability of contact increases as the number of sites decreases, allowing for smaller shared memories of significant sites and an increased chance of two or more groups choosing to visit the same site at the same time. As shared memory decreases further, eventually approaching zero, the probability of contacting other groups begins to decrease; reaching levels of transmission similar to model runs without spatial memory. Our model therefore predicts that the density of high resource sites influences the probability of contacting other groups (Fig. 4).

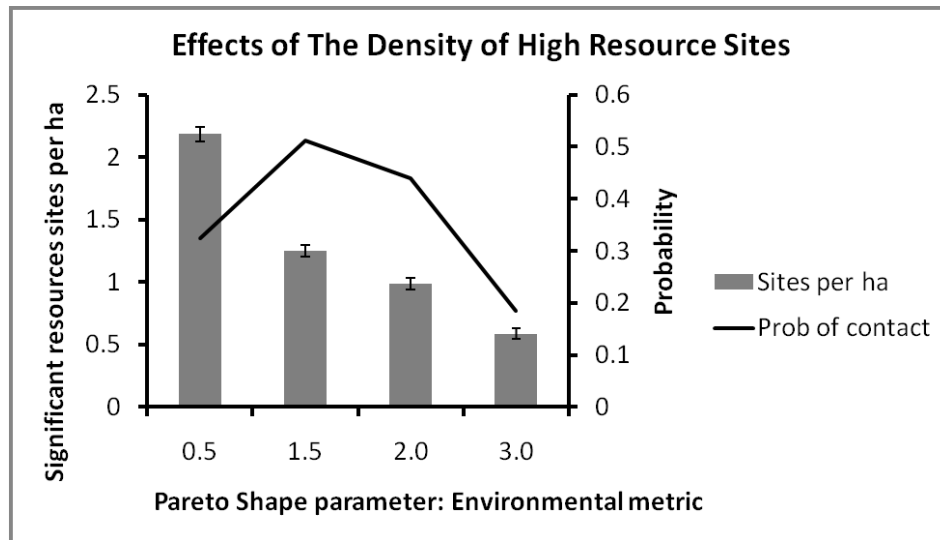


Fig. 4: Effects of density of high resource sites on the probability of contact between groups. The probability of contact is estimated from the percentage of simulations in which all groups became infected by the microparasite. (Simulation results from a total of 297 runs).

The cost of the parasite to the host, modeled here as the probability that the parasite will kill the host after its infectious period, had an effect on overall transmission within the population of red colobus hosts. When immunity to the parasite was not acquired by the hosts, highly virulent microparasites increase overall population survival; with maximum population loss at an intermediate virulence value. At higher virulence, infectious hosts die before being able to infect other hosts, while also reducing the size of their groups and hence reducing their groups' movement. This is somewhat expected from epidemiological theory, regarding parasites that are reliant on their host movement for transmission success (Ewald, 1995). In regards to how the distribution of resources affects this trade-off we find that the virulence value in which the population is reduced the most, shifts as the distribution of resources facilitates overall contact rates within the population, enabling more virulent strains to have a greater impact on the population as well as to spread over larger areas (Fig. 5).

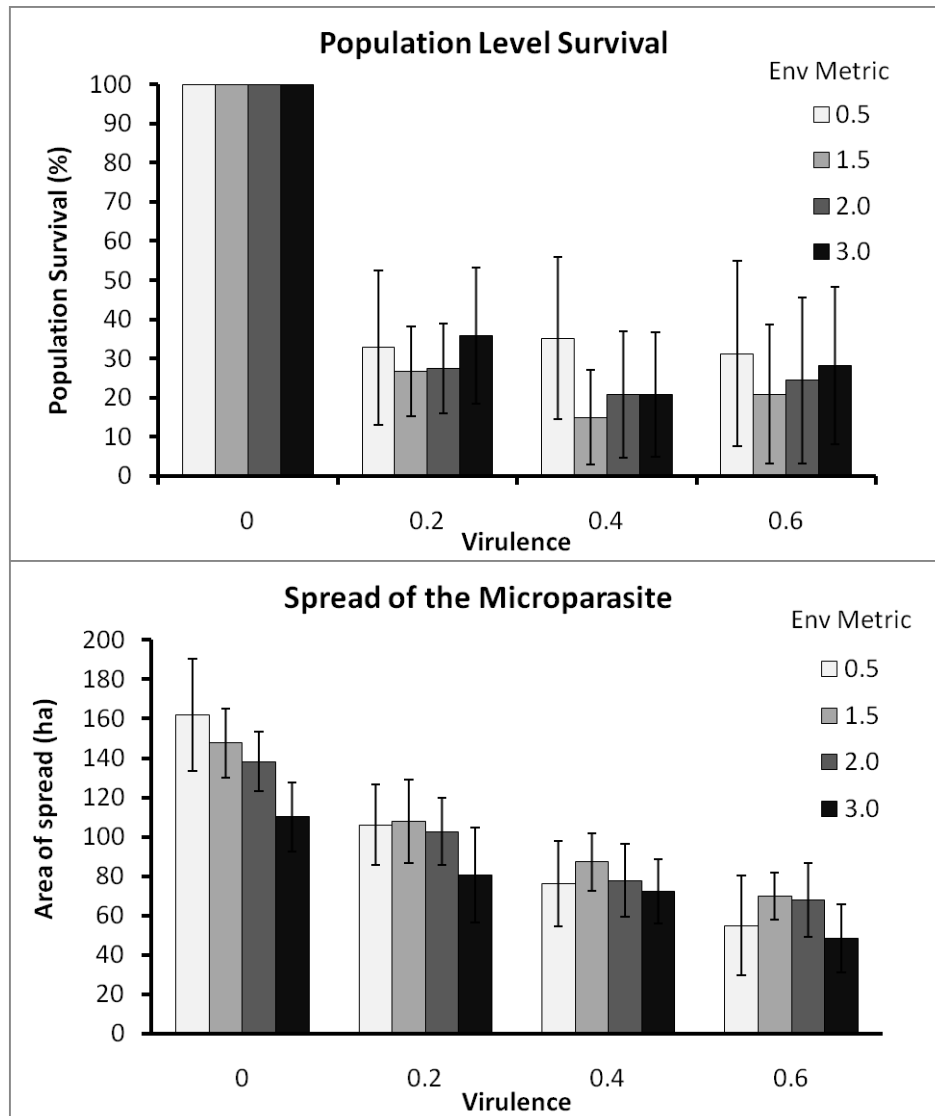


Fig. 5: Effects of varying resource distribution on a) overall population survival rates, and b) spread of the microparasite (ha). The environmental metric, pareto shape parameter, is used to modify the distribution of resources in the simulation. We see that population survival is lower and overall microparasite spread is higher at an optimal resource distribution (pareto shape parameter: 1.5). (Simulation was run a minimum of 300 times for each environmental metric: 0.5, 1.5, 2.0, and 3.0).

4. Discussion

Insights from this model focusing on the effects of changing resource distributions, lead us to predict that resource clumping within habitat patches could be an important factor affecting transmission rates within populations of animals that use patch environments. May animal species, including red colobus, use spatially clumped resources, suggesting that this may represent a general phenomenon. The model also highlights the fact that transmission rates would be spatially dependant, with certain parts of the landscape being hot spots of transmission, such as habitats with few high resource sites.

To generalize our results, we found that there were two forest characteristics that described the distribution of resources that created landscape hot spots of transmission rates; 1) the overall variation in forest tree age classes and 2) the number of high resource sites. The former represented a level of heterogeneity in the forest, whereas the latter represented the effects of shared resource sites; both are not entirely independent of each other. These two characteristics represented two different processes, and their different effects reflect the different spatial scales on which they operate. Changes to the level of heterogeneity of the landscape affects the foraging patterns of groups at the scale of neighbouring resources, influencing short range movement patterns and overall spread of the groups and effecting transmission dynamics within the group. Whereas the number of high resource sites affects foraging patterns at spatial scales closer to that of a group's home range, influencing long range movement patterns, and affecting transmission between groups. The relative importance of these two characteristics in determining overall parasite transmission was determined in our simulation by the importance that spatial memory plays in determining red colobus movement patterns in the wild, suggesting that the density of high resource sites would be the most significant factor in determining the

spread of the microparasite.

Species of non-human primates are thought to retain spatial memory with which they develop cognitive maps of their environments resources (Poucet, 1993; Di Fiore and Suarez, 2007; Janson and Byrne, 2007). Our model suggests that resources sites which are shared in memories of individuals in a group would be an important factor for the transmission of disease. We predict therefore that there could be different evolutionary strategies with regards to a species' use of a cognitive map towards limiting contact with other groups, such as the development of territoriality or strong preference for familiar feeding sites. We might therefore expect that in heavily parasitized communities, factors which reduce the similarities of inter group cognitive maps might be selected for.

If we apply the simulation results to our specific study site, to nearby areas that were selectively logged in the late 1960s, we can compare forest structure characteristics of undisturbed and disturbed areas. We observe a general trend between the intensity of logging and a decrease in the proportion of large trees and an increase in the proportion of mid-sized trees. This suggests that logged areas, with little to no large trees, will produce a resource landscape which consists of a low resource medium populated with very few high values food sites, whereas unlogged areas consist of many high value resource sites. From our plot data (5 plots within logged area; 11 from the undisturbed area) we gain an estimate of significant resource sites (trees > 50 cm DBH) within each area: unlogged 39.5 significant sites/ha, logged 18 significant sites/ha. From the results of our model we would predict that, in the logged areas, the density of resources would act to increase transmission rates between groups leading to less variation in disease states among groups and possibly an increase in parasite levels. However it must be noted that this is under the assumption that the estimated high value resource sites are

randomly distributed and that host densities are similar in both areas and that logging does not induce other stressor that could influence immunity (e.g., a decrease in available foods).

GIS capabilities were embedded and used within our agent-based model, following past approaches which have highlighted the benefits of ABM/GIS systems (Brown et al., 2005). We found the ability of agents to use the spatial functions built within our GIS to be an effective spatial decision support tool (buffer analysis, logical tests (contained within polygon), nearest neighbour analysis). It has also allowed for easier transfer to and from external GIS software, facilitating access to field data, as well as more in-depth analysis on model output. The GIS capabilities allowed us to develop a complex landscape on which we could run our individual based SEIR model, similar to approaches taken by Linard et al (2008), and to apply such a model in the context of an important wildlife host-parasite system. As long as the context considered is a general one (e.g., organisms using patchy resources), then the model will be broadly applicable to a number of wildlife-disease interactions.

The fact that red colobus group movements and foraging patterns are closely associated with environmental characteristics (Snaith and Chapman, 2008), creates an opportunity to simulate how changes to environmental characteristics can affect aspects of host parasite interactions within a population. A similar approach, using detailed environmental data along with behavioural data, should be applicable to other host-parasite systems as well. Examples of systems that would benefit from these approaches might be: 1) directly transmittable parasites reliant on a mobile host for transmission, where connectivity amongst the host population is determined from landscape properties (e.g., rabies) 2) parasites who have a life stage in the external environment and are relatively immobile, where overlap and range use of the definitive host would be an important (e.g., gastro-intestinal parasites), 3) parasites in which a specific host is responsible for spreading disease amongst separate reservoir populations (e.g., Lyme disease, poxviruses, ebola virus). The value of our study is therefore not only to predict how

disease transmission in red colobus in Kibale might respond to changing forest structure, but also to provide ground work for predicting similar effects in other host species where environmental characteristics are important drivers affecting transmission of parasites. Predictions from such models should be useful in constructing informed management plans for endangered species that account for the transmission of infectious disease across real landscapes (depicted with remote sensing data), as well as for predicting the disease-related effects of a changing climate, habitat fragmentation, logging, or other similar anthropogenic changes to wildlife habitats.

Acknowledgements

Funding the Ugandan research was provided by Canada Research Chairs Program, Wildlife Conservation Society, Natural Science and Engineering Research Council of Canada, National Geographic, and the National Science Foundation to CC, and to the generous support of the Morris Animal Foundation through its support of the Kibale EcoHealth Project (award #D07ZO-024 to TG). Support for the modelling work was provided by Natural Science and Engineering Research Council of Canada (RS, CC). Permission to conduct the research in Uganda was given by the National Council for Science and Technology and the Uganda Wildlife Authority. We would like to extend a special thanks to Mel Lefebvre, Lauren Chapman, Julian Zhunping, and Andre Costopoulos for helpful comments on this manuscript, as well as to Margaret Kalácska for providing access to computer resources.

Conclusion

The effect of the spatial structure of habitats is thought to influence transmission rates of disease within hosts; the extent of which depends on the behaviour of the hosts and parasites. The effects of barriers (physical and ecological), connectivity between habitat patches, high value areas and size of habitats are thought to be some of the factors that can influence disease (Ostfeld et al., 2005). Spatial factors, such as these, are not taken into account in traditional epidemiological models (Keeling and Eames, 2005).

Within this chapter I use a spatially explicit ABM as a virtual laboratory to test and develop theories regarding how forest structure influences disease transmission in red colobus monkeys. I tested the effects of changing resource distributions, and made inferences on how these changes influence contact, and thus transmission rates, between red colobus. We find that there is an optimal distribution of resources for which disease transmission is facilitated, and that the characteristics of this distribution fit that of a logged area of forest within our study area. As is the case with all laboratory experiments, in which observations are made under controlled environments, further tests are needed, especially in natural conditions, to increase the strength of inference of the given results. However laboratory analysis can lead to a clearer understanding of what is taking place in the natural world, and maybe more importantly, generate more in-depth and structured questions about the system of interest. For example this ABM helped generate questions regarding: the interplay between spatial memory in an organism and transmission of disease, the density of high value sites and rates of contact within a host species, the effects of logging on the transmission of disease in folivorous primates, as well as to the use

of such spatially explicit models in real time zoonotic outbreaks for interventions and public health planning.

The role of environmental structure in affecting host-parasite relationships is one that is becoming increasingly important to understand (Daszak et al., 2000). Our effect as a species on the environment of our planet largely outweighs any other species past, or present; the implications of which are slowly being understood by societies today (Vitousek et al., 1997; Scheffer et al., 2001). The above chapter, and subsequent research, is an attempt, albeit small and only an initial one, at understanding some of these implications when it comes to infectious disease and our impact on the environment of our planet.

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Appendix 1: Maps of Kibale

Figure 1: Location of Kanyawara within Kibale National Park, Uganda.

