

1 **Designing green plasticizers: influence of alkyl chain length on biodegradation and**  
2 **plasticization properties of succinate based plasticizers**

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21 **Abstract**

22 Phthalate diesters such as di (2-ethylhexyl) phthalate (DEHP) are considered ubiquitous contaminants and  
23 are poorly biodegraded in the environment. Moreover, both the parent compound and stable metabolites  
24 such as mono (2-ethylhexyl) phthalate (MEHP) are linked to several negative impacts on the environment  
25 and human health. Earlier work established that saturated diester compounds, such as succinates, showed  
26 better biodegradation characteristics and comparable plasticizer properties compared to DEHP. In this  
27 work we examine the effect of alkyl chain length of succinate molecules on plasticizer and biodegradation  
28 properties. This included both the side chains (n-ethyl to n-octyl) as well as substituents on the middle part  
29 of the succinate molecule. We showed that the common soil bacterium *Rhodococcus rhodocrous* could  
30 rapidly break down all unsubstituted succinates, without the appearance of stable metabolites.  
31 Furthermore, the organisms used the plasticizer metabolites as carbon source. The introduction of a large  
32 cyclohexyl substituent on the succinate resulted in a poorer degradation rate. Glass Transition  
33 Temperature (T<sub>g</sub>) measurements were performed to evaluate plasticizer properties and showed that longer  
34 side chains reduced the T<sub>g</sub> more efficiently, while large cyclohexyl substituents on the succinate  
35 decreased this effect. However, all compounds performed better or equal to DEHP at reducing the T<sub>g</sub>.

36

37 **Keywords**

38 Green Plasticizer, Phthalates, Biodegradation, Succinate

## 39 **Introduction**

40           Among the many additives to polymers that serve to improve certain properties, plasticizers  
41 account for a large fraction, as they can make up to 40% of the overall material (Mersiowsky et al., 2001).  
42 These are usually relatively small compounds, which serve to render polymers flexible and workable  
43 (Sears and Darby, 1982). Most plasticizers are not covalently bound to the polymer matrix which means  
44 that they can leach out of the material (Jaeger and Rubin, 1970). Approximately 80% of all plasticizers  
45 produced worldwide are used in poly(vinyl chloride) (PVC) formulations (Stevens, 1999; Murphy, 2001)  
46 and about 90% of the plasticizers used for PVC are diesters of phthalic acid (Murphy, 2001).

47           Di (2-ethylhexyl) phthalate (DEHP) is particularly well studied owing to its wide range of toxic  
48 effects (Lloyd and Foster, 1988; Akingbemi et al., 2001; Foster et al., 2001; Gazouli et al., 2002;  
49 Akingbemi et al., 2004; Horn et al., 2004). Studies have found DEHP in samples from many different  
50 types of environments including the interior of buildings (Thuren and Larsson, 1990; Bauer and  
51 Herrmann, 1997; Cartwright et al., 2000; Staples et al., 2000; Butte et al., 2001; Becker et al., 2004; Horn  
52 et al., 2004). Due to this, DEHP and its metabolites are generally considered as ubiquitous environmental  
53 contaminants (Wams, 1987). DEHP and its stable metabolites mono (2-ethylhexyl) phthalate (MEHP), 2-  
54 ethyl hexanol and 2-ethyl hexanoic acid have been shown to be toxic to microorganisms (Nalli et al.,  
55 2002; Horn et al., 2004) and linked to antiandrogenic activities in humans (Richburg and Boeckelheide,  
56 1996; Swan et al., 2005; Pant et al., 2008; Fan et al., 2010; Piche et al., 2012).

57           Several bans on the use of phthalates in various jurisdictions (EU/2005/84/EC, 2005; CPSIA,  
58 2008; HPA, 2010) have spurred interest in developing “greener plasticisers”. Research has been done for  
59 example on modified dibenzoate (Firlotte et al., 2009; Kermanshahi et al., 2009b), oligo-esters made from  
60  $\epsilon$ -caprolactone (Shi et al., 2011) and small diesters based on four carbon diacids (Erythropel et al., 2012).  
61 Stuart et al. worked with diesters of succinic acid, a natural molecule involved in the citric acid cycle of  
62 mammals, and showed that it was compatible with PVC (Stuart et al., 2010). In this work, we present data  
63 demonstrating the influence of structural characteristics on plasticizing as well as biodegradation

64 properties of various succinate-based plasticizers. Alkyl chain length was varied either by altering the  
65 alcohol used or the central diacid leading to several series of homologous compounds.

66

## 67 **Materials and Methods**

### 68 **Plasticizers**

69 Two of the plasticizers tested, diethyl succinate (DES; 99%) and di (2-ethylhexyl) phthalate  
70 (DEHP; 99%) were purchased from Sigma Aldrich. All others (Figure 1) were synthesized using a Dean-  
71 Stark esterification as described earlier (Erythropel et al., 2012).

72 Dihexyl 2-methyl succinate (DHMS) 2-methyl succinic acid (4.45 g, 33.7 mmol, 99%, Sigma  
73 Aldrich) and 1-hexanol (6.46 g, 63.2 mmol, 98%, Sigma Aldrich) were dissolved in 100 ml of toluene  
74 (99%, Sigma Aldrich) with catalytic amounts of concentrated sulphuric acid (Fisher Scientific). This  
75 mixture was refluxed at 120 °C over night in a 250 ml flask. A Dean-Stark trap was used between the  
76 flask and the reflux condenser to remove water. After cooling, the clear solution was washed with three  
77 aliquots of 50 ml of a saturated aqueous NaHCO<sub>3</sub> solution. The combined aqueous phases were set aside  
78 to recover any monoester produced and the combined organic phases were washed with deionized water,  
79 dried with sodium sulphate (Anachemia) and then the solvents were removed on a rotatory evaporator at  
80 95° C at a pressure of 1.2 kPa. (Büchi RE III with Heidolph Rotavac Valve Control). A colorless, oily  
81 liquid was obtained. Yield: 89%.

82 All other syntheses were carried out in a similar manner, keeping the molar ratio of diacid to  
83 alcohol at 1 to 2. Whenever possible, the anhydride was used preferentially over the free diacid. The  
84 purity of each compound was verified by NMR spectroscopy, the spectra are shown in Appendix A. The  
85 yield of each reaction was obtained as follows: DBS: 90%; DHS: 99%; DOS: 82%, DEHS: 81%  
86 (Erythropel et al., 2012); DHMS: 89%; DHCHS: 88%; DHMCH: 85%.

87 Succinic anhydride (1.0 g, 10.0 mmol, 99%, Fisher Scientific) and 1-butanol (0.78 g, 10.5 mmol,  
88 99%, Fisher Scientific) were dissolved in 10 ml of chloroform (99%, Fisher Scientific) and heated to  
89 80 °C in a hermetically sealed vial. Once a clear solution was obtained, the vial was allowed to cool to

90 room temperature and the chloroform was removed on a rotatory evaporator (Büchi RE III with Heidolph  
91 Rotavac Valve Control). The products were recrystallized from hexanes (99%, Fisher Scientific).

92 All other succinate monoesters were prepared in a similar manner, keeping an equimolar ratio of  
93 succinic anhydride (or acid in some cases) to alcohol. Again, the structures of the products were verified  
94 using NMR, the spectra are shown in Appendix A. The yield of each reaction was obtained as follows:  
95 MBS: 85%; MHS: 45%; MOS: 70%.

96 The synthesis of the mixture of monohexyl 2-methylsuccinate (MH2MS) and monohexyl 3-  
97 methylsuccinate (MH3MS) differed significantly as it was recovered as a by-product from the synthesis of  
98 the corresponding diester, DHMS (as described above). The combined aqueous phases from the washing  
99 step were brought to pH 2 with hydrochloric acid (36%, Sigma Aldrich), and extracted with three aliquots  
100 of 30 ml of chloroform. The combined organic phases then washed with a saturated aqueous solution of  
101 NaHCO<sub>3</sub>. Then this aqueous phase underwent a chloroform extraction followed by one more cycle of  
102 extractions. The solvent in the final organic phase was removed using a rotatory evaporator (Büchi RE III  
103 with Heidolph Rotavac Valve Control).

104

## 105 **Extrusion**

106 A conical intermeshing twin-screw extruder (Haake Minilab, Thermo Electron Corporation, screw  
107 diameter 5/14 mm conical, screw length 109.5 mm) was used to incorporate the plasticizers into  
108 unplasticized PVC (UPVC, Solvay Benvic, France). The batch feed size was 3 g, rotation speed of the  
109 screws was set to 60 min<sup>-1</sup> and the operating temperature was between 110 °C and 130 °C. Due to the  
110 difference in viscosity between UPVC and the liquid plasticizers, the addition of plasticizer had to be  
111 carried out in a series of steps. The first step was to prepare a blend containing 20 parts per hundred rubber  
112 (phr, 16.6 wt.%), 4 phr of epoxidized soy bean oil (Chemtura Corporation) as heat stabilizer and 5 phr of  
113 stearic acid (Fisher Scientific) as lubricant. This was collected and recycled through the extruder a second  
114 time to ensure homogeneity. This blend of 20 phr served as master batch to prepare blends of higher

115 plasticizer in steps that added an additional 20 phr of the plasticizer each time. No additional amounts of  
116 heat stabilizer or lubricant were added.

117

### 118 **Differential Scanning Calorimetry**

119 Temperature modulated differential scanning calorimetry (TA Instruments Q100) was carried out  
120 to determine glass transition temperatures. Samples were cut into slices of 1-2 mg each, which were  
121 placed into a standard DSC pan (TA Instruments, model #070221). A top was crimped on and the total  
122 weight and weight of the sample was recorded. Pans were loaded into the autosampler of the instrument,  
123 along with an empty pan for calibration. After quenching the sample at -90 °C for 5 min, a heating rate of  
124 2 °C min<sup>-1</sup> was applied, which was superimposed by a sinusoidal modulation of 1.27 °C with a period of  
125 60 s. Once 100 °C was reached, the sample was held at this temperature for 5 min. A preliminary cycle  
126 was done to erase the thermal history of the sample. The data from the second cycle was analysed using  
127 the software TA Universal Analysis, plotting reversible heat flow against temperature to determine the  
128 glass transition temperature according to ASTM D-3418 (ASTM D-3418, 2003).

129

### 130 **Biodegradation study**

131 The microorganism *Rhodococcus rhodocrous*, American Type Culture Collection (ATCC) 13808  
132 was used to evaluate the biodegradation properties of the pure plasticizers. For each measurement, it was  
133 necessary to extract the entire contents of a shake flask. Taking several samples from one flask led to  
134 errors because of the difficulty in obtaining a homogeneous sample. For each experiment, sets of 500 ml  
135 Erlenmeyer flasks fitted with foam caps were prepared, containing 100 ml Minimum Mineral Salt  
136 Medium (MMSM: 6 g L<sup>-1</sup> Na<sub>2</sub>HPO<sub>4</sub>, 4 g L<sup>-1</sup> NH<sub>4</sub>NO<sub>3</sub>, 4 g L<sup>-1</sup> KH<sub>2</sub>PO<sub>4</sub>, 0.2 g L<sup>-1</sup> MgSO<sub>4</sub>•7H<sub>2</sub>O, 0.014 g L<sup>-1</sup>  
137 Na<sub>2</sub>EDTA, 0.01 g L<sup>-1</sup> CaCl<sub>2</sub>•2H<sub>2</sub>O and 0.01 g L<sup>-1</sup> FeSO<sub>4</sub>•7H<sub>2</sub>O, Fisher Scientific), 10 mM pure plasticizer  
138 and 0.1 g L<sup>-1</sup> yeast extract. Other sets of flasks were prepared the same manner, but these included 2 g L<sup>-1</sup>  
139 hexadecane as an easily used carbon source. The flasks were autoclaved at 121 °C and 100 kPa for 15  
140 minutes (Steris Amsco Lab 250), cooled and then inoculated with 1 ml of cell broth from a previously

141 grown culture, using sterile techniques in a laminar fumehood (Baker Company, Model VBM600). The  
142 flasks were cultured in an incubator-shaker at 30 °C and 140 RPM (Multitron II, Infors AG).

143           The complete contents of one flask of each set were extracted on day 0. Subsequent flasks  
144 were extracted at intervals of 1 to 5 days depending on the degradation rate. All of these extractions were  
145 done in the same manner. First, the contents were adjusted to pH 2 with concentrated sulphuric acid and  
146 then extracted with 20 ml of chloroform containing 2 g L<sup>-1</sup> pentadecane, as an internal standard. The  
147 isolated organic phase was stored at 4 °C until analysis.

148           In order to compare the results of the several biodegradation experiments, rate constants for the  
149 first hydrolysis step of the diesters were approximated using a first order fit. With these, values for half  
150 lives were calculated (Table 1 and Table 2).

151

## 152 **Gas Chromatography**

153           Determination of plasticizer concentrations in the extracts of the flasks, as well as detection and  
154 quantification of metabolites, was done using a gas chromatograph (Trace GC Ultra with AI3000  
155 Autosampler, Thermo Scientific). After appropriate dilution of the samples, 1 µL was injected into the  
156 GC, which was fitted with a Restek RTX®-5 column (length 30 m, 0.32 mm internal diameter, 0.25 µm  
157 film). A flame ionization (FID) was used. In order to calculate concentrations from the ratios of peak areas  
158 of the compounds to those of the internal standard pentadecane, calibration curves were prepared for each  
159 compound of interest.

160

## 161 **NMR Spectroscopy**

162           <sup>1</sup>H - Nuclear Magnetic Resonance spectroscopy was carried out using one of two spectrometers:  
163 Varian Mercury-300 (<sup>1</sup>H = 300 MHz) and Varian Unity-500 (<sup>1</sup>H = 500 MHz). The solvent used for all  
164 measurements was deuterated chloroform (CDCl<sub>3</sub>), with tetramethylsilane (TMS) as internal standard. The  
165 chemical shifts δ are indicated in ppm.

166

## 167 **Statistics**

168 Statistical Analysis was done using the software GraphPad Prism 5; one-sample t-tests and one-  
169 way ANOVA tests with Bonferroni post test were carried out as indicated. A p value less than 0.05 was  
170 taken as significant.

171

## 172 **Results**

### 173 **Syntheses**

174 The synthesis of the monoester of methyl succinic acid resulted in two isomers, depending on  
175 which carboxylic function was esterified, as shown in Figure 1. Figure 2 contains the relevant parts of the  
176 gas chromatogram (a) as well as the NMR spectrum (b). By integrating the GC data it can be seen that one  
177 of the monoesters was preferentially formed during synthesis. The identity of the major and minor isomers  
178 was resolved with NMR. This is shown in an enlargement of the region for the single proton on the optical  
179 centre of the methylated carbon atom between the two carboxylate groups of the diacid core of either  
180 isomer (see Figure 1 for structures). The higher shifted signal belongs to the proton in MH3MS, as this  
181 proton is in the  $\beta$ -position to the COOH group, resulting in a slightly higher chemical shift, while the  
182 corresponding proton in MH2MS is in the  $\gamma$ -position to the COOH group, **but** in the  $\beta$ -position to the ester  
183 function. These results suggest that the peak at 8.8 min belongs to MH2MS (smaller peak areas), and the  
184 peak at 8.9 min to MH3MS (larger peak areas). Since the pure compounds were not obtained, the  
185 calibration curve for the unsubstituted monohexyl succinate was used, assuming the effect of the added  
186 methyl group would have a negligible effect in gas chromatography.

187

### 188 **Plasticizing Studies**

189 Glass Transition Temperature ( $T_g$ ) data for blends of 40 phr (28.6 wt.%) in PVC are shown in  
190 Figure 3. Comparing the data for the branched di (2-ethylhexyl) succinate and the unbranched dihexyl  
191 succinate there was no significant difference in  $T_g$  reduction (one sample t-test,  $p = 0.2254$ ). A significant  
192 statistical difference was found between DHS and DEHP, however (one sample t-test,  $p = 0.0002$ ).



193 Figure 4 shows the glass transition temperatures for several succinates with varying substituents  
194 on the central part of the molecule, while the alcohol used to esterify all compounds was hexanol.  
195 Amongst all data, a significant statistical difference was found between the means (one-way ANOVA,  
196  $p < 0.0001$ ). DHS was not statistically different than DHMS (addition of a methyl group), however it was  
197 statistically different than DHMCH and DHCHS (Bonferroni post test, both  $p < 0.0001$ ), which both  
198 contain larger substituents. DHMS was also found statistically different than DHMCH and DHCHS  
199 (Bonferroni post test, both  $p < 0.0001$ ). The results for both compounds with large substituents, DHCHS  
200 and DHMCH, were found to not be significantly different.

201

## 202 **Biodegradation Studies**

203 Figure 5 shows examples of typical biodegradation experiments. Figure 5A shows the  
204 biodegradation pattern for dihexyl succinate. The parent compound is quickly hydrolysed and a  
205 corresponding increase in monohexyl succinate and hexanol are observed, however by day 3 and day 7,  
206 respectively, the presence of these compounds have decreased to trace amounts. No significant build-up of  
207 any other metabolite, including the corresponding hexanoic acid, was detected. Figure 5A also shows the  
208 concentration of hexadecane in the solution and its concentration only starts decreasing once all of the  
209 other compounds have disappeared.

210 Figure 5B shows the biodegradation pattern for the dihexyl ester of the methyl-substituted  
211 succinate. An equally quick hydrolysis of the parent compound is observed, and there is only a temporary  
212 build-up of the metabolite hexanoic acid. As expected, two metabolites with very close retention times in  
213 GC could be detected. These corresponded to the two structurally different monoesters MH2MS and  
214 MH3MS. After day 3, the concentrations of the monoesters remained at a constant ratio of approximately  
215 5:1 of MH3MS to MH2MS.

216 Table 1 shows data for the half lives and time of maximum concentration during the  
217 biodegradation studies conducted with succinate esters of varying side chain length as well as results for  
218 several substituted succinates with constant side chain length. No data is presented for the small diethyl

219 succinate and its metabolites because of their very quick removal from the broth, as well as for the bulky  
220 dihexyl 2-cyclohexylsuccinate, which showed almost no hydrolysis of one of the ester bonds, resulting in  
221 no significant amount of metabolites detected.

222 For the remaining compounds, a quick hydrolysis of the diester was observed, except for the  
223 compound in which R'' was altered from a hydrogen to a cyclohexyl group (Figure 1). Table 1  
224 summarizes the observed metabolites and their stability. Complete removal of the corresponding  
225 monoester within the timeframe of the degradation experiments was only observed if R' was six carbons  
226 or less and no hydrogen was replaced in the central structure of the succinate molecule (R''). Both  
227 corresponding alcohol and acid were quickly removed for all compounds with R' equal to or smaller than  
228 six.

229 Table 2 shows a comparison of half lives of selected succinate compounds being biodegraded  
230 with and without hexadecane present. For all compounds, hexadecane did not affect the half life and the  
231 onset of hexadecane use was correlated to the disappearance of the plasticizer compounds.

232

233

234

## 235 **Discussion**

236 DEHP has been the object of intense scrutiny in North America and Europe due to its potential  
237 toxic effects. A safe replacement of DEHP requires the design of compounds that adequately plasticizes  
238 PVC, biodegrades upon disposal, and avoids the buildup of stable and toxic metabolites. In this work we  
239 present a series of succinate plasticizers with good biodegradation kinetics and similar plasticizing  
240 properties compared to DEHP.

241 Earlier work has shown that there is a strong correlation between the glass transition temperature  
242 of PVC mixtures and their tensile properties such as elongation at break and secant modulus (Erythropel et  
243 al., 2012). From the glass transition temperature measurements in Figure 3, it is clear that the length of  
244 the alcohol portion of the plasticizers is an important aspect of their plasticizing properties. It should be

245 noted that increasing the length of the alcohol used to esterify by one carbon atom must increase the  
246 length of the longest chain of atoms in these plasticizers by two carbon atoms. The longer the alcohol used  
247 to make the diester plasticizer, the more effective the diester is at lowering the Tg. A decrease in Tg  
248 indicated that the PVC was becoming more flexible. So, these data can be interpreted as showing that as  
249 the plasticizer molecule becomes longer the interaction with the polymer chains becomes more effective at  
250 reducing the crystallinity of the PVC.

251 The Tg data for the branched 2-ethylhexyl succinate is almost identical to the Tg of the succinate  
252 made with hexanol. While the 2-ethyl containing compound has a higher molecular weight, the lengths of  
253 the longest chains of atoms in the two plasticizers are identical (Figure 1). By comparison, the Tg of the  
254 succinate made with octanol (longest chain of C and O atoms of 22 atoms) is significantly lower than that  
255 for the compound made with 2-ethyl hexanol, despite that these two compounds have an identical  
256 molecular weight. This supports the argument that the relative length of the plasticizer is the important  
257 factor in lowering Tg (Figure 3).

258 The data for the modified succinate plasticizer, in which one of the central hydrogen atoms has  
259 been replaced with a methyl group, shows little or no effect on Tg reduction, compared to the non-  
260 substituted succinate with the same alcohol (Figure 4). Again, this is suggesting the key factor is the  
261 length of the longest chain of atoms. Adding a larger group to the central diacid also significantly affects  
262 the overall plasticizing properties. When succinic acid is substituted with a cyclohexyl group (DHCHS)  
263 or when the central part of the diacid is part of a cyclohexane ring (DHMCH), Tg reduction is not as  
264 effective. These results can still be explained by considering the properties of these compounds relative to  
265 the plasticizer made with simple succinic acid. The cyclohexyl substituent would hinder rotation about  
266 the central bond of the plasticizer and no rotation is possible in case of the DHMCH. Rotational freedom  
267 is considered to play an important role in plasticizing abilities of an additive (Erythropel et al., 2012) and  
268 thus it is not surprising that these two compounds are poorer plasticizers than the compound made from  
269 the simpler succinic acid.

270 All of the above compounds are thermally stable. However, it has been shown that plasticizers  
271 inevitably leach out of the formulations and enter the environment (Jaeger and Rubin, 1970; Thuren and  
272 Larsson, 1990; Staples et al., 1997; Cartwright et al., 2000; Becker et al., 2004; Horn et al., 2004). In  
273 some cases, these released compounds can be found to have long half lives and therefore accumulate in  
274 the environment. It would be desirable to develop plasticizers that can be biodegraded quickly and not  
275 produce stable toxic metabolites.

276 *Rhodococcus rhodocrous*, a common soil bacterium, was chosen as the organism due to its known  
277 ability to degrade a broad range of compounds, including aliphatic hydrocarbons (Jones and Goodfellow,  
278 2012). Previous studies have shown that *R.rhodocrous*. has the capability to hydrolyse diester plasticizers  
279 rapidly in the timeframe of these experiments, thus allowing to compare relative hydrolysis rates between  
280 different compounds (Nalli et al., 2002; Nalli et al., 2006c, Sauvageau et al., 2009).

281 Taken together, the biodegradation data for all of the plasticizers tested here are consistent with  
282 the mechanism of biodegradation worked out for other diesters in previous experiments (Nalli et al.,  
283 2006b; Nalli et al., 2006c; Kermanshahi et al., 2009a). In the presence of bacteria, the diesters are first  
284 hydrolysed to the corresponding monoester and an equivalent of the alcohol used to synthesize the diester.  
285 The monoester is then hydrolysed to ultimately yield the diacid and two equivalents of the corresponding  
286 alcohol. The alcohols are then oxidized to yield the corresponding carboxylic acids and will readily  
287 undergo  $\beta$ -oxidation (Nalli et al., 2006a) . The intermediate aldehydes are inferred but were not observed,  
288 which can be explained by the volatility of these compounds as explained in earlier work (Nalli et al.,  
289 2006c).

290 As with the plasticizing properties in PVC mixtures, the type of alcohol used to make the  
291 succinate diesters influences the rate of biodegradation Table 1. The longer the aliphatic chain of the  
292 alcohol – and thus of the overall length of the plasticiser – the slower is the overall degradation rate. This  
293 can be attributed to the increasing hydrophobicity of the compounds as the length of the hydrophobic  
294 portions of the plasticizers increase. This would lead to a decrease in the water-solubility of the  
295 compounds, which would make a plasticizer less available to the microbes in their aqueous environment.

296           The di (2-ethylhexyl) succinate is degraded more slowly than the equally long dihexyl succinate,  
297 yet only slightly more quickly than the di-n-octyl succinate (Table 1). This could be attributed to its 2-  
298 ethyl branches adding to the hydrophobicity of the molecule but it is just as likely that the ethyl branches,  
299 located near the ester bond, are sterically hindering its interaction with the appropriate enzymes.

300           It is important to note that all of the succinate diesters were degraded in a few weeks, which is  
301 significantly faster than the rate for the commercial plasticizer DEHP. In earlier work with di (2-  
302 ethylhexyl) esters, we were able to show that one of the key parameters to hydrolysis of the ester bonds is  
303 the structure of the central part of the diester molecule (Erythropel et al., 2012). With unsaturated diesters  
304 that had their two ester groups in a cis orientation, similar to DEHP, the first hydrolysis step to the  
305 monoester was very slow. Yet when saturated compounds, such as a succinate diester were used, the first  
306 hydrolysis step was much faster. With the compounds in this study, the change in the alcohol used makes  
307 biodegradation even more facile.

308           While it is important that the first hydrolysis step is much quicker than that observed for DEHP, it  
309 is also essential to consider the possibility of the accumulation of stable metabolites. As expected, the use  
310 of straight-chained alcohols was advantageous with regard to the metabolites as reported earlier (Nalli et  
311 al., 2006a). None of these alcohols resulted in accumulation of stable, toxic metabolites (Table 1) as was  
312 observed with any di (2-ethylhexyl) compounds, namely 2-ethyl hexanoic acid (Nalli et al., 2002, 2006a,  
313 2006b; Erythropel et al., 2012).

314           The monoester of DEHP, MEHP, has been studied extensively due to its suspected effect on the  
315 endocrine system (Richburg and Boekelheide, 1996; Swan et al., 2005; Pant et al., 2008; Fan et al., 2010;  
316 Piche et al., 2012). MEHP has long been suspected as the active metabolite of DEHP, although more  
317 recent work has shown 2-ethylhexanal and 2-ethyl hexanol to also be biologically active (Piche et al.,  
318 2012). We showed that the stability of the monoesters is influenced by the choice of the central diacid  
319 (Erythropel et al., 2012), and in this study we show the influence of the side chain on the monoester  
320 stability. The monoesters were observed in the degradation experiments with all of the compounds (Table  
321 1). The shortest diester, diethyl succinate, degraded quickly as did its monoester as it was not observed.

322 The next two compounds in the series, dibutyl and dihexyl succinate (Figure 5A), did generate small  
323 amounts of their monoesters, however these were hydrolysed within a week. The exact time is obscured  
324 by the fact that monoesters continued to be released as the parent diesters were hydrolyzed. It was also  
325 found that the removal of this compound was even faster in the presence of a second carbon source;  
326 hexadecane. Future studies will evaluate the toxicity and potential antiandrogenic nature of these  
327 compounds.

328 The longest monoester in this series was monooctyl succinate and this was the slowest to  
329 biodegrade. This means that, as observed with the parent diesters, as the series of compounds becomes  
330 longer, the rate of biodegradation of the monoesters becomes slower. Again, this is probably related to the  
331 solubility of these compounds in water. However, none of these monoesters are as intractable as those  
332 observed with some of the compounds, such as MEHP, studied in earlier work (Nalli et al., 2002; Horn et  
333 al., 2004; Sauvageau et al., 2009). Another important consideration is that the earlier work was done in  
334 the presence of hexadecane. This work shows that while hydrolysis of the monoester is faster when  
335 hexadecane is present, it is not essential (Table 2).

336 The study's results with the compounds made with the branched alcohol, 2-ethylhexanol, show  
337 these to be stable as has been reported previously (Nalli et al., 2006c; Erythropel et al., 2012). It should be  
338 noted that without the presence of hexadecane as an easily used carbon source, almost none of the 2-  
339 ethylhexyl diester was degraded during the course of the experiment (data not shown). However, it was  
340 not essential when the unbranched alcohols were used. Table 2 shows that there was no significant  
341 difference between any of the pairs of experiments with or without hexadecane in the media. In fact, in  
342 every case, there was a pattern of diauxic growth and there was no significant decrease in hexadecane  
343 concentration until the last of the plasticizer remnants had been removed (Table 2). Figure 5A shows this  
344 behaviour in an experiment with dihexyl succinate. This shows that the bacteria use the released alcohols  
345 and/or succinic acid as carbon sources for growth, which is not surprising as succinic acid is part of the  
346 TCA cycle. These results are important because rapid metabolism of the fragments released after the  
347 initial steps of biodegradation reduces concerns about their potential environmental impact. On the other

348 hand, the ethyl branch in the branched compounds blocks  $\beta$ -oxidation, making this a poor carbon source to  
349 support growth of the organism and resulting in an accumulation of 2-ethylhexanoic acid.

350 The other metabolites are alcohols and acids and these were observed for every type of compound  
351 studied (Table 1). As stated above, the straight chain alcohols or the resulting organic acids can be  
352 metabolized and accumulation does not seem to be significant. The longest compounds, octanol and  
353 octanoic acid, are significantly slower to degrade than the others, but this is again attributed to relative  
354 solubility in water and these compounds are not expected to reach significant concentrations in the  
355 environment as many microbes would be expected to use all of the straight chain alcohols and acids as  
356 carbon sources.

357 The effects of modifying the central diacid on biodegradation were more complex. The addition  
358 of a methyl group to succinic acid did not prevent hydrolysis of the first ester bond (Figure 5B). However,  
359 the addition of a cyclohexyl group completely inhibited degradation of the diester (Table1). This  
360 compound would be significantly less polar and hence less soluble in water. However, the dramatic  
361 change in biodegradation rate would seem to indicate that the bulky cyclohexyl substituent also interferes  
362 with the interaction with the enzyme to affect the first hydrolysis step. This can also be seen in the pattern  
363 of the first hydrolysis step with dihexyl methylsuccinate. There are two possible monoesters depending  
364 on which of the ester bonds was hydrolyzed, MH2MS and MH3MS (see Figure 1). These two monoesters  
365 had significantly different rates of production meaning that one ester bond was more susceptible to  
366 hydrolysis than the other. This cannot be attributed to steric hindrance of the methyl substituent on  
367 hydrolysis of the diester because this would cause a reduction in the rate of hydrolysis of the ester bond  
368 closer to the methyl substituent. In fact, the opposite is true so there may be a slight preferential  
369 stabilization of an intermediate in the preferred hydrolysis pathway.

370 However, there appears to be a steric inhibition of the biodegradation of the monoesters that are  
371 produced by this first step. Neither monoester is degraded quickly. This was not observed for the  
372 degradation of the monoester from the unsubstituted dihexyl succinate (Figure 5A).

373

374

## 375 **Conclusion**

376           Based on an analysis of both the plasticizing properties and the biodegradation properties of the  
377 new diesters, the majority of them are potential green plasticizers. All of the compounds showed  
378 plasticization properties comparable to DEHP but there were trends that showed the importance of  
379 structure. Within the range tested, an increasing overall molecule length positively influenced the  
380 plasticizing properties of unsubstituted succinate compounds, while a negative influence on plasticizer  
381 properties was found when alkyl substituents were added to the middle part of the succinate compounds.  
382 This negative influence became larger with increasing size of these substituents, which relates to the  
383 ability of these compounds to rotate around the axis connecting the ester functions.

384           Structure was important to biodegradation of these compounds: larger, less water-soluble  
385 compounds were slower to disappear. There is some evidence that steric hindrance near the ester bonds  
386 inhibited the rate of hydrolysis. However, the overall observation was that most of these compounds  
387 could be biodegraded and most of the metabolites were also removed quickly. Earlier work on  
388 biodegradation of some commercial plasticizers had shown that the presence of an easily degraded carbon  
389 source was important. The biodegradation of most of the diesters in this study was faster if hexadecane  
390 was included in the medium but it was not essential. The metabolites such as small diacids and  
391 unbranched alcohols were readily used as carbon sources by the bacterium.  
392 Thus, a compromise has to be made between longer molecules favourable for plasticization and shorter  
393 molecules favourable for biodegradation. A good compromise found here would be dihexyl succinate. It  
394 was a comparable plasticizer to DEHP based on the Tg measurements and it was completely biodegraded  
395 in about one week, including all observed metabolites.

396

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